

1                   **The limits of predicting maladaptation to future**  
2                   **environments with genomic data**

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## 13 Abstract

14 Anthropogenically driven changes in land use and climate patterns pose unprecedented  
15 challenges to species persistence. To understand the extent of these impacts, genomic  
16 offset methods have been used to forecast maladaptation of natural populations to future  
17 environmental change. However, while their use has become increasingly common, little  
18 is known regarding their predictive performance across a wide array of realistic and  
19 challenging scenarios. Here, we evaluate four offset methods (Gradient Forests, the Risk-  
20 Of-Non-Adaptedness, redundancy analysis, and LFMM2) using an extensive set of  
21 simulated datasets that vary demography, adaptive architecture, and the number and  
22 spatial patterns of adaptive environments. For each dataset, we train models using either  
23 *all*, *adaptive*, or *neutral* marker sets and evaluate performance using *in silico* common  
24 gardens by correlating known fitness with projected offset. Using over 4,850,000 of such  
25 evaluations, we find that 1) method performance is largely due to the degree of local  
26 adaptation across the metapopulation ( $LA_{\Delta SA}$ ), 2) *adaptive* marker sets provide minimal  
27 performance advantages, 3) within-landscape performance is variable across gardens and  
28 declines when offset models are trained using additional non-adaptive environments, and  
29 4) despite (1), performance declines more rapidly in novel climates for metapopulations  
30 with higher  $LA_{\Delta SA}$  than lower  $LA_{\Delta SA}$ . We discuss the implications of these results for  
31 management, assisted gene flow, and assisted migration.



## 32 1 | Introduction

33 The impacts of climate change,  
 34 habitat loss, and extreme weather  
 35 events pose urgent challenges to the  
 36 management of species, communities,  
 37 habitats, and ecosystem services  
 38 (Bonan, 2008; Doney et al., 2012;  
 39 Hoegh-Guldberg & Bruno, 2010).  
 40 Traditional methods used to infer  
 41 environmental suitability, such as  
 42 reciprocal transplants and common  
 43 gardens, require time and resources that  
 44 may not be available or feasible for  
 45 many organisms of management  
 46 concern, particularly for long-lived  
 47 organisms where reproductive stages  
 48 occur after several decades of  
 49 development. Ecological forecasting  
 50 models have therefore become  
 51 increasingly germane to support  
 52 environmental decision making by  
 53 managers across both terrestrial and  
 54 marine systems.

55 In the context of population  
 56 viability in the face of environmental  
 57 change, many of these models rely on  
 58 theoretical expectations that the limits  
 59 of species' distributions are primarily  
 60 determined by the distribution of  
 61 environmental conditions (e.g., Good  
 62 1931), and that occupancy of highly  
 63 suitable habitat enables increased  
 64 abundance through greater survival and  
 65 reproduction (i.e., fitness) of individuals  
 66 (Brown, 1984). Such methods, termed  
 67 species distribution models or ecological  
 68 niche models (see Elith & Leathwick,  
 69 2009 for a discussion on terminology)  
 70 are correlative approaches that are

71 often used to predict (relative) habitat  
 72 suitability for a single species (Lee-Yaw et al.,  
 73 2022). This information is used to understand  
 74 potential impacts on the species from future  
 75 climate change. However, these methods often  
 76 ignore aspects of the species' evolutionary  
 77 history that could be important for predicting  
 78 long-term population persistence, such as the  
 79 environmental drivers of local adaptation or  
 80 spatial patterns of adaptive genetic variation  
 81 (Waldvogel et al., 2020).

82 Subsequent methods, termed genomic  
 83 offsets (reviewed in Capblancq et al., 2020;  
 84 Rellstab et al., 2021), have attempted to  
 85 address these shortcomings by modeling  
 86 relationships between environmental and  
 87 genetic variation to predict maladaptation of  
 88 natural populations to either future climates  
 89 *in situ*, or to predict the relative suitability of  
 90 these populations for the specific environment  
 91 of a restoration site. Empirical attempts to  
 92 confirm predictions from genomic offset  
 93 models are rare and, compared to attempts *in*  
 94 *silico* (Láruson et al., 2022), have found  
 95 relatively weaker relationships between  
 96 predicted maladaptation to common garden  
 97 climates and the measurement of phenotypic  
 98 proxies for fitness from individuals grown in  
 99 these same environments (e.g., Capblancq &  
 100 Forester, 2021; Fitzpatrick et al., 2021; Lind  
 101 et al., 2024). Even so, these empirical results  
 102 have consistently shown the expected  
 103 negative relationship between predicted offset  
 104 and common garden performance. Further,  
 105 many of these studies found that genomic  
 106 offsets often perform better than climate or  
 107 geographic distance alone (e.g., Capblancq &  
 108 Forester, 2021; Fitzpatrick et al., 2021;  
 109 Láruson et al., 2022; Lind et al., 2024).

110 Across empirical and *in silico* studies, little  
 111 difference in performance was found between

112 models trained using only adaptive  
113 markers (i.e., known *in silico*, or  
114 candidates from empirical genotype-  
115 environment [GEA] associations) and  
116 those chosen at random, suggesting that  
117 genome-wide data may be sufficient to  
118 capture signals relevant to  
119 environmental adaptation.

120 Together, these results suggest that  
121 genomic offset methods may provide  
122 valuable insight for management. Little  
123 is known, however, about how robust  
124 these methods are across a wide array  
125 of realistic empirical scenarios, nor the  
126 extent to which independent methods  
127 will arrive at similar conclusions when  
128 analyzing the same data. Indeed,  
129 concerns regarding the accuracy of  
130 ecological forecasting models present a  
131 primary limitation towards  
132 incorporating inferences from these  
133 models into management (Clark et al.,  
134 2001; Schmolke et al., 2010) and  
135 genomic offset models are no exception.  
136 Major questions still remain about how  
137 performance is affected by aspects of  
138 the evolutionary history of sampled  
139 populations, the type of signals in  
140 putatively ideal datasets that may  
141 mislead offset inference, the importance  
142 of identifying environmental drivers of  
143 local adaptation *a priori*, and the  
144 consistency of predictive performance  
145 across contemporary environmental  
146 space. Finally, because novel climates  
147 with no recent analog are expected to  
148 increase in the future (Lotterhos et al.,  
149 2021; Mahony et al., 2017) there is also  
150 uncertainty regarding the performance  
151 of forecasting models when predictions  
152 are made to environments that

153 drastically differ from those used to train and  
154 build the models themselves (Fitzpatrick et  
155 al., 2018; Lind et al., 2024).

156 While much uncertainty remains  
157 regarding the predictive performance of  
158 genomic offsets, the domain of applicability  
159 (i.e., the circumstances under which a method  
160 is acceptably accurate) for these methods can  
161 be more precisely defined using simulated  
162 data (Lotterhos et al., 2022). Simulated data,  
163 where there is no error in the estimation of  
164 allele frequencies, environmental variables,  
165 individual fitness, or the knowledge regarding  
166 the drivers of local adaptation, present ideal  
167 circumstances for understanding the limits of  
168 genomic offsets and the circumstances under  
169 which data from naturally occurring taxa will  
170 provide useful inference. To provide relevant  
171 inference regarding the domain of  
172 applicability, simulations should capture the  
173 complexities of empirical data with biological  
174 realism (e.g., clinal or patchy environments),  
175 present contrasting cases of differing scenarios  
176 while controlling for important features of the  
177 data (e.g., varying population connectivity  
178 but controlling for mean differentiation), and  
179 challenge methods using adversarial scenarios  
180 that capture extreme characteristics of  
181 empirical data (e.g., prediction to novel  
182 environments with no current analog  
183 available for model training; Lotterhos et al.  
184 2022).

185 Here, we use a wide array of previously  
186 published biologically realistic, contrasting,  
187 and adversarial simulations from Lotterhos  
188 (2023) in an attempt to more precisely define  
189 the limits of predictive performance of five  
190 genomic offset methods (Table 1): Gradient  
191 Forests ( $GF_{\text{offset}}$ ; *sensu* Fitzpatrick & Keller,  
192 2015), the Risk Of Non-Adaptedness (RONA,  
193 Rellstab et al., 2016), Latent Factor Mixed

194 Models (LFMM2<sub>offset</sub>, *sensu* Gain &  
 195 François, 2021, and redundancy analysis  
 196 (RDA<sub>offset</sub>, *sensu* Capblancq & Forester,  
 197 2021). The main goal of this study was  
 198 to understand how the evolutionary and  
 199 experimental parameters used in the  
 200 training and evaluation of offset  
 201 methods affect the accuracy of the  
 202 methods' projections of maladaptation  
 203 under ideal empirical scenarios (i.e.,  
 204 using data with no inherent error).  
 205 Using these scenarios, we ask the  
 206 following six questions: 1) Which  
 207 aspects of the past evolutionary history  
 208 affect performance of offset methods? 2)  
 209 How is offset performance affected by  
 210 the proportion of loci with clinal alleles  
 211 in the data? 3) Is method performance  
 212 driven by causal loci or by genome-wide  
 213 patterns of isolation-by-environment?  
 214 4) What is the variation of model  
 215 performance across the landscape? 5)  
 216 How does the addition of non-adaptive  
 217 nuisance environments in training affect  
 218 performance? 6) How well do offset  
 219 models extrapolate to novel  
 220 environments outside the range of  
 221 environmental values used in training?

## 222 2 | Methods

223 Throughout this manuscript we will  
 224 be citing code used to carry out specific  
 225 analyses in-line with the text.  
 226 Supplemental Notes S1-S2 outlines and  
 227 describes the sets of scripts or, most  
 228 often, jupyter notebooks, used to code  
 229 analyses. Scripts and notebooks are  
 230 both referenced as Supplemental Code  
 231 (SC) using a directory numbering  
 232 system (e.g., SC 02.05). More

233 information regarding the numbering system,  
 234 archiving, and software versions can be found  
 235 in the Data Availability section.

### 236 2.1 / *Explanation of Simulations and* 237 *Training Data*

238 To train offset methods we used single  
 239 nucleotide polymorphism (SNP) and  
 240 environmental data from a set of previously  
 241 published simulations (225 levels with 10  
 242 replicates each) of a Wright-Fisher  
 243 metapopulation of 100 demes on a 10 x 10 grid  
 244 evolving across a heterogeneous landscape  
 245 (Lotterhos, 2023). Each dataset was  
 246 simulated under a combination of the  
 247 following four evolutionary parameters: i)  
 248 three landscapes (10 populations x 10  
 249 populations) that varied in vicariance and  
 250 environmental gradients (*Estuary - Clines*;  
 251 *Stepping Stone - Clines*; and *Stepping Stone -*  
 252 *Mountain*), ii) five demographies that varied  
 253 population size and migration rates across the  
 254 landscape, iii) three genic levels that varied in  
 255 the effect size and number of mutations  
 256 underlying adaptation (mono-, oligo-, and  
 257 polygenic), and iv) five pleiotropy levels that  
 258 varied the number of quantitative traits under  
 259 locally stabilizing selection ( $n_{\text{traits}} \in \{1, 2\}$ ),  
 260 presence of pleiotropy (when  $n_{\text{traits}} = 2$ ), and  
 261 variability of selection strength across  
 262 individual traits (see Fig. 1 in Lotterhos  
 263 2023).

264 The adaptive trait(s) were under selection  
 265 by a different environmental variable, where  
 266 the optimum trait value was given by the  
 267 local environment on the landscape. The  
 268 adaptive trait(s) undergoing selection  
 269 responded to either a latitudinal temperature  
 270 gradient (*temp*;  $n_{\text{traits}} = 1$ ), or to both *temp*  
 271 and a longitudinal “*Env2*” gradient ( $n_{\text{traits}} =$

272 2). *Env2* represented distinct biological  
 273 analogies depending on the context: in  
 274 the *Stepping Stone - Mountain*  
 275 landscape *Env2* was analogous to  
 276 elevation (e.g., as with tree species),  
 277 whereas in the *Estuary - Clines*  
 278 landscape the *Env2* environment was  
 279 analogous to gradients of salinity within  
 280 coastal inlets connected only by the  
 281 outer marine (ocean) environment (e.g.,  
 282 as with stickleback or oyster species).

283 Twenty independent linkage groups  
 284 were simulated. Of these, mutations  
 285 that had effects on one or more  
 286 phenotypes under selection (i.e.,  
 287 quantitative trait nucleotides, QTNs)  
 288 were allowed to evolve on only ten  
 289 linkage groups, and neutral mutations  
 290 were added to all 20 linkage groups with  
 291 tree sequencing (for details see  
 292 Lotterhos 2023). Adaptive traits were  
 293 determined additively by effects of  
 294 QTNs.

295 In all simulations, phenotypic clines  
 296 evolved between each trait and the  
 297 selective environment (Lotterhos,  
 298 2023), where populations became  
 299 locally adapted to their environment,  
 300 measured at the metapopulation level  
 301 as the mean difference of demes in  
 302 sympatry minus allopatry ( $LA_{\Delta SA}$ ,  
 303 Blanquart et al., 2013).  $LA_{\Delta SA}$  equates  
 304 to the average levels of local adaptation  
 305 at the deme level which can be  
 306 calculated for each deme by both home-  
 307 away ( $LA_{\Delta HA}$ ) and local-foreign  
 308 ( $LA_{\Delta LF}$ ) measures.

309 These simulations represent a wide  
 310 array of realistic, contrasting, and  
 311 adversarial scenarios in which we could  
 312 more precisely define the domain of

313 applicability of offset methods. For instance,  
 314 in the *Stepping Stone - Mountain* landscape,  
 315 geographic distance and environmental  
 316 distance were not strongly correlated, whereas  
 317 in the *Stepping Stone - Clines* and *Estuary -*  
 318 *Clines* they were. Additionally, the proportion  
 319 of mutations with monotonic frequency  
 320 gradients (i.e., allelic clines) underlying local  
 321 adaptation varied across the simulated  
 322 datasets (Lotterhos, 2023), which may also  
 323 affect offset performance. These simulations  
 324 also presented demographic scenarios in  
 325 which selection was confounded with genetic  
 326 drift or population genetic structure.  
 327 For each simulation, ten individuals were  
 328 randomly chosen per population for a total of  
 329 1000 individuals. Individual genotypes were  
 330 coded as counts of the derived allele. Alleles  
 331 with global minor allele frequency (MAF) <  
 332 0.01 were removed. Using all 100 populations,  
 333 population-level derived allele frequencies and  
 334 current environmental values were used as  
 335 input to train offset methods.

336 In addition to the 2250 simulated Wright-  
 337 Fisher datasets (225 levels \* 10 replicates), we  
 338 also included a non-Wright-Fisher case with  
 339 range expansion from three refugia and  
 340 secondary contact (Lotterhos 2023). This  
 341 simulation evolved variable degrees of  
 342 admixture across the landscape. Six  
 343 moderately polygenic environmental traits  
 344 ( $n_{\text{traits}} = 6$ ) were under selection from the  
 345 environment. Environments were based on six  
 346 weakly correlated environmental variables  
 347 taken from Bioclim environmental measures  
 348 of western Canada. The simulation evolved  
 349 local adaptation at all six traits with  
 350 unconstrained pleiotropy. For more details on  
 351 simulations, see (Lotterhos, 2023).

## 352 2.2 / Evaluation of Offset Methods

353 We investigated the performance of  
 354 five implementations of four genomic  
 355 offset methods (Table 1):  $GF_{\text{offset}}$ ,  
 356  $RDA_{\text{offset}}$ ,  $LFMM2_{\text{offset}}$ , and RONA.  
 357 While  $GF_{\text{offset}}$ ,  $RDA_{\text{offset}}$ , and  
 358  $LFMM2_{\text{offset}}$  can use multivariate  
 359 environmental data to train models,  
 360 RONA can only account for a single  
 361 environment at one time (Table 1).  
 362 Additionally, while  $GF_{\text{offset}}$  and RONA  
 363 do not apply correction for population  
 364 genetic structure,  $LFMM2_{\text{offset}}$  does by  
 365 default, and structure correction with  
 366  $RDA_{\text{offset}}$  is optional. We thus evaluate  
 367  $RDA_{\text{offset}}$  with (RDA-corrected) and  
 368 without (RDA-uncorrected) population  
 369 genetic structure correction (Table 1).  
 370 For additional specifics related to the  
 371 implementation of each offset method,  
 372 see Supplemental Note S1.1-S1.4 and  
 373 Fig. S1, Fig. S2, Fig. S3.

374 We varied construction of genomic  
 375 offset training datasets for each  
 376 replicate of the 1-, 2-, and 6-trait  
 377 simulations by varying the marker set  
 378 used in model training (Fig. 1A, Table  
 379 2; see *Q3* below). Each model was  
 380 trained using genetic and  
 381 environmental data from all 100  
 382 populations. The environmental vari-  
 383 ables used were only those imposing  
 384 selection pressure. We predict offset  
 385 from each model for each population to  
 386 all 100 within-landscape common  
 387 gardens from a full factorial *in silico*  
 388 reciprocal transplant design (Fig. 1B).  
 389 For each common garden, we quantified  
 390 offset model performance as the rank  
 391 correlation (Kendall's  $\tau$ ) between the

392 population mean fitness (averaged over  
 393 sampled individuals, Equation 3 in Lotterhos  
 394 2023) and projected population offset (Fig.  
 395 1C). Strong negative relationships between  
 396 fitness and predicted offset indicate higher  
 397 performance of the method (note y-axes of  
 398 Kendall's  $\tau$  are inverted within figures to  
 399 show more intuitive performance  
 400 relationships, Fig. 1C-11). We refer to the  
 401 preceding processing of data as the *Adaptive*  
 402 *Environment* workflow (Fig. 1, Table 2).

403 To explore the impact of the choice of  
 404 environmental variables used (see *Q5* below),  
 405 we used a workflow similar to the *Adaptive*  
 406 *Environment* workflow, except instead of  
 407 using only adaptive environmental variables,  
 408 we used additional non-adaptive (i.e.,  
 409 nuisance) environmental variables in training  
 410 and prediction (second row, Table 2). These  
 411 nuisance variables had relatively weak  
 412 correlation structure with adaptive  
 413 environments and each other (Fig. S4). We  
 414 refer to each of these nuisance levels by the  
 415 number of traits under selection and the  
 416 number of nuisance environments used (e.g.,  
 417 *1-trait 3-nuisance*). We refer to this workflow  
 418 as the *Nuisance Environment* workflow .

419 Finally, to contrast with within- landscape  
 420 evaluations, we explored predictive  
 421 performance of *Adaptive Environment* offset  
 422 models in novel environments that are beyond  
 423 the range of values of those used in training  
 424 (see *Q6* below). In these novelty cases, we use  
 425 11 common gardens, each progressively more  
 426 distant from the average environment used in  
 427 training (i.e., climate center) and evaluate  
 428 performance in each garden. We refer to this  
 429 workflow as the *Climate Novelty* workflow.  
 430 See Supplemental Note S3 and Fig. S5 for  
 431 details regarding the choice of environmental  
 432 values for novelty scenarios.

433 **2.3 / Study Questions**

434 *Q1 - Which aspects of the past*  
 435 *evolutionary history affect within-*  
 436 *landscape performance of offset*  
 437 *methods?*

438 For each offset method, we used a  
 439 fixed-effects type II ANOVA model to  
 440 test for significant differences in the  
 441 performance from 2-trait *Adaptive*  
 442 *Environment* models trained using *all*  
 443 markers using the following factors:  
 444 landscape (*Estuary - Clines, Stepping*  
 445 *Stone - Clines, Stepping Stone -*  
 446 *Mountain*), demography (five levels  
 447 describing population size and  
 448 migration patterns across the  
 449 landscape), genic level of architecture  
 450 (three levels from oligogenic to  
 451 polygenic), presence or absence of  
 452 pleiotropy, proportion of loci with clinal  
 453 allele frequencies (as defined in  
 454 Lotterhos, 2023), degree of local  
 455 adaptation ( $\Delta SA$ ), and common garden  
 456 ID. Specifically,

$$\begin{aligned}
 457 \quad Y_{ij} = & L_i + D_i + GL_i + P_i + \\
 458 & p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} + \\
 459 & p_{cNeut,Env2,i} + LA_{\Delta SA,i} + G_j \\
 460 & \hspace{15em} \text{(Eq. 1)}
 \end{aligned}$$

461 where  $Y_{ij}$  is the within-landscape  
 462 performance (Kendall's  $\tau$ ) of a single  
 463 method for garden  $j$  in simulation  $i$ ,  
 464 with factors for landscape ( $L$ ),  
 465 demography ( $D$ ), genic level ( $GL$ ),  
 466 presence of pleiotropy ( $P$ ), proportion  
 467 of QTN or neutral alleles with *temp*  
 468 clines (respectively  $p_{cQTN,t,i}$  and  
 469  $p_{cNeut,t,i}$ ), proportion of QTN or neutral  
 470 alleles with *Env2* clines (respectively

471  $p_{cQTN,Env2,i}$  and  $p_{cNeut,Env2,i}$ ), degree of local  
 472 adaptation ( $LA_{\Delta SA}$ ), and garden ID ( $G$ ). The  
 473 first four factors are illustrated in Fig. 1 of  
 474 Lotterhos (2023).

475 *Q2 - How is offset performance affected by the*  
 476 *proportion of clinal alleles in the data? (Q1B)*

477 Clinal alleles (i.e., alleles with monotonic  
 478 gradients in frequency across space) that  
 479 covary with environmental clines could be  
 480 weighted more heavily in offset models that  
 481 emphasize loci whose allele frequencies  
 482 explain significant variation across local  
 483 environmental values. Using 2-trait models  
 484 trained using *all* markers from the *Adaptive*  
 485 *Environment* workflow, we used an ANOVA  
 486 model (Eq. 2) to test the hypothesis that  
 487 clinal alleles differentially impact model  
 488 performance, independent from the other  
 489 factors from Eq. 1:

$$\begin{aligned}
 490 \quad Y_{ij} = & p_{cQTN,t,i} + p_{cNeut,t,i} + p_{cQTN,Env2,i} + \\
 491 & \hspace{15em} p_{cNeut,Env2,i} \\
 492 & \hspace{15em} \text{(Eq. 2)}
 \end{aligned}$$

493 The factors representing clinal alleles in Eq.  
 494 2 are the same as those in Eq. 1.

495 *Q3 - Is method performance driven by causal*  
 496 *loci or by genome-wide patterns of Isolation*  
 497 *By Environment? (Q2A)*

498 For each offset method and workflow, we  
 499 varied the set of input markers for 1-, 2- and  
 500 6-trait simulations that were used in training  
 501 to determine if performance of a method was  
 502 driven by properties of the evolutionary forces  
 503 shaping genotype-environment relationships:  
 504 1) *adaptive* markers (i.e., QTNs with effects  
 505 on at least one trait), 2) *neutral* markers  
 506 (SNPs on linkage groups without QTNs), and  
 507 3) *all* markers (union of *adaptive* and *neutral*

508 markers, as well as non-QTN markers  
 509 on the same linkage groups as QTNs).  
 510 Only loci that passed MAF filtering  
 511 were included in marker sets. If offset  
 512 performance is determined solely by  
 513 adaptive signals in genetic data, offsets  
 514 trained using *adaptive* markers should  
 515 have better performance than *all* or  
 516 *neutral* markers, and *all* markers should  
 517 have better performance than *neutral*  
 518 markers.

519 If the marker set has little impact on  
 520 offset performance, this could indicate  
 521 that offset methods are giving weight to  
 522 genome-wide signals present in the  
 523 data. Previously, some (e.g., Lachmuth,  
 524 Capblancq, Keller, et al., 2023; Lind et  
 525 al., 2024) have postulated that this  
 526 signal may be related to isolation by  
 527 environment ((IBE, i.e., when genetic  
 528 and environmental distances are  
 529 positively correlated, independent of  
 530 geographic distance; Wang &  
 531 Bradburd, 2014).

532 If IBE is driving patterns of offset  
 533 performance, we expect 1) performance  
 534 to be similar between offsets estimated  
 535 using *adaptive* markers and those  
 536 estimated using *neutral* markers; 2) a  
 537 greater proportion of variation in  
 538 performance to be explained by  $p_{cNeut}$   
 539 than  $p_{cQTN}$  (from Q2); 3) a strong,  
 540 positive relationship between  
 541 performance and  $LA_{\Delta SA}$ ; and 4) the  
 542 difference in IBE between two marker  
 543 sets to be positively correlated with the  
 544 difference in performance of two models  
 545 trained with those markers. We  
 546 measure IBE as the rank correlation  
 547 (Spearman's  $\rho$ ) between population  
 548 pairwise  $F_{ST}$  (Weir & Cockerham, 1984)

549 and Euclidean climate distance of adaptive  
 550 environmental variables.

551 *Q4 - What is the variation of model*  
 552 *performance across the landscape? (Q3a)*

553 Within a landscape, offset methods may  
 554 not have high predictive performance at every  
 555 site or every environment. Understanding  
 556 variability in the predictive performance of  
 557 offset models across the landscape is  
 558 particularly relevant when offsets are used for  
 559 restoration or assisted gene flow initiatives  
 560 (i.e., ranking sources for a given site). If  
 561 predictive performance is variable across the  
 562 landscape, this may limit the usefulness of  
 563 genomic offsets for such purposes even if  
 564 model performance is validated in one  
 565 common garden. Using the *Adaptive*  
 566 *Environment* workflow, we visualized  
 567 variation of 1- and 2-trait within-landscape  
 568 performance with boxplots for each common  
 569 garden for each method and landscape. To  
 570 understand if variation in predictive  
 571 performance was a function of the model  
 572 quality, we investigated the relationship  
 573 between a model's performance variability  
 574 (i.e., standard deviation across 100 common  
 575 gardens) and the model's median  
 576 performance.

577 *Q5 - How does the addition of non-adaptive*  
 578 *nuisance environments in training affect*  
 579 *performance? (Q2B)*

580 In practice, the environments imposing  
 581 selection are rarely known *a priori*.  
 582 Additionally, the inclusion of environmental  
 583 measures that are not correlated with the  
 584 main axes of selection may reduce model  
 585 performance compared to models trained  
 586 using only causal environments. To

587 investigate the sensitivity of offset  
 588 methods to environmental input we  
 589 compared *Adaptive Environment*  
 590 workflow models from 1-, 2-, and 6-trait  
 591 simulations – where only the adaptive  
 592 environment(s) are used in training (*0-*  
 593 *nuisance*) – to models from the  
 594 *Nuisance Environment* workflow  
 595 trained with the same data but with the  
 596 addition of nuisance environments (*N-*  
 597 *nuisance*, where  $N > 0$ ; Table 2).

598 We use nuisance environmental  
 599 variables from Lotterhos (2023) that  
 600 were real BioClim variables (*TSsd*,  
 601 *PSsd*, and *ISO*) taken from British  
 602 Columbia and Alberta, Canada, which  
 603 have minimal correlation with causal  
 604 environments and each other (Fig. S4).  
 605 These three nuisance environments  
 606 differ from previous implementations of  
 607 such variables (Láruson et al. 2022) in  
 608 that they are spatially autocorrelated  
 609 whereas nuisance environments in  
 610 Láruson et al. (2022) were not. For 1-  
 611 trait scenarios, *Env2* was also used as a  
 612 nuisance environmental variable.

613 If offset methods are unaffected by  
 614 the addition of nuisance environmental  
 615 variables, performance should not differ  
 616 between *0-nuisance* and *N-nuisance*  
 617 implementations. Finally, in  
 618 empirical settings the set of adaptive  
 619 environments are not known *a priori*.  
 620 We also explored whether GF would  
 621 rank adaptive environments higher  
 622 than nuisance environments using  
 623 weighted importance output from GF.

624 *Q6 - How well do offset models extrapolate to*  
 625 *novel environments outside the range of*  
 626 *environments used in training? (Q2C)*

627 Even if offset methods have high within-  
 628 landscape performance, this does not directly  
 629 address situations where future  
 630 environmental conditions are vastly different  
 631 from the environmental conditions used for  
 632 training (i.e., novel environments). If  
 633 performance decreases with increasing  
 634 environmental novelty relative to training  
 635 data, this raises questions about the utility of  
 636 genomic offsets for predicting 1) relative *in*  
 637 *situ* vulnerability of populations to future  
 638 climate change, and 2) the relative suitability  
 639 of populations to restoration sites that differ  
 640 drastically than those used in training.

641 To understand if offset performance  
 642 degrades with environmental novelty relative  
 643 to training data, we predicted offset to 10  
 644 novel environmental scenarios for the 1-, 2-,  
 645 and 6-trait simulations using the *Climate*  
 646 *Novelty* workflow (Table 2). The novel  
 647 environmental scenarios were a set of common  
 648 garden environments,  $z_E$ , extending outward  
 649 from the training populations and exceeding  
 650 values observed on the landscape for all  
 651 adaptive environmental variables  
 652 (Supplemental Note S3). We represent these  
 653 scenarios as standard deviations from the  
 654 center of environmental values used in  
 655 training:  $z_E \in \{1.72, 2.35, 2.74, 3.13, 3.53,$   
 656  $3.92, 4.31, 4.70, 5.09, 5.48, 5.88\}$ . Fitness in  
 657 novel environments was estimated assuming  
 658 that the phenotypic optimum continues to  
 659 have a linear relationship with the  
 660 environmental variable (Equation 3 in  
 661 Lotterhos 2023).



### 662 3 | Results

663 *Q1 - Which aspects of the past*  
 664 *evolutionary history affect within-*  
 665 *landscape performance of offset*  
 666 *methods?*

667 The ANOVA model (Eq. 1)  
 668 indicated that the degree of local  
 669 adaptation of the metapopulation  
 670 ( $LA_{\Delta SA}$ ) was the primary factor  
 671 influencing offset performance, followed  
 672 by common garden location,  
 673 demography, and landscape (Table S1;  
 674 Fig. S6). Within the simulations,  $LA_{\Delta SA}$   
 675 was impacted by pleiotropy, the  
 676 relative strength of selection, and  
 677 landscape, (Fig. S7; see also Figs. S2A,  
 678 S2B in Lotterhos, 2023), so there may  
 679 be some confounding among these  
 680 factors.

681 In line with the ANOVA model, the  
 682 performance of specific offset methods  
 683 generally increased with increasing  
 684  $LA_{\Delta SA}$  (Fig. 2), but there were some  
 685 interesting differences among methods.  
 686 For instance,  $GF_{\text{offset}}$ ,  $LFMM2_{\text{offset}}$ ,  
 687 RDA-uncorrected, and  $RONA_{\text{temp}}$  all  
 688 improved as  $LA_{\Delta SA}$  increased, while  
 689 RDA-corrected and  $RONA_{\text{Env2}}$  showed  
 690 relatively weaker relationships.

691 Across landscapes, offset methods  
 692 generally had higher performance in  
 693 *Stepping Stone - Clines* landscapes than  
 694 *Stepping Stone - Mountain* landscapes  
 695 (Fig. 2B) despite similar levels of  $LA_{\Delta SA}$   
 696 (Fig. 2A). Offset methods also generally  
 697 performed better in the two *Stepping*  
 698 *Stone* landscapes than the *Estuary -*  
 699 *Clines* landscape (Fig. 2B). However,  
 700 there were some interactions between

701 method and landscape (Fig. 2C). For  
 702 instance, RDA-corrected performed better in  
 703 the *Estuary - Clines* compared to the two  
 704 *Stepping Stones* landscapes, while the RDA-  
 705 uncorrected showed the opposite pattern:  
 706 performance was higher in the two *Stepping*  
 707 *Stones* landscapes compared to *Estuary -*  
 708 *Clines*.

709 The performance of methods was similar  
 710 across genic levels but increased slightly as  
 711 the number of QTNs underlying adaptation  
 712 became more polygenic (Fig. S8).  
 713 Additionally, while demography primarily  
 714 influenced population differentiation across  
 715 the landscape with little impact on  $LA_{\Delta SA}$   
 716 within simulations (Table S2 in Lotterhos  
 717 2023), migration breaks between populations  
 718 and latitudinal clines in population size  
 719 generally decreased offset performance for  
 720  $LFMM2_{\text{offset}}$ ,  $GF_{\text{offset}}$ , and RDA- uncorrected  
 721 (Fig. S9).

722 *Q2 - How is offset performance affected by the*  
 723 *proportion of clinal alleles in the data? (Q1B)*

724 The sum of squares from Eq. 1 indicated  
 725 that the proportion of clinal alleles did not  
 726 account for meaningful variation in offset  
 727 performance (Table S1). Even so, results from  
 728 an ANOVA model with just the proportion of  
 729 clinal loci as explanatory variables (Eq. 2)  
 730 indicated that  $p_{cNeut}$  accounted for 4.14-9.65  
 731 times the variation than did  $p_{cQTN}$  for  $GF_{\text{offset}}$ ,  
 732  $LFMM2_{\text{offset}}$ , and RDA-corrected. For  $GF_{\text{offset}}$   
 733 and RDA-uncorrected,  $p_{cNeut,Env2}$  accounted  
 734 for >16% of the sum of squares (Table S2,  
 735 Fig. S10).

736 Overall, relationships between  
 737 performance and  $p_{cNeut}$  (second column, Fig.  
 738 S11) had stronger relationships than between  
 739 performance and  $p_{cQTN}$  (first column, Fig.

740 S11). However, sometimes performance  
 741 increased with  $p_{cNeut}$  and sometimes it  
 742 decreased, depending on the method  
 743 (Fig. S11), indicating that each method  
 744 is differentially sensitive to clinal alleles  
 745 in the data. Ultimately, strong  
 746 population genetic structure along  
 747 environmental clines in 2-trait  
 748 simulations (Fig. S12) drove  
 749 relationships with  $p_{cNeut}$  (Fig. S13)  
 750 which in turn drove relationships with  
 751 performance (Fig. S14, Fig. S11).

752 *Q3 - Is method performance driven by*  
 753 *causal loci or by genome-wide patterns*  
 754 *of Isolation-By-Environment? (Q2A)*

755 Overall, 1- and 2-trait *Adaptive*  
 756 *Environment* models had relatively  
 757 similar performance among marker sets.  
 758 For instance, models trained using *all* or  
 759 *neutral* markers had similar  
 760 performance while models trained using  
 761 *adaptive* markers performed slightly  
 762 higher than the other sets. The median  
 763 increase in performance from *adaptive*  
 764 compared to *all* or *neutral* models was  
 765 less than 3%. In total, using *adaptive*  
 766 markers outperformed 68% of models  
 767 using *neutral* markers and 67% of  
 768 models using *all* markers, while 74% of  
 769 models using *all* markers outperformed  
 770 *neutral* models (Fig. 3A-C). For RDA-  
 771 corrected the *neutral* markers  
 772 performed slightly better than either  
 773 *adaptive* or *all* markers in 2-trait  
 774 evaluations (Fig. 3E). *Adaptive*  
 775 markers from 6-trait evaluations  
 776 provided varied performance  
 777 advantages across methods (Fig. 4).

778 The *adaptive* marker sets had relatively  
 779 elevated levels of *IBE* compared to sets of  
 780 *neutral* or *all* markers in 1- and 2-trait  
 781 simulations, but levels of *IBE* were  
 782 nonetheless quite similar between marker sets  
 783 (Fig. S15). Consequently, performance of  
 784 models trained with *adaptive* markers  
 785 generally had stronger relationships with *IBE*  
 786 than  $LA_{\Delta SA}$  but this was not the case for  
 787 models trained with either *all* or *neutral*  
 788 markers (Fig. S16).

789 Intriguingly, levels of *IBE* found within a  
 790 landscape (Fig. S17A) did not correspond to  
 791 the degree of  $LA_{\Delta SA}$  that developed (Fig.  
 792 S17B). Even so, while *IBE* was generally  
 793 unrelated to  $LA_{\Delta SA}$  across all simulations,  
 794 there were generally positive relationships  
 795 between *IBE* and  $LA_{\Delta SA}$  when controlling for  
 796 the number of traits and differences in  
 797 strengths of selection (Fig. S18). As such, *IBE*  
 798 from *all* markers explained very little  
 799 variation in performance when added as a  
 800 factor to the ANOVA model from Eq. 1 (SC  
 801 02.02.01), but accounted for some variation in  
 802 ANOVA models with only  $LA_{\Delta SA}$  and *IBE* as  
 803 explanatory variables (0-34% for *IBE* vs 0-  
 804 74% for  $LA_{\Delta SA}$ ; Table S3). Except for RONA,  
 805 the differences in performance between two  
 806 models trained with different marker sets was  
 807 generally unrelated to the differences in *IBE*  
 808 between the two marker sets used to train the  
 809 models (Fig. S19).

810 Together these results indicate that while  
 811 higher degrees of local adaptation may lead to  
 812 increased levels of *IBE* in the genome, the  
 813 signal of *IBE* of input markers generally has  
 814 minimal and varied impact on performance  
 815 differences for the scenarios evaluated here.  
 816 Alternatively, the levels of *IBE* present in the  
 817 simulated genomes may exceed a minimum  
 818 threshold of *IBE*, beyond which differences in

819 performance between marker sets are  
820 minimized.

821 *Q4 - What is the variation of model*  
822 *performance across the landscape?*  
823 *(Q3a)*

824 All 1- and 2-trait models exhibited  
825 variation in the predictive performance  
826 across gardens within a landscape, from  
827 essentially no predictive performance to  
828 very high predictive performance (Fig.  
829 S20, Fig. S21, Fig. S22, Fig. S23).  
830 Variation in performance was also  
831 observed for 6-trait models (Fig. 4).

832 While there was variability in  
833 predictive performance of 1- and 2-trait  
834 models within each landscape, in many  
835 cases the best performing models had  
836 the lowest levels of performance  
837 variation (Figs. S24, S25, S26).  
838 Ultimately we found no strong indicator  
839 for predicting when a model will be  
840 highly variable. Indeed, while  
841 performance generally increased with  
842  $LA_{\Delta SA}$  (Fig. 2), variability in  
843 performance was not strongly related to  
844 the variability in deme-level  $LA$  on the  
845 landscape (Figs. S27, S28, S29). Despite  
846  $LA_{\Delta SA}$  driving performance more  
847 generally (from Q1), this indicates that  
848 variation in model performance across  
849 the landscape is not strongly driven by  
850 metapopulation levels of, nor deme-  
851 level variation in,  $LA$ .

852 *Q5 - How does the addition of non-*  
853 *adaptive nuisance environments in*  
854 *training affect performance? (Q2B)*

855 Training offset models with the  
856 addition of non-adaptive nuisance

857 environmental variables generally reduced  
858 offset method performance (Fig. 5). This  
859 decline was most dramatic for offset trained  
860 on 1-trait simulations (Fig. 5A) compared to  
861 the decline observed for 2-trait (Fig. 5B) and  
862 6-trait (Fig. 5C) simulations. The only  
863 instances for which median performance did  
864 not decrease monotonically with nuisance  
865 level were for 2-trait simulations evaluated  
866 with  $GF_{\text{offset}}$  (Fig. S30).

867 Overall, landscape had the most influence  
868 over performance differences due to non-  
869 adaptive nuisance environments (Fig. S30),  
870 whereas there was little difference across  
871 other simulation parameters (not shown  
872 except in SC 02.02.06). Even so, *adaptive*  
873 markers seemed to provide some advantages  
874 in the presence of nuisance environments,  
875 particularly for 1-trait datasets where the  
876 advantages were more substantial compared  
877 to 2-trait datasets (Fig. S31, Fig. S32).

878 In some cases, the rankings of weighted  
879 environmental importance output from  $GF$   
880 ranked nuisance variables higher than at least  
881 one adaptive environment (Table S4). Across  
882 1- and 2-trait *N-nuisance* models trained with  
883 *all* markers,  $GF$  incorrectly ranked  
884 environmental drivers in 26.9% (133/495) of  
885 the cases. Rankings improved somewhat for  
886 models trained with *adaptive* markers,  
887 incorrectly ranking environmental variables in  
888 20.6% (102/495) of the cases (Table S4).

889 *Q6 - How well do offset models extrapolate to*  
890 *novel environments outside the range of*  
891 *environments used in training? (Q2C)*

892 The datasets that had the greatest within-  
893 landscape performance (i.e., those with higher  
894 levels of  $LA_{\Delta SA}$ ) were also those that  
895 experienced the steepest decline in

896 performance with increasing climate  
 897 novelty (red shade, Fig. 6).  
 898 Importantly, declines in performance  
 899 for datasets with greater  $LA_{\Delta SA}$  were  
 900 not due to instances where all  
 901 populations had zero fitness (and thus  
 902 performance was undefined and  
 903 manually set to 0; Supplemental Note  
 904 S4, Fig. S33). Despite little change in  
 905 the median performance for datasets  
 906 with low levels of LA, most performance  
 907 scores from these datasets were below  
 908 Kendall's  $\tau=0.5$ , and therefore had  
 909 little predictive value in novelty  
 910 scenarios.

911 Advantages of *adaptive* marker sets  
 912 were much less prevalent across  
 913 methods for *Climate Novelty* scenario  
 914 performance than either *Adaptive*  
 915 *Environment* or *Nuisance Environment*  
 916 scenarios (Fig. S34).

## 917 4 | Discussion

918 Solutions are needed to mitigate the  
 919 negative impacts of global change on  
 920 biodiversity. In the last decade,  
 921 genomic offset methods have been  
 922 identified as a complement to other  
 923 ecological forecasting models because  
 924 they incorporate intraspecific variation  
 925 (Keller & Fitzpatrick, 2015; Capblancq  
 926 et al., 2020; Rellstab et al., 2021). Our  
 927 evaluations show that offset methods  
 928 are differentially impacted by both the  
 929 evolutionary history of sampled  
 930 populations as well as the decisions  
 931 made during model training. Our  
 932 analyses emphasize the importance of  
 933 sampling locally adapted populations,

934 identifying the drivers underlying  
 935 environmental selection pressures *a priori*,  
 936 and restricting offset projections to climates  
 937 similar to those used in training. Below, we  
 938 discuss the implications of these findings  
 939 towards restoration, conservation, and the  
 940 management of biodiversity.

### 941 4.1 / *The importance of local adaptation*

942 A basic assumption of genomic offset  
 943 methods is that the sampled populations are  
 944 adapted to their local environment (Rellstab  
 945 et al., 2016, 2021), but this assumption has  
 946 not been formally tested. Our analyses show  
 947 that indeed the degree of local adaptation  
 948 ( $LA_{\Delta SA}$ ) is one of the primary factors that  
 949 determine model performance for most  
 950 methods. A value of  $LA_{\Delta SA} \sim 0.5$  indicates  
 951 that fitness in demes is on average 50% higher  
 952 in sympatry than allopatry. Values of  $LA_{\Delta SA}$   
 953 represent the average deme-level magnitudes  
 954 of  $LA_{\Delta HA}$  and  $LA_{\Delta LF}$  across the  
 955 metapopulation (Blanquart et al., 2013).  
 956 Previous metaanalyses of studies measuring  
 957 local adaptation of natural populations have  
 958 used different measures of  $LA$  from the ones  
 959 we calculate here, but do show that some  
 960 species evolve large fitness differences among  
 961 populations (Hereford, 2009; Leimu &  
 962 Fischer, 2008). Given the prevalence of  $LA$   
 963 found previously (Hereford, 2009; Leimu &  
 964 Fisher, 2010), we may therefore expect some  
 965 genomic offset methods to do reasonably well  
 966 when local adaptation in the metapopulation  
 967 is high (i.e., when  $LA_{\Delta SA} > 0.5$ ).

### 968 4.2 / *The importance of the signals* 969 *within genomic marker sets*

970 Because of the assumption that locally  
 971 adapted populations will be necessary for

972 satisfactory model performance, initial  
 973 implementations of genomic offset  
 974 models focussed on putatively adaptive  
 975 markers where this signal may be  
 976 strongest (Keller & Fitzpatrick, 2015;  
 977 Rellstab et al., 2016). More recently,  
 978 investigators have varied the set of  
 979 markers used to train models but have  
 980 found little influence on performance  
 981 (Fitzpatrick et al., 2021; Lachmuth,  
 982 Capblancq, Keller, et al., 2023; Láruson  
 983 et al., 2022; Lind et al., 2024). Our  
 984 results are similar to previous  
 985 investigations, finding that the *adaptive*  
 986 marker sets provide minimal advantage  
 987 over *all* or *neutral* marker sets, but not  
 988 universally or by great margins.

989 One hypothesis put forth as to why  
 990 adaptive marker sets perform similarly  
 991 to all markers is that genome-wide data  
 992 captures sufficient signatures of IBE  
 993 (Lachmuth, Capblancq, Keller, et al.,  
 994 2023; Lind et al., 2024). Our analysis  
 995 found weak positive relationships  
 996 between performance and levels of *IBE*  
 997 within marker sets. Even so, and except  
 998 for RONA, there were no universal  
 999 relationships within methods between  
 1000 the difference in *IBE* of marker sets and  
 1001 the difference in performance of the  
 1002 models trained with these markers.

1003 While we found little impact of  
 1004 levels of *IBE* on overall performance,  
 1005 the way in which we measured IBE may  
 1006 have masked causative relationships.  
 1007 For instance, in our measure of IBE we  
 1008 correlated environmental distance with  
 1009 pairwise  $F_{ST}$ . In doing so, our measure  
 1010 of IBE distills genetic distance down to  
 1011 a single value from a large number of  
 1012 loci, and gives less weight to loci with

1013 rare alleles. In future studies, creating a  
 1014 marker set by ranking loci by single-locus  
 1015 measures of IBE offers another opportunity to  
 1016 understand the impact of IBE on  
 1017 performance. Such marker sets could be used  
 1018 to compare to performance from putatively  
 1019 adaptive marker sets or marker sets composed  
 1020 of all or random loci. Empirical datasets will  
 1021 also be able to specifically address  
 1022 geographical distances while quantifying IBE  
 1023 (e.g., Bradburd et al., 2013).

1024 While measures of IBE are one signal  
 1025 remaining to be explored in future analyses,  
 1026 the proportion of clinal neutral loci within  
 1027 marker sets was shown to impact  
 1028 performance, sometimes being positively  
 1029 related to performance and sometimes  
 1030 negatively depending on the context. These  
 1031 and other signals within data that could  
 1032 improve or mislead offset models also warrant  
 1033 further investigation.

### 1034 ***4.3 / The importance of adaptive*** 1035 ***environmental variables***

1036 In empirical settings, the environ- mental  
 1037 drivers of local adaptation are rarely known *a*  
 1038 *priori*. Even so, our results emphasize the  
 1039 importance of identifying these variables  
 1040 before training offset models, as there were  
 1041 often declines in performance between models  
 1042 trained using only adaptive environmental  
 1043 variables (*0-nuisance*) and those trained using  
 1044 additional non-adaptive nuisance  
 1045 environmental variables (*N-nuisance*).

1046 The importance of identifying these  
 1047 selective environments may be particularly  
 1048 germane to two general empirical scenarios. In  
 1049 the first empirical scenario., sparsely sampling  
 1050 an environmentally heterogeneous range may  
 1051 enrich genetic signals (e.g., coincident

1052 population structure) most correlated  
1053 to environmental variables that  
1054 maintain a gradient across this extent,  
1055 and miss signals relevant to more local  
1056 scales. In the second empirical scenario,  
1057 identifying the environmental variables  
1058 underlying selection is particularly  
1059 important when a specific genomic  
1060 offset method is ill-suited to  
1061 differentiate importance among input  
1062 variables. For instance, RDA (and  
1063 therefore RDA<sub>offset</sub>) assumes that the  
1064 environmental variables used to build  
1065 models are not collinear; (as  
1066 implemented here; Capblancq &  
1067 Forester, 2021; Legendre & Legendre,  
1068 2012). Because of this, empirical  
1069 datasets must be limited to a subset of  
1070 available environmental measures. The  
1071 process of excluding environmental  
1072 variables in this way may omit signals  
1073 of adaptive drivers (particularly when  
1074 true drivers are not well measured), or  
1075 perhaps incorporate environmental  
1076 variables that do not coincide with  
1077 drivers of selection. In these cases,  
1078 performance is likely to decline. As  
1079 such, this may indicate that methods  
1080 such as RDA<sub>offset</sub> are likely to perform  
1081 worse in, or less uniformly across,  
1082 realistic empirical settings than what  
1083 our current findings suggest.

1084 On the other hand, users of GF may  
1085 be tempted to include a large number  
1086 of environmental variables in training,  
1087 hoping that GF can accurately  
1088 attribute the correct environmental  
1089 variation to adaptive genetic structure.  
1090 Our results show that it is not  
1091 necessarily the case that GF will give  
1092 the highest importance values to the

1093 true adaptive environmental variables.  
1094 Indeed, weighted feature importance scores  
1095 from GF models still incorrectly ranked the  
1096 adaptive environments below neutral  
1097 environments in 20%-27% of the datasets,  
1098 depending on which marker set was used.  
1099 These importance values ultimately affect the  
1100 model predictions. Including all available  
1101 environmental variables may therefore  
1102 negatively impact GF<sub>offset</sub> performance, and  
1103 could have weakened overall performance in  
1104 previous empirical evaluations that used a  
1105 large number of environmental measures in  
1106 training (e.g., Lind et al., 2024).

1107 There are some differences between the  
1108 nuisance environmental variables  
1109 implemented here and those that have been  
1110 implemented previously. For instance,  
1111 Láruson et al. (2022) created nuisance  
1112 variables by randomly sampling a  
1113 multivariate normal distribution. In contrast  
1114 to findings here, Láruson et al. (2022) found  
1115 that model performance was relatively  
1116 unaffected with the addition of nuisance  
1117 variables. The minimal influence of nuisance  
1118 variables on performance found by Láruson et  
1119 al. (2022) may differ from the performance  
1120 declines reported here because the nuisance  
1121 variables we used were spatially  
1122 autocorrelated, while those from Láruson et  
1123 al. (2022) were not. Inclusion of nuisance  
1124 variables that are spatially autocorrelated  
1125 may mislead offset models more generally  
1126 than variables with little spatial  
1127 autocorrelation because of the spurious  
1128 relationship between environmental structure  
1129 and genetic structure.

1130 **4.4 / The effect of environmental**  
 1131 **novelty**

1132 While within-landscape perfor-  
 1133 mance generally increased with  $LA_{\Delta SA}$ ,  
 1134 the datasets with the greatest levels of  
 1135  $LA_{\Delta SA}$  were also the datasets where  
 1136 performance declined most readily with  
 1137 climate novelty. This occurred because  
 1138 locally adapted metapopulations were  
 1139 under strong selection to be fine-tuned  
 1140 to their environment, and as a result  
 1141 most individuals suffered severe fitness  
 1142 declines with environmental change. In  
 1143 contrast, less locally adapted  
 1144 metapopulations were under weaker  
 1145 selection, and suffered less steep fitness  
 1146 declines with environmental change.  
 1147 This result highlights an interesting  
 1148 paradox: offset methods that have the  
 1149 highest performance in common garden  
 1150 transplants under current climates  
 1151 (because of strong local adaptation)  
 1152 may have the lowest performance in  
 1153 predicting “genomic vulnerability” as  
 1154 the range of climate variables become  
 1155 more novel compared to the ranges used  
 1156 in training the model.

1157 Thus, genomic offset models are  
 1158 likely ill-suited for estimating fitness  
 1159 ranks of populations in environments  
 1160 that differ drastically from those used  
 1161 to train the models themselves. This is  
 1162 particularly relevant for applications of  
 1163 offset methods that attempt to estimate  
 1164 *in situ* risk of climate change to years or  
 1165 climate scenarios where the  
 1166 environment is expected to be  
 1167 increasingly novel. While climate  
 1168 novelty is often measured with respect  
 1169 to historical variability (e.g., Lotterhos

1170 et al., 2021; Mahony et al., 2017; Williams et  
 1171 al., 2007), indices of local climate change  
 1172 indicate that local environments in terrestrial  
 1173 systems could experience change in excess of  
 1174 three standard deviations relative to historic  
 1175 values (Williams et al., 2007). Similar indices  
 1176 in marine systems indicate potential for even  
 1177 greater novelty (Lotterhos et al., 2021). We  
 1178 observed performance declines below the  
 1179 analogous  $z_E=3.13$  *Climate Novelty* scenario,  
 1180 indicating offset predictions will likely be  
 1181 inaccurate in many real-world climate change  
 1182 predictions. These issues are also germane to  
 1183 measures derived from offset values  
 1184 (Gougherty et al., 2021; Lachmuth,  
 1185 Capblancq, Keller, et al., 2023; Lachmuth,  
 1186 Capblancq, Prakash, et al., 2023), which  
 1187 currently do not consider the degree of  
 1188 climate novelty in the prediction (but see  
 1189 DeSaix et al., 2022).

1190 Our results present a best-case scenario for  
 1191 predicting performance in novel  
 1192 environments, as in many cases there will be  
 1193 biological reasons as to why climate-fitness  
 1194 relationships will differ in future  
 1195 environments from relationships measured  
 1196 within the contemporary climate space (see  
 1197 Fig. 5 in Capblancq et al., 2020). For the  
 1198 simulations here, the relationship between  
 1199 contemporary and novel environments with  
 1200 fitness was the same.

1201 **4.5 / Genomic offsets in practice**

1202 Our evaluations show that genomic offset  
 1203 methods hold promise for predicting  
 1204 maladaptation to environmental change  
 1205 within a historical baseline, in  
 1206 metapopulations that evolve strong local  
 1207 adaptation. However, our analyses also  
 1208 emphasize the limits of these methods in some

1209 systems or scenarios. In practice,  
1210 species that are locally adapted to  
1211 measurable environmental variables  
1212 will be best suited for offset methods  
1213 when predicting the relative  
1214 performance of populations in a  
1215 contemporary common garden, but  
1216 paradoxically these species may be least  
1217 suited to using these methods to predict  
1218 their vulnerability to novel climates.

1219 Together, these results indicate that  
1220 some genomic offset methods may be  
1221 suited to guide initiatives such as near-  
1222 term assisted gene flow, where targeted  
1223 restoration sites within a species range  
1224 have climates that are similar to those  
1225 used to train offset models. Even so, our  
1226 results also show that the performance  
1227 of these methods are often variable  
1228 across a landscape, indicating that high  
1229 performance at one site does not mean  
1230 the offset model will perform well at  
1231 another. While genomic offset methods

1232 may be suitable for assisted gene flow  
1233 initiatives, they may be less suited for assisted  
1234 migration programs where populations are  
1235 moved outside of their native range and  
1236 environments differ from training data.

1237 Before genomic offsets can be incorporated  
1238 into management plans, considerable thought  
1239 must be put into the sensitivity of model  
1240 outcomes from input data (Lind et al., 2024),  
1241 the uncertainty inherent in environmental or  
1242 climate forecasts (Lachmuth, Capblancq,  
1243 Keller, et al., 2023), as well as the degree of  
1244 novelty of future climates (DeSaix et al., 2022,  
1245 this study). While accurate predictions are  
1246 limited for novel climates of the future, these  
1247 offset methods could still be used to guide  
1248 management in the intervening time in a  
1249 stepwise manner where experiments can be  
1250 used to validate model performance in  
1251 practice. Using simulations tailored to the life  
1252 history of target species also presents a  
1253 promising avenue to understand limitations of  
1254 these methods for specific management cases.



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1257 bodies did not have any role in the design of the study, analysis, interpretation of results, or  
1258 in writing of the manuscript.

## 1259 **Author Contributions**

1260 KEL received funding. KEL and BML conceptualized the project and methodology. With  
1261 input, editing, and feedback from KEL, BML wrote code to train and evaluate offset models,  
1262 created figures, curated coding and records for archiving, and wrote the manuscript.

## 1263 **Conflict of Interests**

1264 The authors declare no conflicts of interest.

## 1265 **Data Availability**

1266 We reference the analysis code in the text of our documents by designating Supplemental  
1267 Code (SC) using a directory numbering system from our servers (as opposed to the order  
1268 listed in the manuscript). Supplemental Code includes both executable scripts (\*.R, \*.py) as  
1269 well as jupyter notebooks (\*.ipynb). For example, for Script 3 in Directory 1, we refer to SC  
1270 01.03; for Notebook 5 in Subfolder 3 of Directory 2, we will refer to SC 02.03.05. Each  
1271 Directory will be archived on Zenodo.org and include a citation below, which will also link  
1272 to the GitHub repository. Notebooks are best viewed within a local jupyter or jupyter lab  
1273 session (to enable cell output scrolling / collapsing), but can also be viewed at  
1274 nbviewer.jupyter.org using the web link in the archive's README on GitHub. Analyses  
1275 were carried out primarily using python v3.8.5 and R v3.5.1 and v4.0.3. Exact package and  
1276 code versions are available at the top of each notebook. More information on coding  
1277 workflows and coding environments can be found in Supplemental Note S1 and Supplemental  
1278 Note S2.

1279  
1280 All directories, notebooks, and scripts can be found on GitHub, and will be archived on  
1281 Zenodo.

1282 <https://github.com/ModelValidationProgram/MVP-offsets>

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Figures for:

# The limits of predicting maladaptation to future environments with genomic data

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**Running Title:** Limits of genomic offsets

**Keywords:** genomic offset, environmental change, climate change, assisted gene flow, genomic forecasting, restoration

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<b>Method</b>	<b>abbr.</b>	<b>Multivariate?</b>	<b>Structure correction?</b>
Gradient Forests <sup>1</sup>	GF <sub>offset</sub>	Yes	No
Redundancy Analysis <sup>2</sup> with population structure correction	RDA-corrected	Yes	Yes, with axes loadings from PCA*
Redundancy Analysis <sup>2</sup> without population structure correction	RDA-uncorrected	Yes	No
Latent factor mixed model from Landscape and Ecological Association Studies R package <sup>3</sup>	LFMM2 <sub>offset</sub>	Yes	Yes, with latent factors
Risk Of Non-Adaptedness <sup>4</sup>	RONA	No	No

\* principal component analysis

**Table 1** Genomic offset methods used for evaluation. Genomic offset methods differ in their capability to use multivariate environmental data in training as well as whether a correction for population genetic structure is applied. Superscripts apply to the following reference citations: 1 - Fitzpatrick & Keller, 2015; 2 - Capblancq & Forester, 2021; 3 - Gain & François, 2021; 4 - Rellstab et al., 2016.



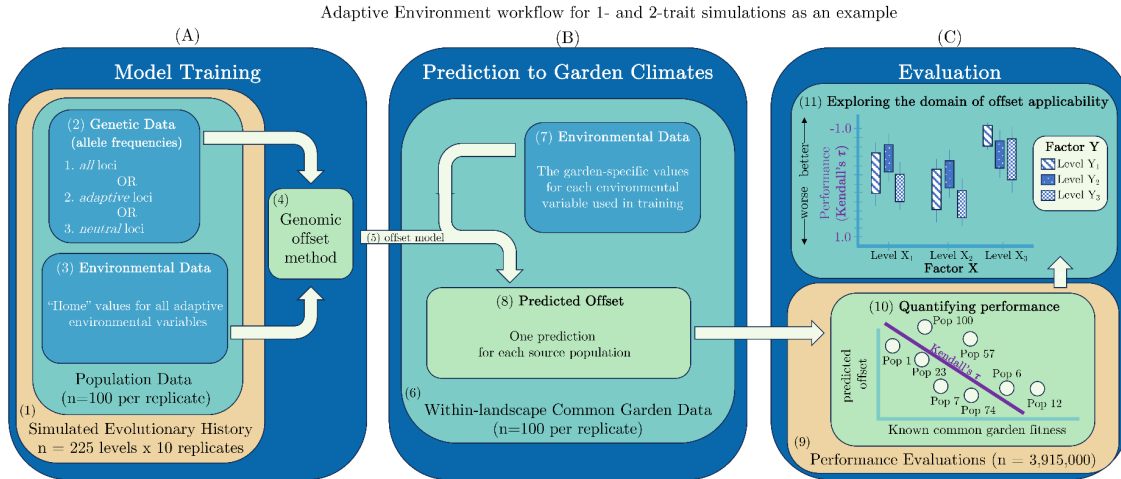
Workflow	$n_{\text{traits}}$	(1) Simulations Levels (replicates per level)	(3, 7) Environmental Data	Training and Prediction?	(6) Within-landscape Evaluation?	(9) Total Performance Evaluations
Adaptive Environment (AE)	1-trait	45 (10)	temp	Yes	Yes	675,000
	2-trait	180 (10)	temp + Env2			3,240,000
	6-trait	1 (1)	MAT + MTwetQ + MTDQ + PDM + PwarmQ + PWM			3,000
Nuisance Environment (NE)**	1-trait	45 (1)	[AE <sub>1-trait</sub> environments + Env2] +/- [ISO + PSsd] +/- [TSsd]	Yes	Yes	175,500
	2-trait	180 (1)	[AE <sub>2-trait</sub> environments + ISO + PSsd] +/- [TSsd]			432,000*
	6-trait	1 (1)	[AE <sub>6-trait</sub> environments + ISO + PSsd + TSsd]			1,200*
Climate Novelty (CN) <sup>†</sup>	1-trait	45 (10)	AE <sub>1-trait</sub> environments	Prediction only (using AE training models)	No	64,800*
	2-trait	180 (10)	AE <sub>2-trait</sub> environments			259,200*
	6-trait	1 (1)	AE <sub>6-trait</sub> environments			144*

\* excludes RONA

\*\* The set of population values for each unique nuisance environment were the same across traits and landscapes

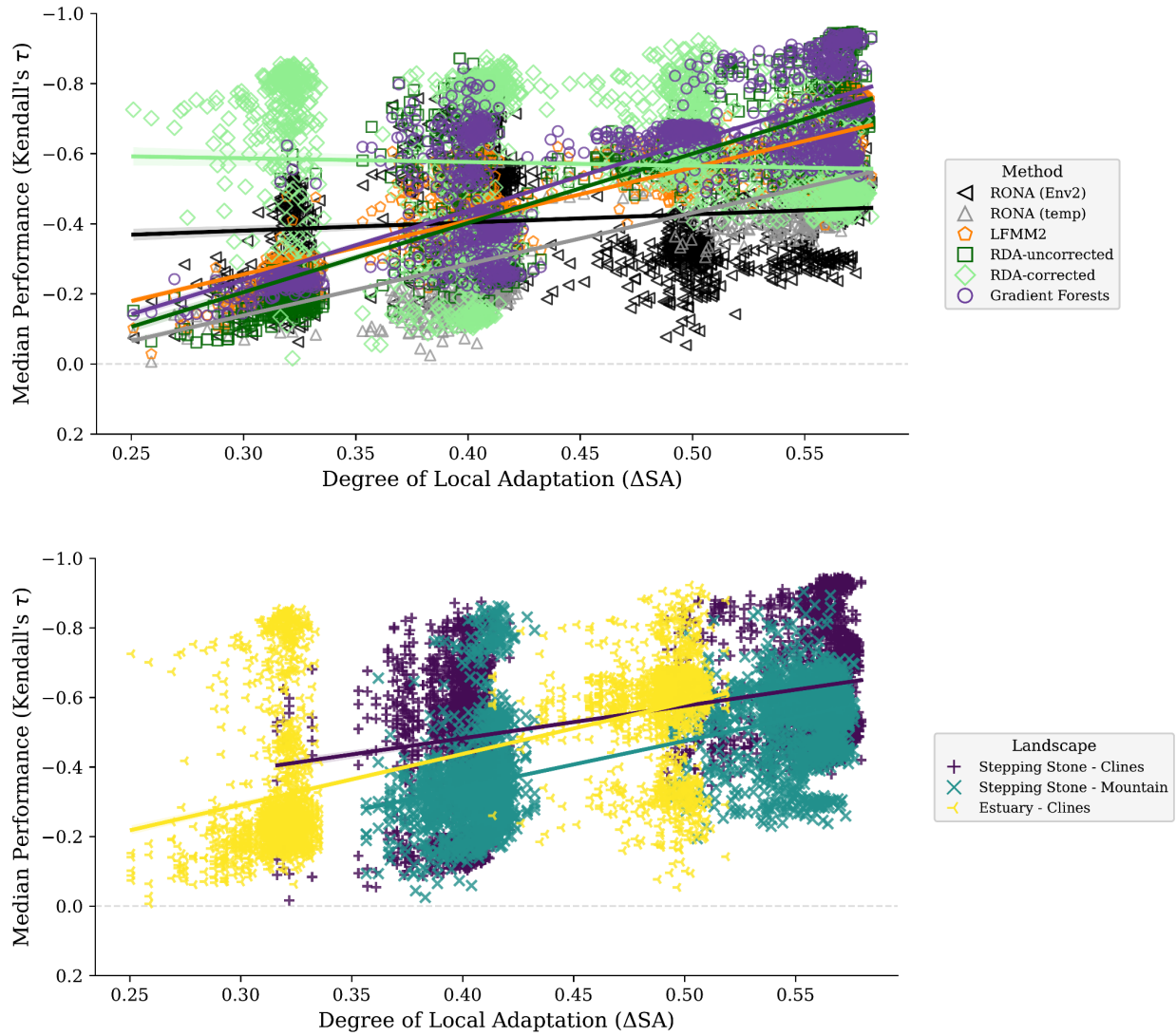
<sup>†</sup> includes evaluation of climate center and eleven *Climate Novelty* scenarios

**Table 2** Workflows used to process simulation data for the evaluation of genomic offset methods. Numbers given in column names refer to locations in schematic of Fig. 1. The *Adaptive Environment* workflow processes all population data from 1- and 2-trait (example shown in Fig. 1) as well as 6-trait simulations using only adaptive environmental variables in training, and evaluates performance in each garden on the metapopulation landscape. The *Nuisance Environment* workflow processes 1-, 2-, and 6-trait simulations similarly to the *Adaptive Environment* workflow, except in addition to adaptive environmental variables used in training, non-adaptive (i.e., nuisance) environmental variables are also used - each bracketed set of environmental variables indicate a distinct nuisance level (e.g., “1-trait 1-nuisance” = [AE<sub>1-trait</sub> environments + Env2] and “1-trait 4-nuisance” = [AE<sub>1-trait</sub> environments + Env2 + ISO + PSsd + TSsd]). The *Climate Novelty* workflow uses trained models from the *Adaptive Environment* workflow (Fig. 1A-5) and evaluates offset in 11 novel environments relative to the range of environments used in training. See Supplemental Note S3 for details regarding the choice of *Climate Novelty* environmental values and visualizations of climate data in principal component space. See Supplemental Notes S1-S2 for descriptions of coding workflows. Counts of evaluations were tabulated in SC 02.10.01.

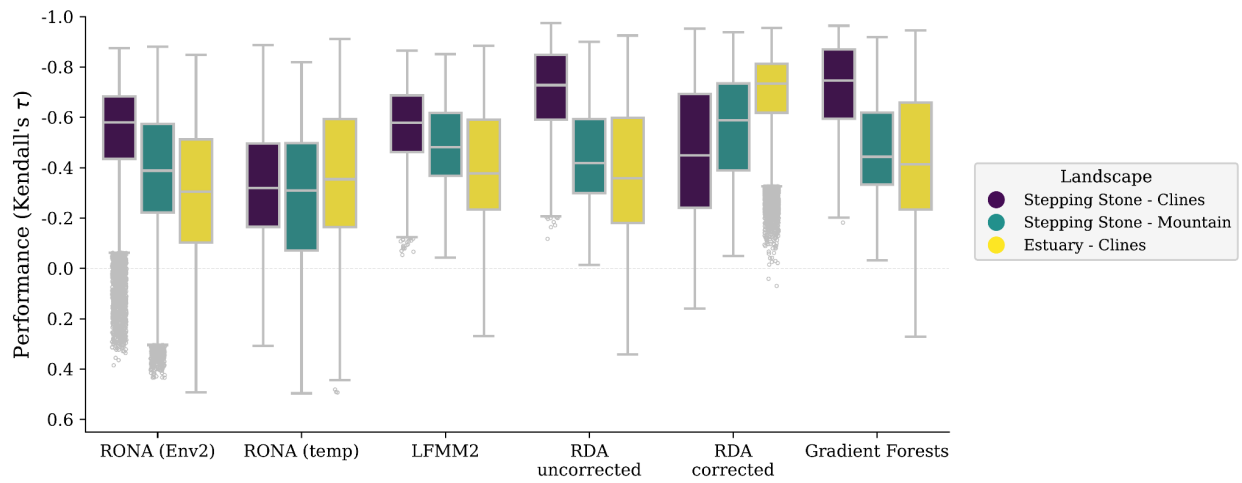


**Figure 1** Analysis of 1-, 2-, and 6-trait simulations included three main phases: A) model training, B) model prediction, and C) evaluation of models. The *Adaptive Environment* workflow is shown as an example of the processing of 1- and 2-trait simulation data for genomic offset evaluation. In total, three general workflows are used to evaluate genomic offset methods (Table 2). Subpanels of this schematic are numbered for referencing in Table 2 and the main text.

(Fig. 2)

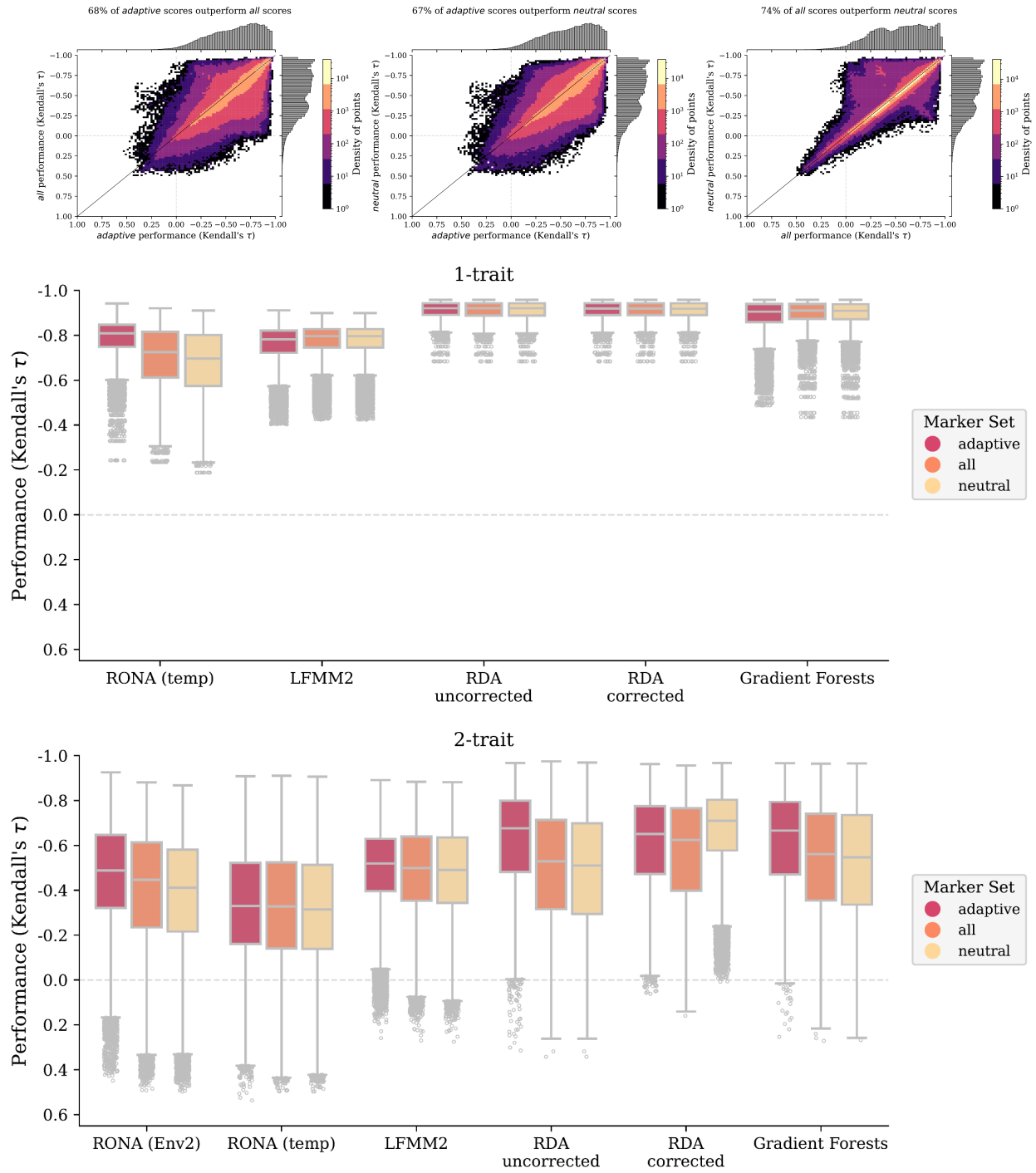


(Fig. 2 continued)

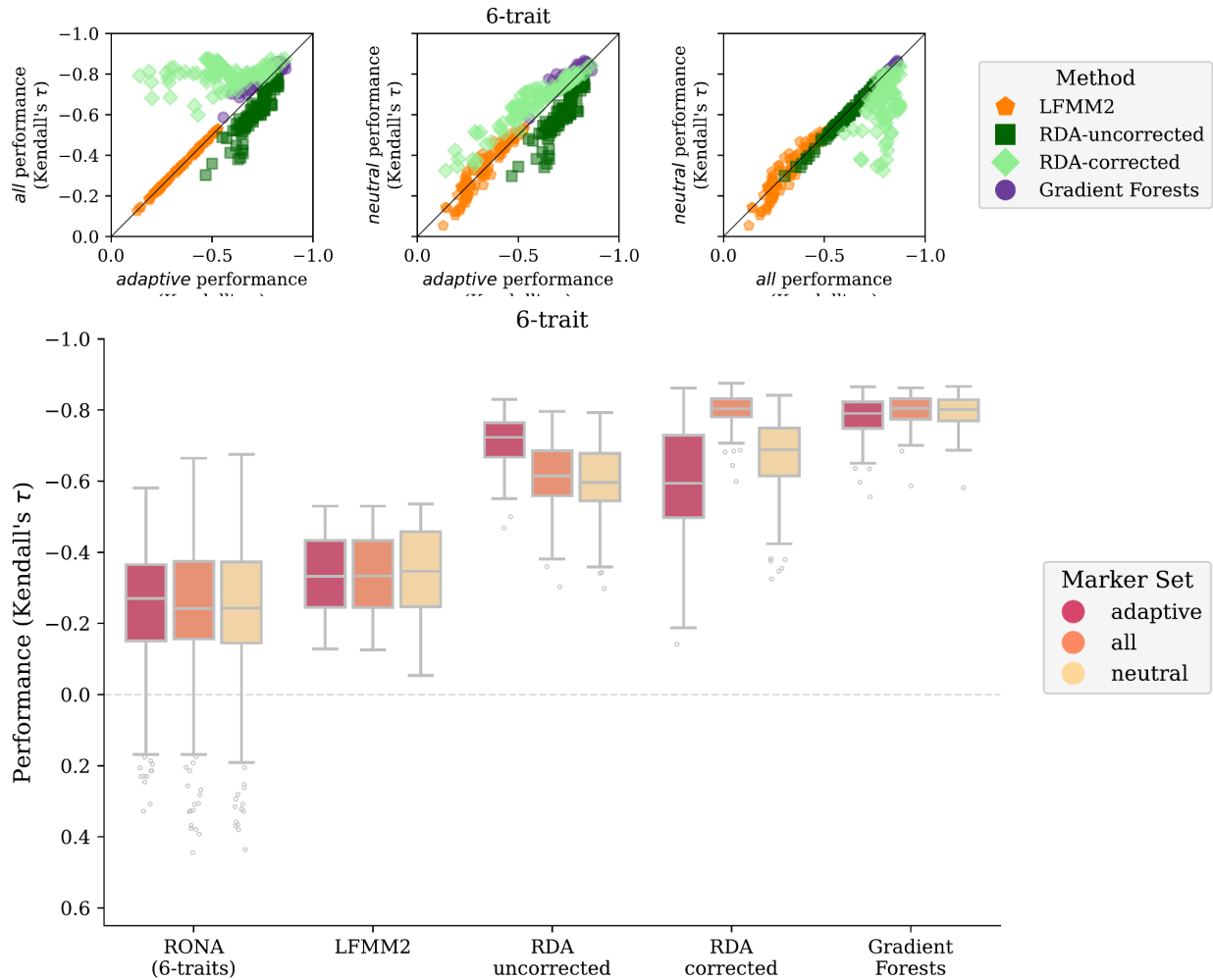


**Figure 2** Predictive performance of genomic offset models (y-axes) is driven by the degree of local adaptation (A) and the spatial patterns of adaptive environments across the landscape (B, C). For each model, a median value from performance scores from 100 common gardens is shown for A and B; C shows scores across all common gardens for each model (note that y-axes are inverted, as more negative values have higher performance). Data included in these figures was processed through the *Adaptive Environment* workflow but only includes models trained using 2-trait simulations and *all* loci. Code to create (A) and (B) can be found in SC 02.02.02; code to create (C) can be found in SC 02.02.01.

(Fig. 3)

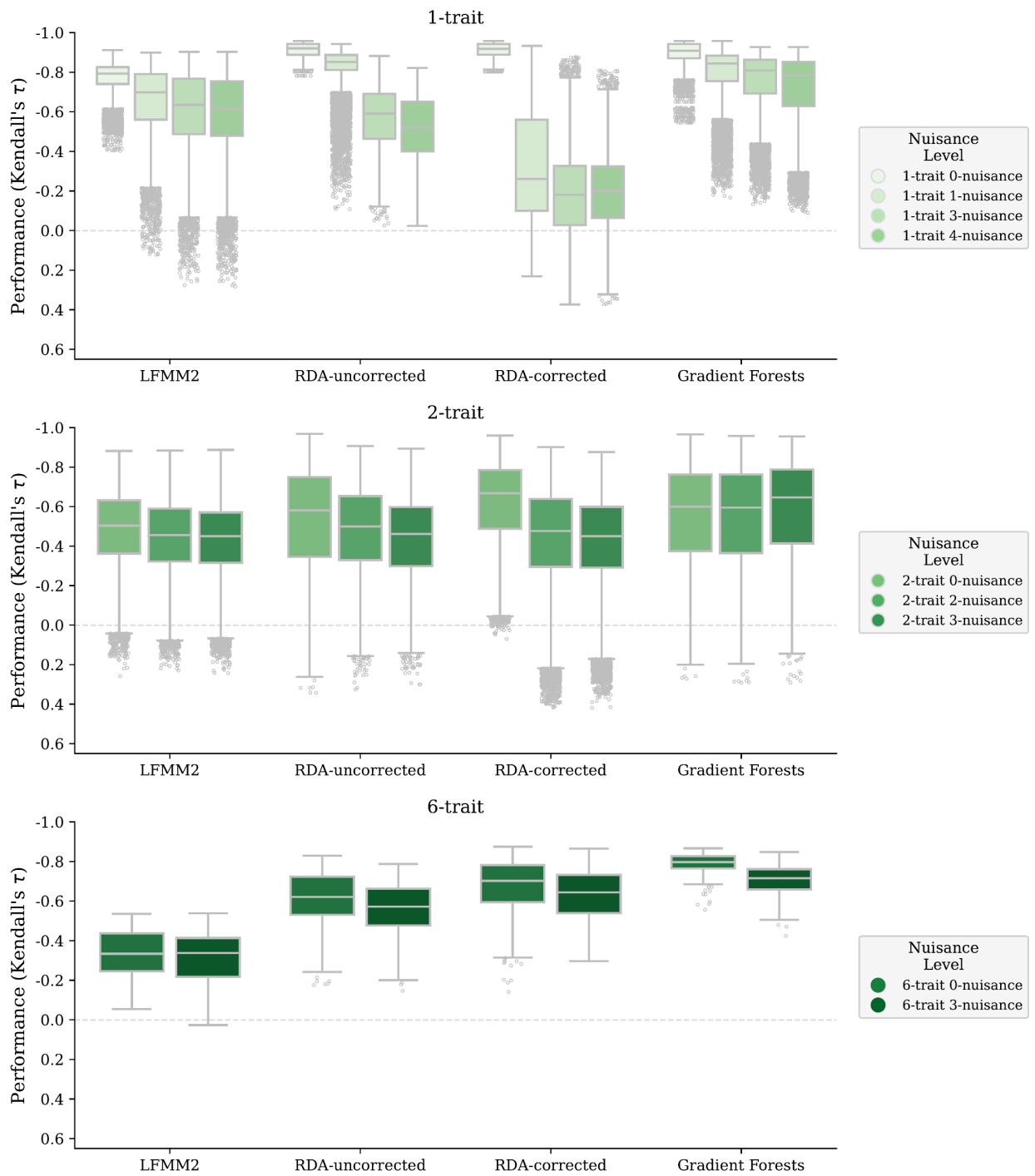


**Figure 3** Comparison of marker choice across genomic offset methods for 1- and 2-trait simulations. A-C are scatterplots of pairwise comparisons of performance between marker sets (histograms in each margin) from both 1- and 2-trait models where density of points is indicated by color in legend (note color scale is different for each figure to accentuate patterns in data). D-E are boxplots from the same data in A-C separated by individual traits. Data included in these figures is from all 1- and 2-trait models from the *Adaptive Environment* workflow. Code to create these figures can be found in SC 02.02.03.



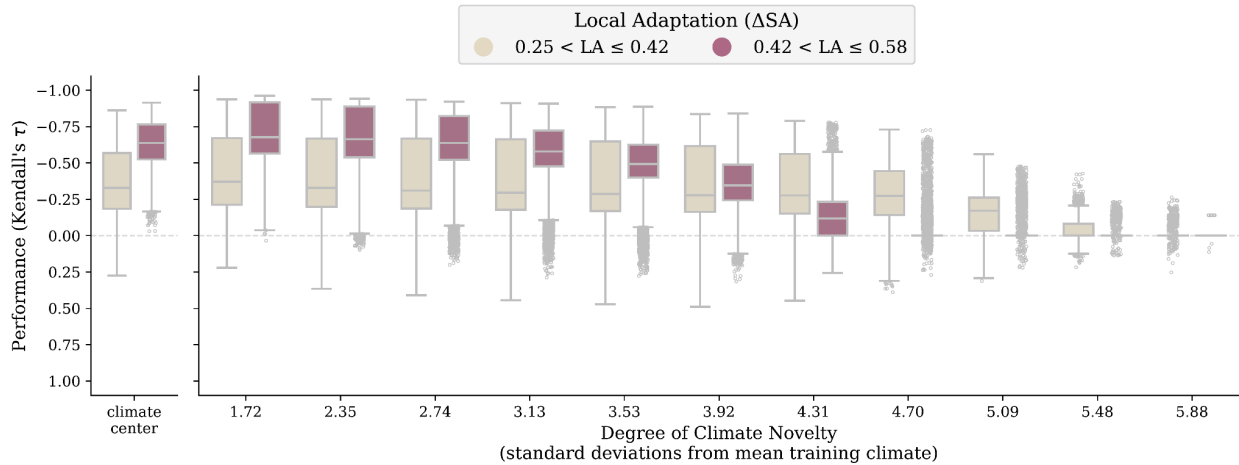
**Figure 4** Comparison of marker choice across genomic offset methods for the 6-trait simulation. A-C are scatterplots of pairwise comparisons of performance between marker sets (RONA is not shown, except in SN 02.05.10). D are boxplots from the same data in A-C (RONA<sub>6-traits</sub> is the combined performance across all six environmental models). Data included in this figure is from the 6-trait models processed through the *Adaptive Environment* workflow. Note there is only one 6-trait replicate, and variation within figures represents the performance across 100 common gardens for each method. Code to create these figures can be found in SN 02.05.10.

(Fig. 5)





**Figure 5** Effect of non-adaptive nuisance environmental variables on offset performance. Shown are evaluations of offsets from 1- and 2-trait models trained using only adaptive environments (*O- nuisance*) or with adaptive environments and the addition of  $N > 0$  non-adaptive environmental variables (*N- nuisance*). RONA is not shown because it is univariate with respect to environmental variables. Nuisance variables are listed in Table 2. Code to create figures can be found in SC 02.02.06 and SC 02.02.08.



**Figure 6** Performance decays with climate novelty relative to training data. Shown is model performance (y-axis) across methods at climate center (A) and across common gardens each representing increasing degrees of climate novelty relative to training data (x-axis of B) where all 100 populations have been transplanted. The standard deviation values (x-axis, B) are applicable to all environments for all landscapes except for *Env2* in the *Stepping Stone - Mountain* landscape; the corresponding standard deviations are 1.55, 2.12, 2.47, 2.82, 3.18, 3.53, 3.88, 4.24, 4.60, 4.95, 5.3. When fitness for all transplanted individuals was zero, a model's performance was undefined and manually set to 0; no method predicted a single offset value for all populations in these situations. Setting undefined performance to 0 did not substantially impact patterns between performance and climate novelty, and is explored in Supplemental Text S3. Data included in this figure are from models trained using 1- and 2-trait simulations from the *Climate Novelty* workflow, which excludes both  $RONA_{temp}$  and  $RONA_{Env2}$ . Code used to create this figure can be found in SC 02.04.05.

1 Supplemental Information

2 **The limits of predicting maladaptation to future**  
3 **environments with genomic data**

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8 **Running Title:** *The limits of genomic offsets*

9 **Keywords:** genomic offset, environmental change, climate change, assisted gene  
10 flow, genomic forecasting, restoration

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15 **Table of Contents**

16 **Supplemental Notes .....6**

17     **S1 - Implementation of Offset Methods ..... 6**

18         1.1 | Gradient Forests .....6

19         1.2 | The Risk Of Non-Adaptedness ..... 7

20         1.3 | Landscape and Ecological Association (LEA) Studies R package .....8

21             Fig S1 Distribution of  $K$  used for the lfm2 genetic.offset function

22             for 1- and 2-trait simulations.. .....9

23             Fig S2 Percent variance explained from principal component (PC)

24             axes from principal component analysis of SNP data from the 6-trait

25             simulation..... 10

26         1.4 | Redundancy Analysis ..... 11

27             Fig S3 Performance of RDA-outlier markers are on par with other

28             marker sets for (A) 1-trait, (B) 2-trait, and (C) 6-trait evaluations of

29             offset estimated with (RDA-corrected) or without (RDA-uncorrected)

30             population structure correction.. ..... 13

31     **S2 - Coding workflows ..... 14**

32         1.1 | The Adaptive Environment coding workflow ..... 14

33             1.1.1 | 1- and 2-trait simulations..... 15

34             1.1.2 | 6-trait simulation..... 16

35         1.2 | The Climate Novelty coding workflow ..... 16

36         1.3 | The Nuisance Environment coding workflow ..... 17

37         1.4 | Misc ..... 17

38     **S3 - Defining Climate Novelty scenarios ..... 18**

39             Fig S5 Differentiation of *Climate Novelty* environments (blue stars,

40             including climate center) from within-landscape environments

41             (black circles) using Principal Component Analysis (PCA) of

42             environmental data..... 19

43     **S4 - Missing data in Climate Novelty evaluations ..... 20**

44             Fig S33 The effect of simulation parameters on missing data for

45             *Climate Novelty* scenarios..... 25

46 **Supplemental Tables ..... 26**

47     Table S1 Results from Type II ANOVAs from regressing simulation

48     factors on offset performance (see Equation 1 of the main text)..... 26

49 Table S2 Results from Type II ANOVAs from regressing the  
50 proportion of clinal QTNs (*cor\_TPR\_tmp* and *cor\_TPR\_sal*) and clinal  
51 neutral alleles (*cor\_FPR\_temp\_neutSNPs*, *cor\_FPR\_sal\_neutSNPs*) on  
52 offset performance (see Equation 2 of the main text).. ..... 27

53 Table S3 Results from Type II ANOVAs regressing two factors -  
54 degree of local adaptation (*final\_LA*) and levels of isolation-by-  
55 environment in all marker sets) on offset performance..... ..... 28

56 Table S4 Gradient Forests (GF) sometimes incorrectly identifies the  
57 environments driving adaptation.. ..... 29

58 **Supplemental Figures** ..... 30

59 Fig S4 Correlation (Spearman’s rho) among environmental  
60 variables faceted by landscape.. ..... 31

61 Fig S6 Percent sum of squares of the various factors from the  
62 ANOVA model in Table S1.. ..... 34

63 Fig S7 Effect of the degree of local adaptation (x-axes) on method  
64 performance (y-axes) colored by the relative strength of selection on  
65 the two traits..... ..... 35

66 Fig S8 Effect of polygenicity on performance of offset methods  
67 trained using all markers on simulations with two adaptive  
68 traits..... ..... 36

69 Fig S9 Effect of demography on performance of offset methods  
70 trained using all markers on simulations with two adaptive  
71 traits..... ..... 37

72 Fig S10 Stacked bar plot of the percent sum of squares from Type II  
73 ANOVAs from regressing the proportion of clinal QTNs and clinal  
74 neutral alleles on offset performance (see Equation 2 of the main  
75 text)..... ..... 38

76 Fig S11 Impact on method performance (y-axes) from the  
77 proportion of QTNs with clinal relationships with *temp* (first  
78 column) or *Env2* (second column).. ..... 43

79 Fig S12 Stacked bar plot showing correlation between  
80 environmental variables and axes of population genetic structure  
81 (Principal Component Analysis axes [PC axes])..... ..... 44

82 Fig S13 Relationship between the proportion of clinal neutral loci  
83 for *temp* (y-axes, first row) or *Env2* (y-axes, second row) with the  
84 strength of the relationship between environmental variables and  
85 axes of population genetic structure. Purple = *Stepping Stone* -

86 *Clines*; teal = *Stepping Stone - Clines*; yellow = *Estuary - Clines*. Data  
87 included in this figure is from all 2-trait simulations.. ..... 45

88 Fig S14 Relationship between median performance and absolute  
89 correlation (Pearson's  $r$ ) between environmental variables and axes  
90 of population genetic structure (principal component analysis  
91 axes)..... 51

92 Fig S15 Adaptive markers contain greater levels of isolation-by-  
93 environment (*IBE*) than other marker sets..... 52

94 Fig S16 The relationship between the degree of local adaptation  
95 ( $LA_{\Delta SA}$ ), levels of *IBE* within marker sets, and median performance of  
96 models trained with one of the three marker sets: (A) *adaptive*, (B)  
97 *all*, and (c) *neutral* marker sets..... 54

98 Fig S17 Levels of isolation-by-environment in marker sets vary  
99 across landscapes (A) and the degree of local adaptation reached by  
100 metapopulations on these landscapes (B).. ..... 55

101 Fig S18 The levels of isolation-by-distance in marker sets (panels)  
102 are weakly correlated with the degree of local adaptation ( $LA_{\Delta SA}$ )  
103 within simulation levels..... 56

104 Fig S19 Differences in levels of *IBE* between marker sets used to  
105 train models is generally unrelated to differences in model  
106 performances..... 60

107 Fig S20 A map of Garden ID (unbolded entries) across each  
108 landscape for 1-, 2- and 6-trait simulations (latitudinal and  
109 longitudinal grids are bolded)..... 61

110 Fig S21 Genomic offset methods have variable performance across  
111 the *Stepping-Stone - Clines* landscape.. ..... 64

112 Fig S22 Genomic offset methods have variable performance across  
113 the *Stepping-Stone - Mountain* landscape.. ..... 67

114 Fig S23 Genomic offset methods have variable performance across  
115 the *Estuary - Clines* landscape..... 70

116 Fig S24 Variability of genomic offset performance (y-axes) for a  
117 given model (+) often decreases with increasing median performance  
118 (x-axes).. ..... 72

119 Fig S25 Variability across evaluations of genomic offsets often  
120 decreases with increasing average performance across marker sets..  
121 ..... 77

122 Fig S26 Variability across evaluations of genomic offsets often

123	decreases with increasing average performance across marker	
124	set.....	74
125	Fig S27 Variability across evaluations of genomic offsets is often	
126	unrelated to the variability in the degree of local adaptation across	
127	populations.....	75
128	Fig S28 Variability across evaluations of genomic offsets is often	
129	unrelated to the variability in the degree of local adaptation across	
130	populations.....	76
131	Fig S29 Variability across evaluations of genomic offsets is often	
132	unrelated to the variability in the degree of local adaptation across	
133	populations.....	77
134	Fig S30 Effect of non-adaptive nuisance environmental variables on	
135	offset performance faceted by landscape.....	78
136	Fig S31 Effect of non-adaptive nuisance environmental variables on	
137	offset performance faceted by marker set.....	79
138	Fig S32 Pairwise comparison of performance differences between	
139	marker sets for <i>Nuisance Environment</i> scenarios.....	84
140	Fig S34 Pairwise comparison of performance differences between	
141	marker sets for <i>Climate Novelty</i> scenarios.....	86
142	<b>Supplemental References.....</b>	<b>87</b>
143		

## 144 Supplemental Notes

### 145 S1 - Implementation of Offset Methods

146

147 See Supplemental Note S2 for specific citations of code.

148

#### 149 1.1 | Gradient Forests

150 For a given set of input loci (*all*, *adaptive*, or *neutral*; see Q3 in Methods), and

151 for all workflows, Gradient Forests ( $GF_{\text{Offset}}$ ) is trained using `ntree=500`,

152 `corr.threshold=0.5`, and `maxLevel=(0.368 *  $\frac{N}{2}$ )`, where  $N$  is the number of

153 populations. Using default linear extrapolation, the trained model is projected onto

154 the landscape using the ``predict`` function and the same environmental values

155 used in training. This creates the “current” projection used to calculate offset

156 below.

157 The trained models are then fit to the climate of each of 100 common gardens

158 on the landscape for the *Adaptive Environment* and *Nuisance Environment*

159 scenarios, or to each of the 11 *Climate Novelty* scenarios. Specifically, for each

160 garden, the ``predict`` function is used to take the trained model and the garden’s

161 climate to create a projection similar to that using current climate data (previous

162 paragraph). Then the Euclidean distance is taken between the current and future

163 projections to calculate offset.



164 **1.2 | The Risk Of Non-Adaptedness**

165 For a given set of input loci (*all*, *adaptive*, or *neutral*; see Q3 in Methods), we  
 166 first discarded any locus that did not have significant ( $p \leq 0.05$ ) linear models  
 167 relating population-level allele frequencies with environmental variables.  $p$ -  
 168 values were not corrected for multiple testing. For each common garden, and once  
 169 for each environmental variable, RONA offset for each population was calculated  
 170 by averaging the absolute allele frequency difference between the population's  
 171 current frequency and that predicted by using each locus' linear model fit using  
 172 climate of the common garden,

$$\text{RONA} = \frac{1}{n} \sum_{i=1}^n |(S_{\text{present}_i} * \text{EF}_{\text{future}} + I_{\text{present}_i}) - \text{AAF}_{\text{present}_i}|$$

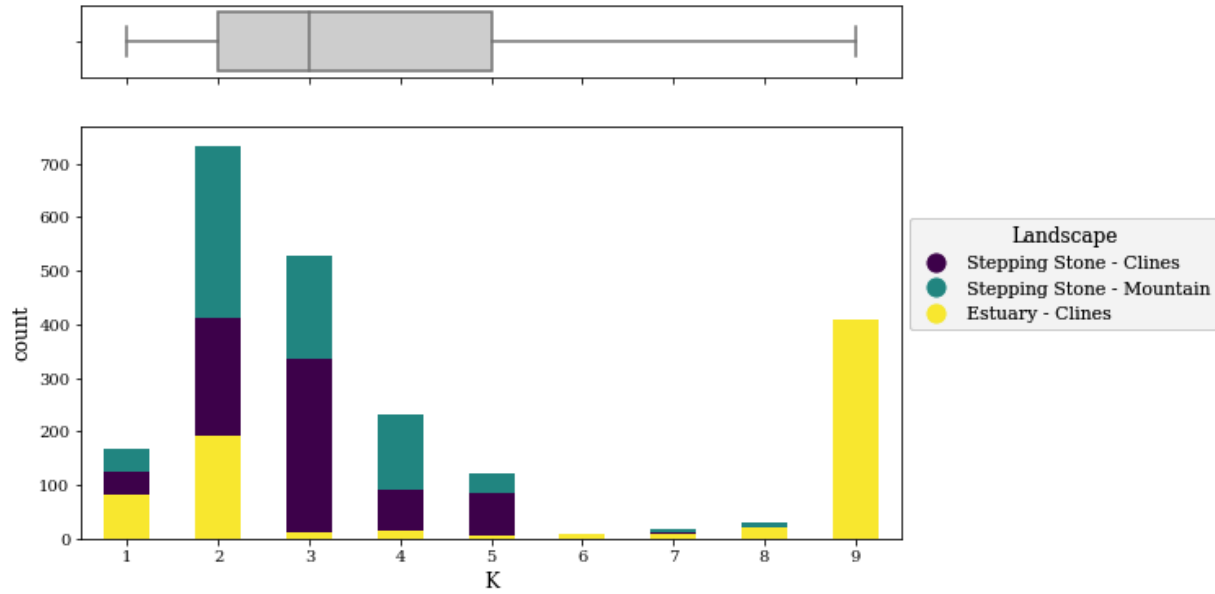
173  
 174 where  $n$  is the total number of loci with significant linear models;  $S_{\text{present}}$  and  $I_{\text{present}}$   
 175 are respectively the slope and intercept from the linear model for locus $_i$  relating  
 176 current climate and allele frequencies from all populations;  $\text{AAF}_{\text{present}}$  is the current  
 177 allele frequency for the population under consideration; and  $\text{EF}_{\text{future}}$  is the  
 178 environmental value for the common garden. RONA can only be calculated for a  
 179 single population and environmental variable at a time.

180 RONA was excluded from *Nuisance Environment* and *Climate Outlier*  
181 workflows because of its poor (Fig. 2A) and variable (Fig. 2C, Fig. 4) performance  
182 from evaluations from the *Adaptive Environment* workflow.

183 Of note, in some instances, particularly *Adaptive Environment* datasets  
184 simulated with oligogenic architectures, there were no loci with significant linear  
185 relationships with environmental variables and these instances were given NA  
186 performance values (i.e., excluded from analyses).

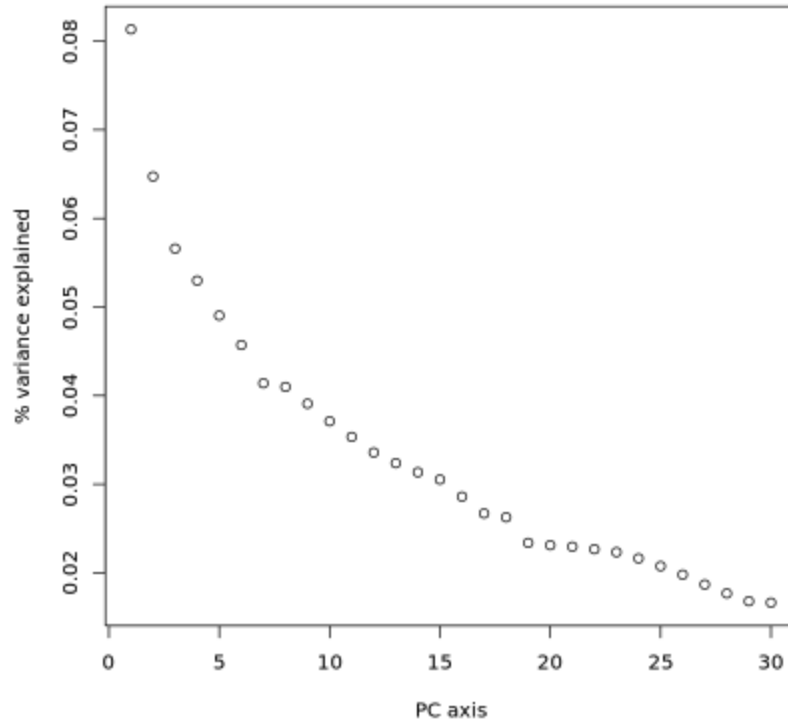
### 187 **1.3 | Landscape and Ecological Association (LEA) Studies R package**

188 We used the `genetic.offset` function in the LEA package to estimate  
189  $LFMM2_{\text{offset}}$  for each workflow (Fig. 1). The `genetic.offset` function was used  
190 with default settings, except for  $K$ , the number of subdivisions within the data. To  
191 determine  $K$  needed for the `genetic.offset` function for 1- and 2-trait  
192 simulations, we first used filtered SNP data (see Section 2.1) to estimate 21 principal  
193 components (PCs) using principal component analysis (PCA). Then we equated  $K$   
194 to the number of PC axes that explain greater than 1.3x the variation of the next  
195 subsequent axis (see line 677-697 of [c-AnalyzeSimOutput.R](#) from Lotterhos, 2023).  
196 This resulted in varied  $K$  across simulation levels and replicates (Fig S1). For the 6-  
197 trait simulation, it was never the case that a PC axis explained  $>1.3x$  the variation  
198 explained by the previous axis, so we used the elbow rule to estimate  $K=7$  (Fig S2).



199

200 **Fig S1** Distribution of  $K$  used for the `lfmm2 genetic.offset` function for 1- and  
 201 2-trait simulations.  $K$  was estimated by determining the number of principal  
 202 component axes that explain at least 1.3x times the amount of variation of the  
 203 subsequent axis. Code used to create this figure can be found in SC 02.09.01.



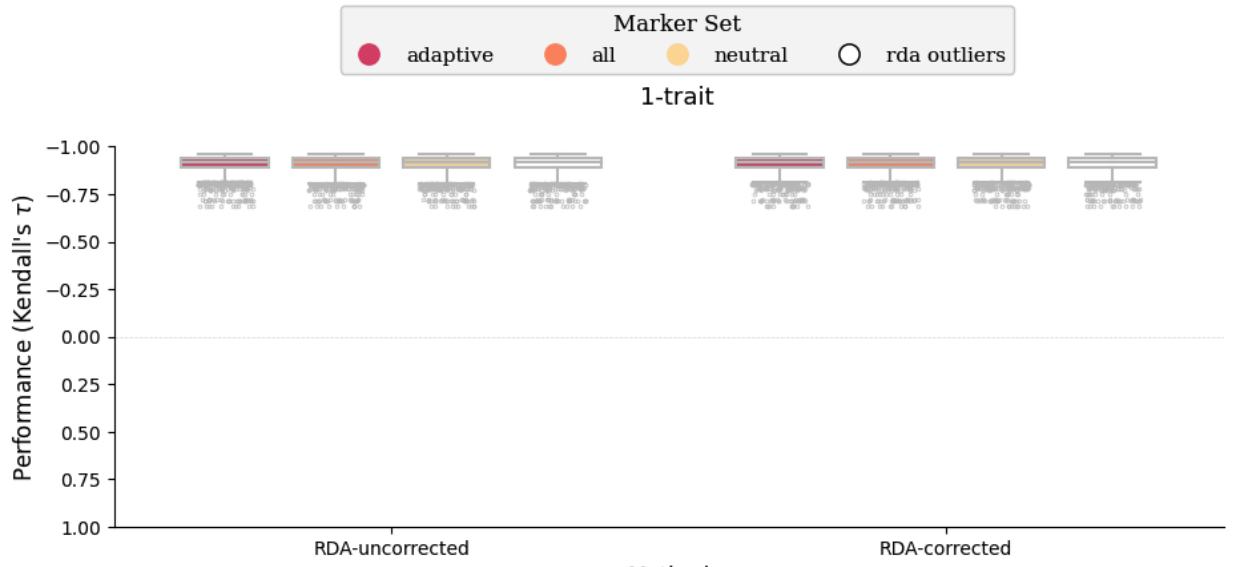
204

205 **Fig S2** Percent variance explained from principal component (PC) axes from  
206 principal component analysis of SNP data from the 6-trait simulation. The “elbow  
207 rule” was used to estimate  $K=7$  for this simulation. Code used to create this figure  
208 can be found in SC 02.05.11.

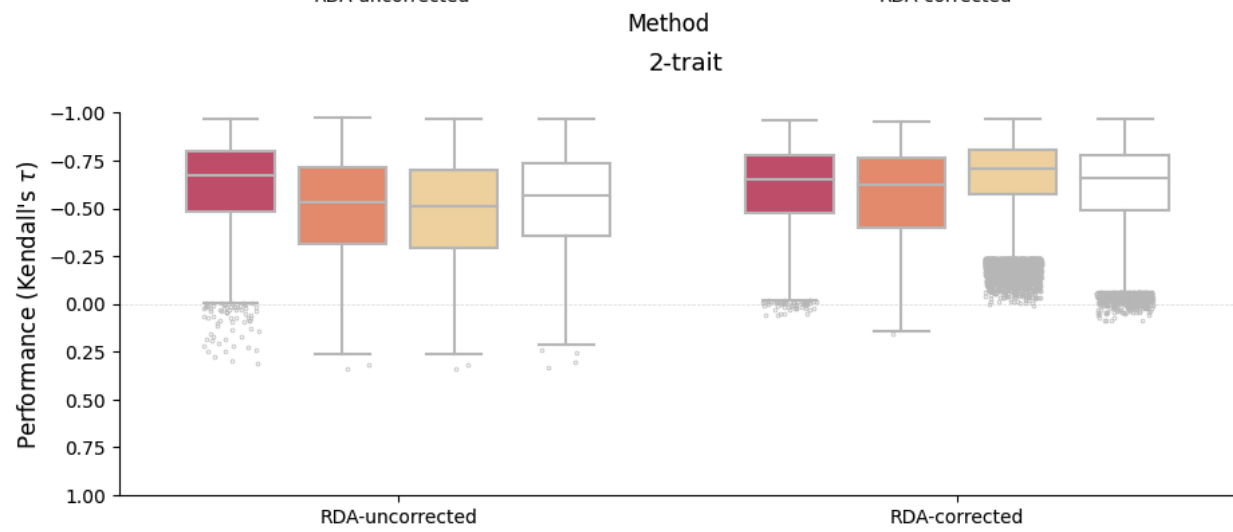
## 209 **1.4 | Redundancy Analysis**

210  $RDA_{\text{offset}}$  was implemented as in Capblancq & Forester (2021). Note that the  
211 environmental variables used here across workflows had minimal correlation, as  
212 required by RDA (Fig. S4). In addition to the three marker sets used as input (*all*,  
213 *adaptive*, or *neutral*; see Q3 in Methods), we also used *RDA-outliers* as input to RDA  
214 offset estimation. *RDA-outlier* loci were those from separate RDA models trained  
215 using *all* loci and adaptive environments, and were included in this study because  
216 of their use in the original implementation of  $RDA_{\text{offset}}$  by Capblancq & Forester  
217 (2021). *RDA-outliers* were identified as in Capblancq et al. (2018) for loci with q-  
218 values  $< 0.05$ . For each 1-, 2-, and 6-trait simulation replicate,  $RDA_{\text{offset}}$  was estimated  
219 with (RDA-corrected) and without (RDA-uncorrected) correction for population  
220 genetic structure. When correcting for structure, the loadings for the first two PCs  
221 from PCA estimated with *all* loci were used. Because *RDA-outliers* performed on  
222 par with or worse than other marker sets in 1-trait (Fig S2A), 2-trait (Fig S3B), and  
223 6-trait (Fig S3C) evaluations from the *Adaptive Environment* workflow (Fig. 1) we  
224 focus on *all*, *neutral*, and *adaptive* marker sets for the main text.

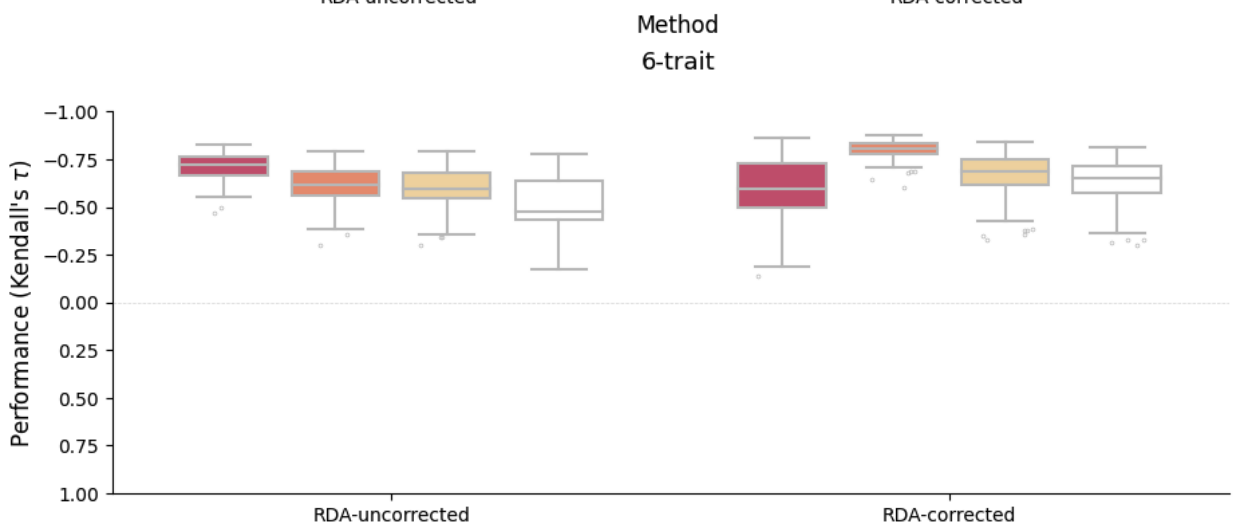
Supplement - *Lind, Lotterhos, and the limits of genomic offsets*



225



226



227

228 **Fig S3** Performance of RDA-outlier markers are on par with other marker sets for  
229 (A) 1-trait, (B) 2-trait, and (C) 6-trait evaluations of offset estimated with (RDA-  
230 corrected) or without (RDA-uncorrected) population structure correction. Data in  
231 this figure is from the *Adaptive Environment* workflow. Code to create this figure  
232 can be found in SC 02.06.02.

## 233 S2 - Coding workflows

234 Below we reference the scripts (\*.R, \*.py) and notebooks (\*.ipynb) used to  
235 analyze data in this manuscript using the naming convention described in the Data  
236 Availability section (e.g., SC 05.02). Scripts are often written using only functions,  
237 instead of a linear development of code. This allows the functions to be  
238 imported/sourced in other scripts or notebooks to avoid code redundancy. At the  
239 top of all script files are detailed instructions for use. The “main” function in many  
240 script files gives a general outline for the code and calls all other functions.

241 All python scripts and notebooks are run in the “mvp\_env” (python v3.8)  
242 Anaconda environment. All GF scripts are run in R within the “r35” (R v3.5)  
243 Anaconda environment. All other R code is run within the “MVP\_env\_R4.0.3” (R  
244 v4.0.3) Anaconda environment. All Anaconda environments can be recreated using  
245 their .yaml files found in the code archive. These files contain all package and  
246 library versions at the time of saving. Package and library versions that were used  
247 are found at the top of each notebook - look for “Click to view session information”  
248 (python notebooks) or printouts from `sessionInfo()` (R notebooks).

249 1- and 2-trait simulations are often processed separately from the 6-trait  
250 simulation. Descriptions of coding workflows reflect this.

251 All scripts referenced by name are in the SC 01 directory.

252 Notebooks used to create figures and tables are not described here (but see  
253 coding archive README). Instead, these notebooks are referenced within the  
254 caption of all figures and tables, or in the main text when appropriate. These  
255 notebooks (mainly within SC 02.02 directory) rely on data processed through the  
256 coding workflows described below. Similarly, code previously described in  
257 Supplemental Note S1 is not redescribed here.

258 Simulation data used below within scripts and notebooks has been  
259 processed from SLiM output separately by Lotterhos (2023) into more user-friendly  
260 forms - see here for more information:  
261 [https://github.com/ModelValidationProgram/MVP-](https://github.com/ModelValidationProgram/MVP-NonClinalAF/tree/main/sim_output_20220428_metadata)  
262 [NonClinalAF/tree/main/sim\\_output\\_20220428\\_metadata](https://github.com/ModelValidationProgram/MVP-NonClinalAF/tree/main/sim_output_20220428_metadata)

### 263 1.1 | The *Adaptive Environment* coding workflow

264 The *Adaptive Environment* workflow represents the general pipeline for  
265 processing simulations and running genomic offset methods, most other



266 processing code is built on top of this main pipeline (i.e., scripts and notebooks  
267 source/import functions from these scripts to avoid code redundancy).

### 268 *1.1.1 / 1- and 2-trait simulations*

269 The *Adaptive Environment* pipeline is kicked off using SC 01.00, which  
270 allows the user to decide which method to run. All analyses were generally run in  
271 batches of 225 simulation levels (one replicate per level). SC 01.00 can call SC 01.01  
272 (for GF), SC 01.05 (for RONA), SC 01.10 (for LFMM), SC 01.07 (for pairwise  $F_{ST}$ ), or  
273 scripts related to RDA (more details below).

274  $GF_{offset}$  : SC 01.01 processes the data into formats suitable for GF input. This  
275 includes converting genotype data into derived allele frequencies, asserting MAF  
276 cutoffs, and reformatting environmental data. This script creates `.sh` files for the  
277 slurm HPC and trains GF models using ``MVP_gf_training_script.R``. The slurm  
278 `.sh` files call SC 01.02, which takes the trained GF model and predicts offset to each  
279 of the 100 environments (population sources) on the landscape using  
280 ``MVP_gf_fitting_script.R``. Performance of GF offset predictions are then  
281 validated using SC 01.03. Performance results are saved in a nested dictionary.  
282 Environmental importance is extracted from each GF model using SC 01.04 within  
283 SC 02.10.02.

284 RONA : Using files created from SC 01.01, SC 01.05 creates files suitable for  
285 RONA analyses and calculates RONA itself. Performance of RONA is validated with  
286 SC 01.06. As with GF, performance results are saved in a nested dictionary.

287  $LFMM_{offset}$  : SC 01.10 creates files suitable for LFMM in R and submits jobs to  
288 the slurm HPC to train LFMM with ``MVP_process_lfmm.R``. SC 01.10 also submits  
289 SC 01.11 to validate LFMM offsets.  
290 ``MVP_watch_for_failure_of_train_lfmm2_offset.py`` watches for failed  
291 jobs and reruns them. Performance of LFMM is validated in SC 01.11. As with GF  
292 and RONA, performance results are saved in a nested dictionary.

293  $RDA_{offset}$  : ``MVP_pooled_pca_and_rda.R`` creates principal component  
294 analysis data and RDA objects using allele frequencies of *all* loci; it also creates  
295 additional files needed downstream. Next, SC 01.12 is run to estimate RDA offset.  
296 Performance of  $RDA_{offset}$  is validated with SC 01.13. As with GF, RONA, and LFMM,  
297 performance results are saved in a nested dictionary.

298 Nested dictionaries containing validation results from each method are  
299 reformatted and combined into a single object in notebooks within the SC 02.01.00

300 directory. These combined objects are used throughout remaining analyses in  
301 jupyter notebooks found in subdirectories of SC 02.

### 302 *1.1.2 / 6-trait simulation*

303 The 6-trait simulation was processed through the *Adaptive Environment*  
304 workflow using code found in the SC 02.05 directory. 6-trait simulations needed  
305 extra formatting in order to be comparable to the 1- and 2-trait evaluations. First,  
306 SC 02.05.00 assigns individuals to populations using a gridded system. Population-  
307 level environmental values are the average climate from assigned individuals on  
308 the landscape (each environmental variable is averaged independently). Genetic  
309 and environmental data was formatted as with 1- and 2-trait simulations. Fitness  
310 for each population in each environment was calculated using  
311 ``MVP_climate_outlier_fitness_calculator.R``. The script  
312 ``MVP_climate_outlier_fitness_calculator.R`` was validated against  
313 previous fitness estimates from 1- and 2-trait simulations in SC 02.05.01.

314 GF was trained using the same script as 1- and 2-trait simulations  
315 (``MVP_gf_training_script.R``). GF offset was predicted manually in SC 02.05.02,  
316 and validated manually in SC 02.05.03. In SC 02.05.04 - 02.05.05 LFMM was trained  
317 and validated manually. Similarly, RDA was trained and validated in SC 02.05.06 -  
318 SC 02.05.07, and RONA trained and validated in SC 02.05.08 - SC 02.05.09.

## 319 **1.2 | The *Climate Novelty* coding workflow**

320 Fitness was calculated for 1- and 2-trait populations within the *Climate*  
321 *Novelty* scenarios (x-axis, Fig. 6, Supplemental Note S3) using  
322 ``MVP_climate_outlier_fitness_calculator.R`` in 02.04.01.

323 Using 1- and 2-trait offset models output from the *Adaptive Environment*  
324 workflow, the following code predicted offset to *Climate Novelty* scenarios (GF: SC  
325 01.14; LFMM: SC 01.16; RDA: SC 01.18; and RONA: SC 01.20) which was subsequently  
326 validated against known fitness (GF: SC 01.15; LFMM: SC 01.17; RDA: SC 01.19; and  
327 RONA: SC 01.21). A few examples of code executions are shown in SC 02.04.03.

328 Fitness of 6-trait populations for *Climate Novelty* scenarios was calculated in  
329 SC 02.04.06. 6-trait GF models used the same scripts as 1- and 2-trait runs (SC 01.14  
330 - SC 01.15); executed from SC 02.04.07. Commands to train LFMM were created in  
331 SC 02.04.07, which called on ``MVP_complex_sims_process_lfmm.R``. RDA was  
332 trained manually in SC 02.04.07. Offset from both LFMM and RDA were validated  
333 manually in SC 02.04.08.

### 334 **1.3 | The *Nuisance Environment* coding workflow**

335 Environmental files for *Nuisance Environment* scenarios were created in SC  
336 02.07.02.02.

337 Files for 1- and 2-trait simulations were created in SC 02.07.02.01 to train GF  
338 using ``MVP_gf_training_script.R``. SC 01.02 and SC 01.03 are used for  
339 predicting and validating GF offset, respectively, executed in SC 02.07.02.02. Code  
340 for LFMM was executed in SC 02.07.02.07 and used ``MVP_process_lfmm.R`` for  
341 training and SC 01.11 for validation. Commands for RDA were created in SC  
342 02.07.02.06 similarly to *Adaptive Environment* workflow (calling  
343 ``MVP_pooled_pca_and_rda.R``) and used ``MVP_nuisance_RDA_offset.R`` for  
344 training and ``MVP_nuisance_rda_validation.py`` for validation.

345 6-trait sims were processed for GF exactly as they were for 6-trait data in the  
346 *Adaptive Environment* workflow (with updated environmental files) and executed  
347 in SC 02.07.02.03, SC 02.07.02.04, and validated manually in SC 02.07.02.10. Code to  
348 train both LFMM and RDA was executed in SC 02.07.02.12, which called on  
349 ``MVP_complex_sims_process_lfmm.R`` for LFMM. LFMM was validated in SC  
350 02.07.02.13; RDA was validated in SC 02.07.02.14.

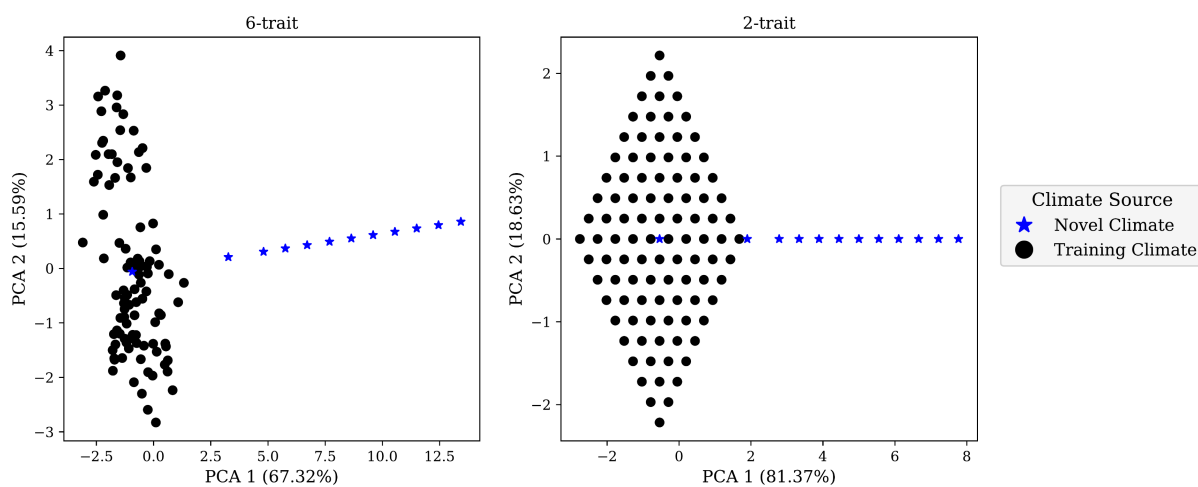
### 351 **1.4 | Misc**

352 ``MVP_summary_functions.py`` contains much of the API used within notebooks  
353 for loading and filtering data as well as creating figures. It is often imported using  
354 the alias ``mvp`` within python scripts and notebooks.

355 S3 - Defining *Climate Novelty* scenarios

356 To understand if genomic offset models maintained predictive performance  
 357 in environments differentiated from training environments, we created 11  
 358 climates, each progressively more distant from the mean training environment.  
 359 Specifically, for each environmental variable, we used a standardized set of z-  
 360 scores ( $z_E \in \{1.72, 2.35, 2.74, 3.13, 3.53, 3.92, 4.31, 4.70, 5.09, 5.48, 5.88\}$ ) to calculate  
 361 corresponding environmental values. In other words, we used the distribution of  
 362 the within-landscape values from which to identify the appropriate value for a  
 363 given z-score for each environmental variable independently. The *temp*  
 364 environment and all six of the 6-trait environments were given positive values for  
 365 *Climate Novelty* scenarios, and *Env2* was given negative values.

366 Novelty climates for 6-trait and 2-trait evaluations are shown Fig S5A and B,  
 367 respectively. In this and other figures related to performance in *Climate Novelty*  
 368 scenarios, we also include  $z_E=0.00$  for comparison of novelty climates to the mean  
 369 training climate (i.e., climate center). We chose z-scores over Mahalanobis  
 370 distances because of 1) the reduced correlation structure among environmental  
 371 variables (where z-scores and Mahalanobis distances should be roughly  
 372 equivalent; Fig. S4), and 2) the large number of combinations of values from  
 373 environmental variables that could be used for a given Mahalanobis distance. The  
 374 standard deviation values that we used are applicable to all environments and for  
 375 all landscapes except for *Env2* in the *Stepping Stone - Mountain* landscape; the  
 376 corresponding standard deviations for this case are  $z_E \in \{1.55, 2.12, 2.47, 2.82, 3.18,$   
 377  $3.53, 3.88, 4.24, 4.60, 4.95, 5.3\}$ .



378

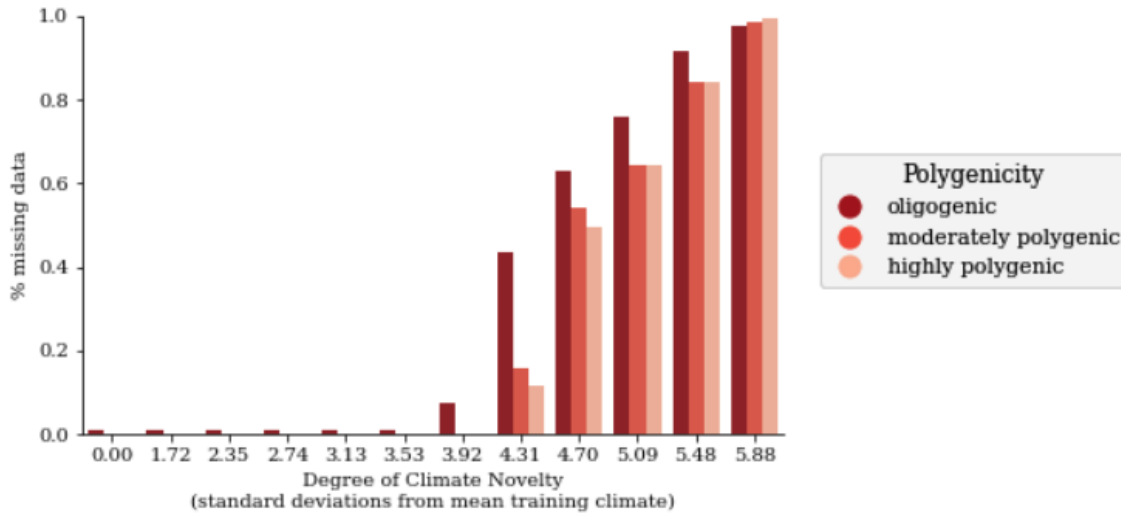
379 **Fig S5** Differentiation of *Climate Novelty* environments (blue stars, including  
380 climate center) from within-landscape environments (black circles) using  
381 Principal Component Analysis (PCA) of environmental data. Environmental data  
382 is centered and standardized relative to the within-landscape environmental  
383 values. Scatter plots show the first two principal components (PCs) of  
384 environmental data used to evaluate 6-trait (A) and 2-trait (B) *Climate Novelty*  
385 scenarios. There is no figure for 1-trait evaluations because there would only be  
386 one PC axis. Code to create these figures can be found in SC 02.04.10.

387 S4 - Missing data in *Climate Novelty* evaluations

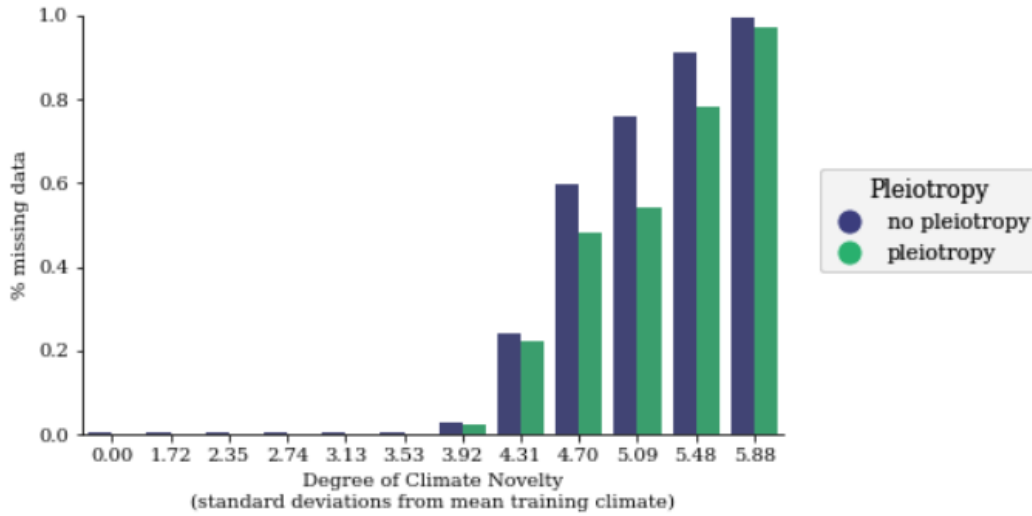
388           When calculating fitness of populations in *Climate Novelty* scenarios, it  
389 could be the case that all populations have zero fitness because of the extremity of  
390 the novel climate. In these cases the calculation of performance is technically  
391 undefined due to the lack of variability in one of the vectors (i.e., the code returns  
392 “NAN”), but for Figure 7 we replaced these undefined values with 0 (because there  
393 was no predictive performance of the offset method). We refer to these cases as  
394 missing data below. It is therefore important to explore the effect of these missing  
395 data points on patterns observed between performance and climate novelty (i.e.,  
396 in the context of Fig. 7 of the main text) to ensure patterns before and after setting  
397 missing data to 0 do not affect inferences.

398           To understand impacts of missing data, we created figures that grouped  
399 simulation and experimental levels across novelty scenarios (Fig. S33). We also  
400 printed out specific scenarios in the code (SC 02.04.05). Importantly, missing data  
401 is not substantial until *Climate Novelty (CN) Scenario 4.31*, which is preceded by  
402 the drop in performance from datasets with elevated  $LA_{\Delta SA}$ . After *CN Scenario 4.31*  
403 missing data begins to increase because of climate novelty, first with datasets  
404 where high levels of  $LA_{\Delta SA}$  take place through oligogenic architectures, then  
405 missing data is more uniform across simulation and experimental parameters for  
406 the remaining *CN Scenarios* (Fig. S33). (Before *CN Scenario 4.31*, missing data is not  
407 due to all populations having zero fitness - instead missing data is primarily due to  
408 1-trait oligogenic scenarios evaluated by RONA where there are no *adaptive* alleles  
409 with significant clines with *temp* in the *Estuary - Clines* landscape (Fig. S33; SC  
410 02.04.05).) Finally, we also explored patterns presented in Fig. 6 before setting  
411 undefined performance scores to zero and found nearly identical trends (not  
412 shown).

413 (Fig S33)

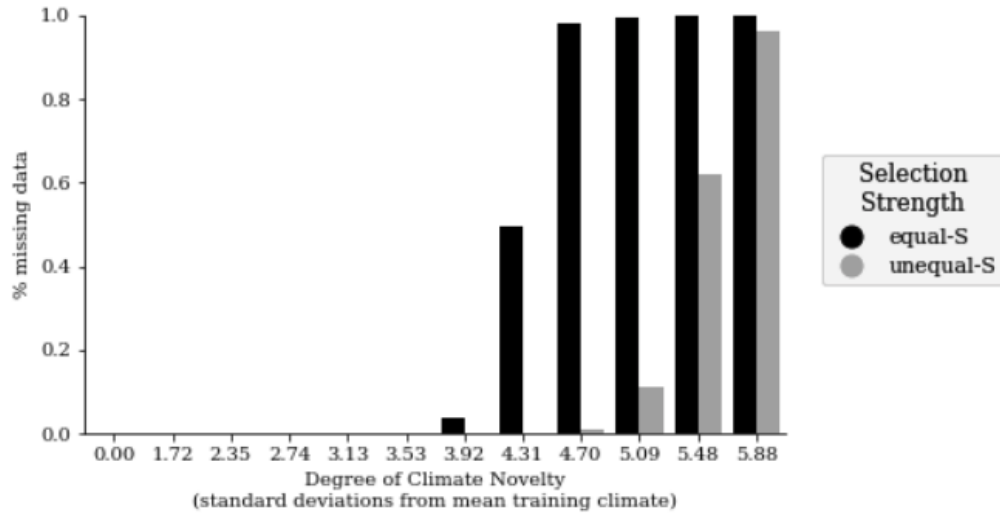


414

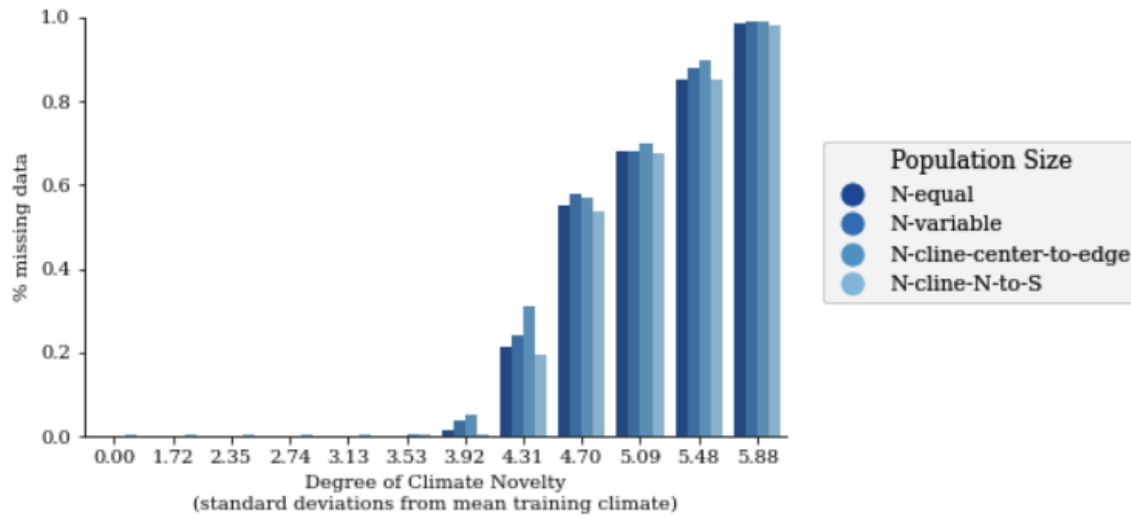


415

416 (Fig S33 continued)



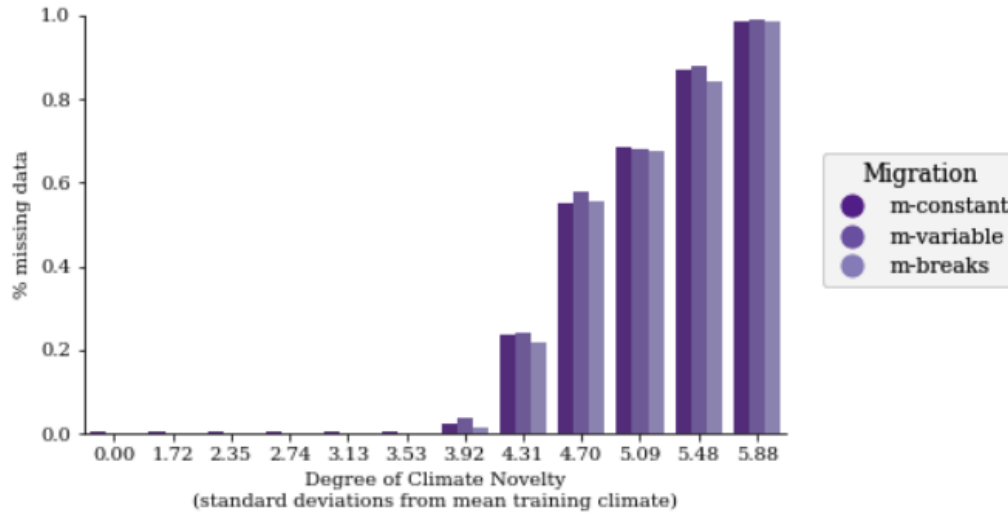
417



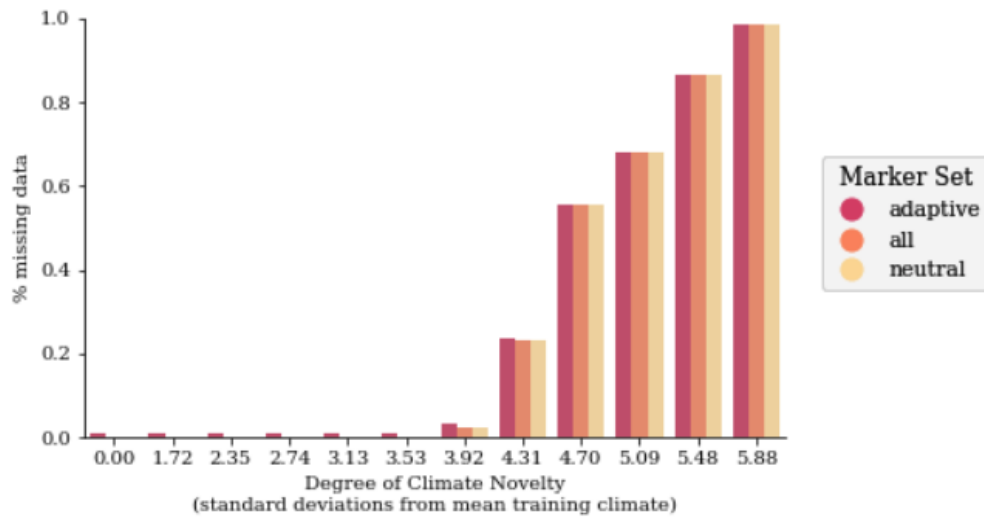
418



419 (Fig S33 continued)

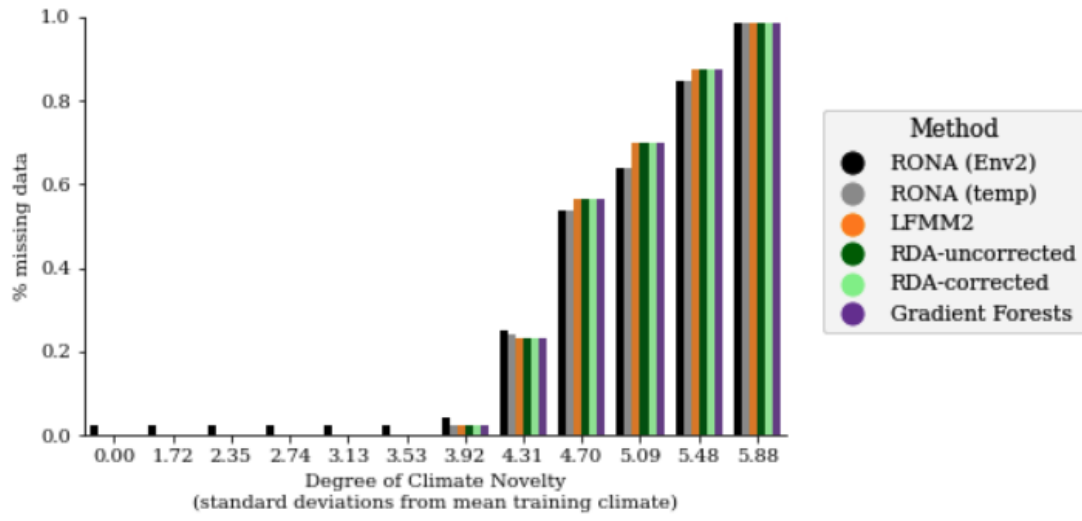


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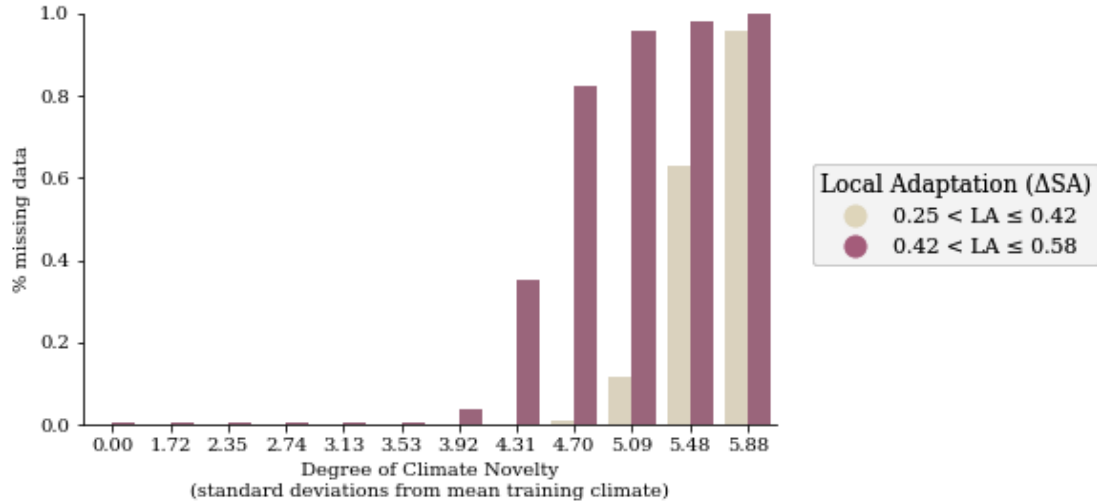


421

422 (Fig S33 continued)

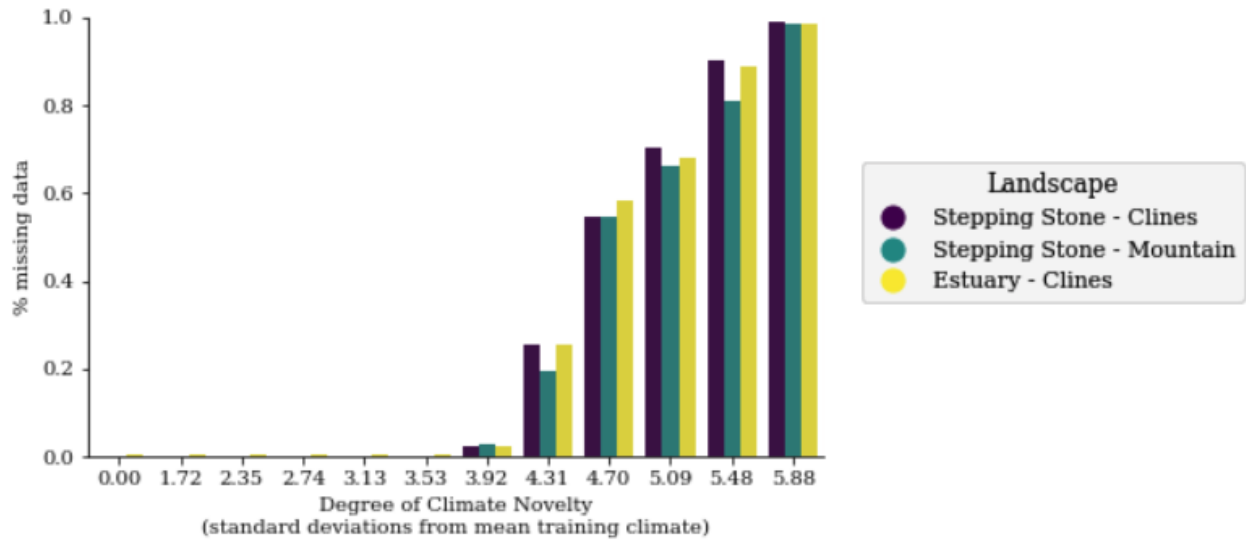


423



424

425 (Fig S33 continued)



426

427 **Fig S33** The effect of simulation parameters on missing data for *Climate Novelty*  
 428 scenarios. Shown are the percent missing data (y-axes) due to experimental and  
 429 simulation parameters (legends). Missing data is when all populations have zero  
 430 fitness in a given novelty scenario, and thus performance cannot be defined  
 431 (though we manually set it to zero for other figures). Data included in these figures  
 432 are from 1- and 2-trait evaluations of *Climate Novelty* scenarios. Code to create  
 433 these figures can be found in SC 02.04.05.

434 Supplemental Tables

RONA-sal\_opt

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	15.864775	2.0	210.709735	3.953193e-92	0.17
landscape	149.664742	2.0	1987.788561	0.000000e+00	1.59
demography	433.301763	4.0	2877.472267	0.000000e+00	4.60
plevel_pleio	0.033619	1.0	0.893039	3.446564e-01	0.00
C(garden)	1941.690095	99.0	520.985219	0.000000e+00	20.61
cor_TPR_temp	0.724989	1.0	19.258049	1.142528e-05	0.01
cor_TPR_sal	3.874653	1.0	102.923263	3.536502e-24	0.04
cor_FPR_temp_neutSNPs	0.897974	1.0	23.853067	1.040651e-06	0.01
cor_FPR_sal_neutSNPs	14.829886	1.0	393.929492	1.433616e-87	0.16
final_LA	87.744615	1.0	2330.779320	0.000000e+00	0.93
Residual	6771.995848	179886.0	NaN	NaN	71.88

RONA-temp\_opt

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	3.264399	2.0	99.857527	4.534018e-44	0.03
landscape	74.659979	2.0	2283.838967	0.000000e+00	0.64
demography	26.679201	4.0	408.056627	0.000000e+00	0.23
plevel_pleio	1.505757	1.0	92.121798	8.249029e-22	0.01
C(garden)	4962.472456	99.0	3066.694571	0.000000e+00	42.52
cor_TPR_temp	1.035753	1.0	63.367084	1.725510e-15	0.01
cor_TPR_sal	0.021308	1.0	1.303608	2.535568e-01	0.00
cor_FPR_temp_neutSNPs	2.155381	1.0	131.865626	1.640714e-30	0.02
cor_FPR_sal_neutSNPs	0.039487	1.0	2.415797	1.201186e-01	0.00
final_LA	3659.697867	1.0	223899.355101	0.000000e+00	31.35
Residual	2940.287212	179886.0	NaN	NaN	25.19

435

GF

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	3.109424	2.0	158.728388	1.336264e-69	0.05
landscape	344.465503	2.0	17584.110829	0.000000e+00	5.24
demography	104.816048	4.0	2675.299841	0.000000e+00	1.59
plevel_pleio	0.373620	1.0	38.144788	6.582481e-10	0.01
C(garden)	392.397617	99.0	404.665183	0.000000e+00	5.97
cor_TPR_temp	3.305288	1.0	337.453511	2.682308e-75	0.05
cor_TPR_sal	0.498155	1.0	50.859112	9.961078e-13	0.01
cor_FPR_temp_neutSNPs	46.646231	1.0	4762.349106	0.000000e+00	0.71
cor_FPR_sal_neutSNPs	37.556091	1.0	3834.290889	0.000000e+00	0.57
final_LA	3881.402690	1.0	396271.989655	0.000000e+00	59.02
Residual	1761.946397	179886.0	NaN	NaN	26.79

lfmm2

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	0.648503	2.0	26.071108	4.776417e-12	0.01
landscape	69.824926	2.0	2807.098902	0.000000e+00	1.34
demography	75.595816	4.0	1519.549983	0.000000e+00	1.45
plevel_pleio	0.088264	1.0	7.096738	7.723125e-03	0.00
C(garden)	391.270017	99.0	317.774174	0.000000e+00	7.50
cor_TPR_temp	3.333636	1.0	268.037430	3.359626e-60	0.06
cor_TPR_sal	1.434571	1.0	115.345148	6.738112e-27	0.03
cor_FPR_temp_neutSNPs	18.408493	1.0	1480.114981	1.679823e-322	0.35
cor_FPR_sal_neutSNPs	0.000221	1.0	0.017772	8.939475e-01	0.00
final_LA	2419.263042	1.0	194518.232194	0.000000e+00	46.37
Residual	2237.278977	179886.0	NaN	NaN	42.88

436

rda-nocorr

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	2.722620	2.0	114.252516	2.583753e-50	0.04
landscape	351.685552	2.0	14758.193717	0.000000e+00	5.17
demography	86.983434	4.0	1825.093984	0.000000e+00	1.28
plevel_pleio	0.043391	1.0	3.641764	5.634878e-02	0.00
C(garden)	384.324179	99.0	325.815086	0.000000e+00	5.65
cor_TPR_temp	4.653284	1.0	390.542466	7.801493e-87	0.07
cor_TPR_sal	1.660051	1.0	139.325332	3.842628e-32	0.02
cor_FPR_temp_neutSNPs	57.083425	1.0	4790.917538	0.000000e+00	0.84
cor_FPR_sal_neutSNPs	39.340402	1.0	3301.774980	0.000000e+00	0.58
final_LA	3728.556367	1.0	312931.576881	0.000000e+00	54.83
Residual	2143.328255	179886.0	NaN	NaN	31.52

rda-structcorr

	sum_sq	df	F	PR(>F)	perc_sum_sq
glevel	19.270968	2.0	303.886012	1.763727e-132	0.29
landscape	54.926356	2.0	866.139726	0.000000e+00	0.81
demography	644.354351	4.0	5080.447174	0.000000e+00	9.54
plevel_pleio	2.784965	1.0	87.832851	7.201110e-21	0.04
C(garden)	184.307173	99.0	58.714343	0.000000e+00	2.73
cor_TPR_temp	13.822608	1.0	435.940440	1.078989e-96	0.20
cor_TPR_sal	0.004197	1.0	0.132380	7.159771e-01	0.00
cor_FPR_temp_neutSNPs	1.803296	1.0	56.872745	4.671127e-14	0.03
cor_FPR_sal_neutSNPs	45.534733	1.0	1436.084373	5.262511e-313	0.67
final_LA	85.096985	1.0	2683.807346	0.000000e+00	1.26
Residual	5703.746286	179886.0	NaN	NaN	84.43

437

438 **Table S1** Results from Type II ANOVAs from regressing simulation factors on  
 439 offset performance (see Equation 1 of the main text). In this table, the common  
 440 garden ID was included as a categorical factor (n=100 per simulation). Code to  
 441 create these tables can be found in SC 02.02.01.

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RONA (Env2)

RONA (temp)

	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
<i>P<sub>c</sub>QTN<sub>temp</sub></i>	11.076279	1.0	203.959434	3.027995e-46	0.11	<i>P<sub>c</sub>QTN<sub>temp</sub></i>	461.435097	1.0	6854.545092	0.000000e+00	3.54
<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	7.130868	1.0	131.308340	2.171984e-30	0.07	<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	321.669143	1.0	4778.344045	0.000000e+00	2.47
<i>P<sub>c</sub>Neut<sub>temp</sub></i>	350.186388	1.0	6448.358576	0.000000e+00	3.40	<i>P<sub>c</sub>Neut<sub>temp</sub></i>	21.729532	1.0	322.788745	4.136405e-72	0.17
<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	165.769820	1.0	3052.497975	0.000000e+00	1.61	<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	119.950781	1.0	1781.849799	0.000000e+00	0.92
<b>Residual</b>	9774.859470	179995.0	NaN	NaN	94.82	<b>Residual</b>	12116.925221	179995.0	NaN	NaN	92.91

442

RDA-uncorrected

RDA-corrected

	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
<i>P<sub>c</sub>QTN<sub>temp</sub></i>	326.478563	1.0	8168.479995	0.0	3.29	<i>P<sub>c</sub>QTN<sub>temp</sub></i>	36.345449	1.0	970.987128	1.344472e-212	0.48
<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	171.117641	1.0	4281.356235	0.0	1.72	<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	26.680765	1.0	712.790185	1.002191e-156	0.35
<i>P<sub>c</sub>Neut<sub>temp</sub></i>	589.314088	1.0	14744.613849	0.0	5.93	<i>P<sub>c</sub>Neut<sub>temp</sub></i>	281.267846	1.0	7514.213388	0.000000e+00	3.70
<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	1649.284886	1.0	41265.038898	0.0	16.61	<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	512.076667	1.0	13680.388309	0.000000e+00	6.74
<b>Residual</b>	7194.056785	179995.0	NaN	NaN	72.45	<b>Residual</b>	6737.472479	179995.0	NaN	NaN	88.72

443

Gradient Forests

LFMM2

	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
<i>P<sub>c</sub>QTN<sub>temp</sub></i>	350.833368	1.0	9184.348121	0.0	3.69	<i>P<sub>c</sub>QTN<sub>temp</sub></i>	256.979852	1.0	8357.260709	0.0	3.72
<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	210.540691	1.0	5511.673549	0.0	2.21	<i>P<sub>c</sub>QTN<sub>Env2</sub></i>	173.645491	1.0	5647.137802	0.0	2.51
<i>P<sub>c</sub>Neut<sub>temp</sub></i>	375.213154	1.0	9822.578289	0.0	3.94	<i>P<sub>c</sub>Neut<sub>temp</sub></i>	718.499204	1.0	23366.365591	0.0	10.40
<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	1702.524610	1.0	44569.816059	0.0	17.89	<i>P<sub>c</sub>Neut<sub>Env2</sub></i>	227.994551	1.0	7414.627593	0.0	3.30
<b>Residual</b>	6875.637914	179995.0	NaN	NaN	72.26	<b>Residual</b>	5534.718859	179995.0	NaN	NaN	80.08

444

445 **Table S2** Results from Type II ANOVAs from regressing the proportion of clinal  
 446 QTNs (*cor\_TPR\_tmp* and *cor\_TPR\_sal*) and clinal neutral alleles  
 447 (*cor\_FPR\_temp\_neutSNPs*, *cor\_FPR\_sal\_neutSNPs*) on offset performance (see  
 448 Equation 2 of the main text). Code to create these tables can be found in SC 02.02.05.

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RONA-sal_opt						RONA-temp_opt					
	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
all	13.066056	1.0	939.445081	2.543708e-166	34.04	all	0.086501	1.0	8.324025	0.003959	0.18
final_LA	0.321152	1.0	23.090730	1.673759e-06	0.84	final_LA	30.251825	1.0	2911.137332	0.000000	61.72
Residual	24.993162	1797.0	NaN	NaN	65.12	Residual	18.673983	1797.0	NaN	NaN	38.10

lfmm2						GF					
	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
all	3.673454	1.0	873.091702	9.874623e-157	8.64	all	8.505224	1.0	749.172746	3.638160e-138	10.54
final_LA	31.290136	1.0	7436.913517	0.000000e+00	73.58	final_LA	51.787680	1.0	4561.657568	0.000000e+00	64.18
Residual	7.560714	1797.0	NaN	NaN	17.78	Residual	20.401019	1797.0	NaN	NaN	25.28

rda-nocorr						rda-structcorr					
	sum_sq	df	F	PR(>F)	perc_sum_sq		sum_sq	df	F	PR(>F)	perc_sum_sq
all	10.763364	1.0	827.118396	6.020460e-150	12.54	all	3.192007	1.0	78.914320	1.537690e-18	4.20
final_LA	51.668736	1.0	3970.521018	0.000000e+00	60.21	final_LA	0.062105	1.0	1.535395	2.154664e-01	0.08
Residual	23.384518	1797.0	NaN	NaN	27.25	Residual	72.686890	1797.0	NaN	NaN	95.71

452 **Table S3** Results from Type II ANOVAs regressing two factors - degree of local  
 453 adaptation (final\_LA) and levels of isolation-by-environment in *all* marker sets) on  
 454 offset performance. Code to create these tables can be found in SC 02.02.11.

455

Nuisance Level	<i>Adaptive</i> models	<i>All</i> models	<i>Neutral</i> models
1-trait 1-nuisance	45/45	45/45	45/45
1-trait 3-nuisance	45/45	43/45	38/45
1-trait 4-nuisance	43/45	36/45	35/45
2-trait 2-nuisance	120/180	119/180	119/180
2-trait 3-nuisance	140/180	119/180	119/180

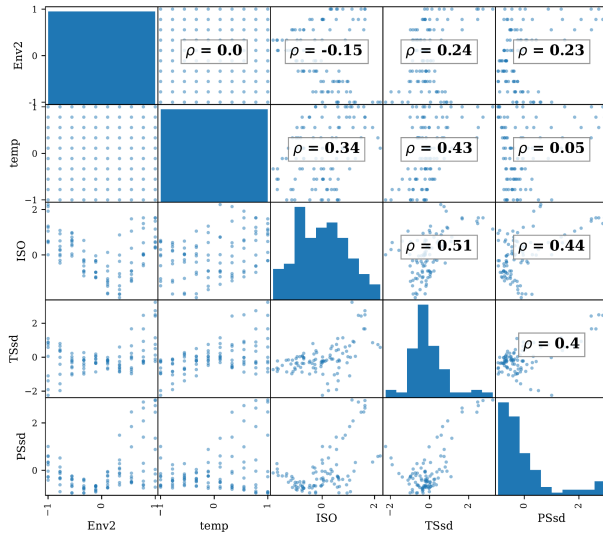
456 **Table S4** Gradient Forests (GF) sometimes incorrectly identifies the environments  
457 driving adaptation. Shown are the proportions of simulation levels ( $N$  1-trait = 45  
458 levels;  $N$  2-trait = 180 levels; one replicate each) where weighted feature importance  
459 output from GF correctly identified the adaptive environments in the top-most  
460 ranks. If at least one nuisance environment was ranked above an adaptive  
461 environment this was counted as incorrect. Data used to create this table is from  
462 the GF models output from the *Nuisance Environment* workflow. Code used to  
463 create this table can be found in SC 02.10.02.

464 **Supplemental Figures**

465 Figs. S1-S3 are in Supplemental Note S1.

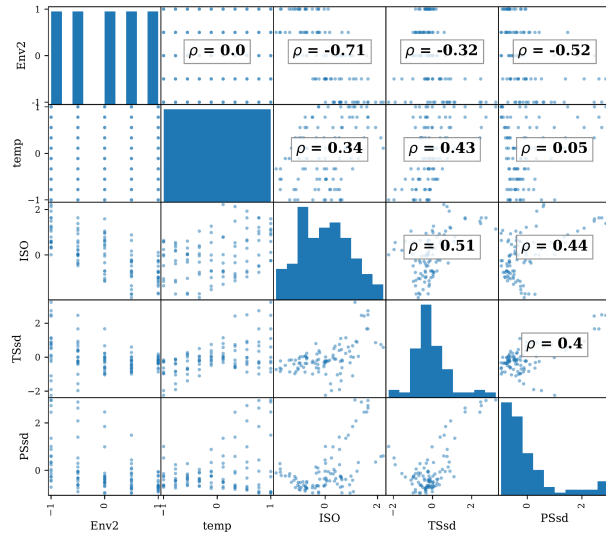


Stepping Stone - Clines



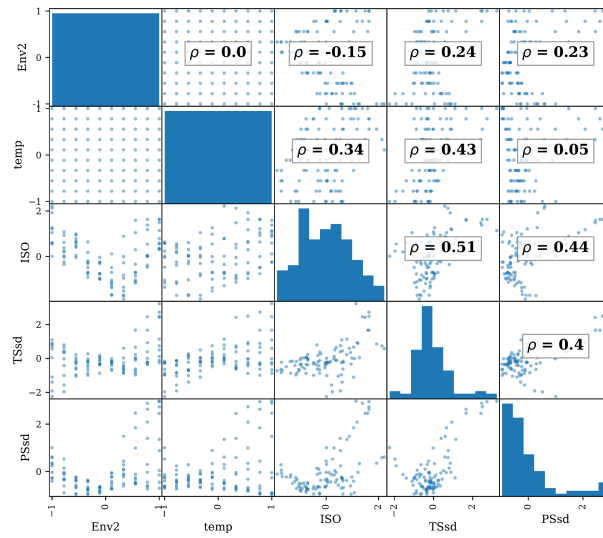
466

Stepping Stone - Mountain



467

Estuary - Clines



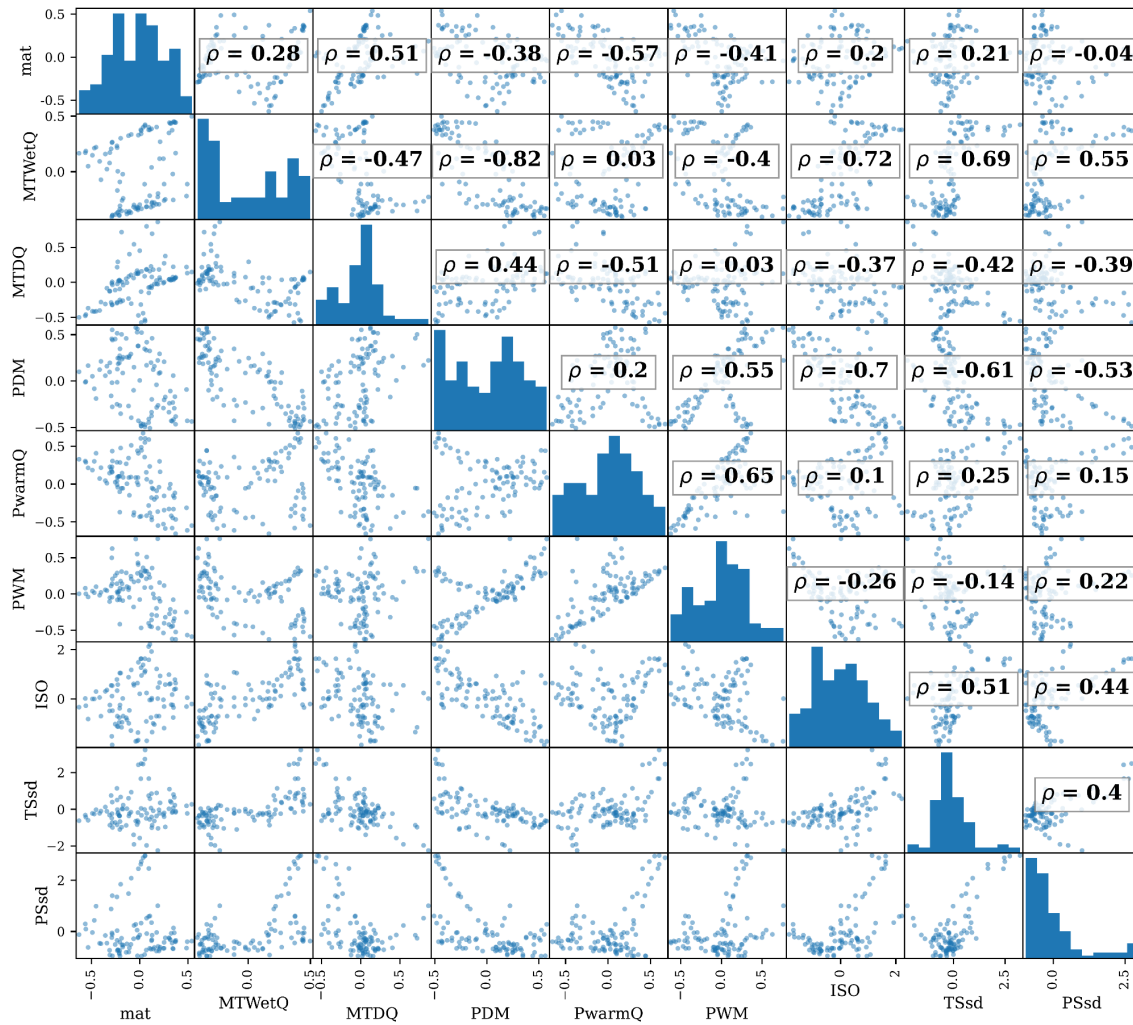
482

**Fig S4** Correlation (Spearman's rho) among environmental variables faceted by landscape. On-diagonal entries are histograms of environmental values, off-diagonal entries are scatter plots between pairwise variables. Included are environmental variables from 1-trait (*temp*), 2-trait (*temp*, *Env2*), and 6-trait simulations (MAT, MTWetQ, MTDQ, PDM, PwarmQ, PWM), as well as nuisance environmental variables (ISO, PSsd, TSsd). Note *Estuary - Clines* and *Stepping Stone - Clines* have the same correlation structure; *Stepping Stones - Mountain* only differs from these two landscapes with *Env2*. Figure continues

483 on the next page. Code to create these figures can be found in SC 02.07.02.11.

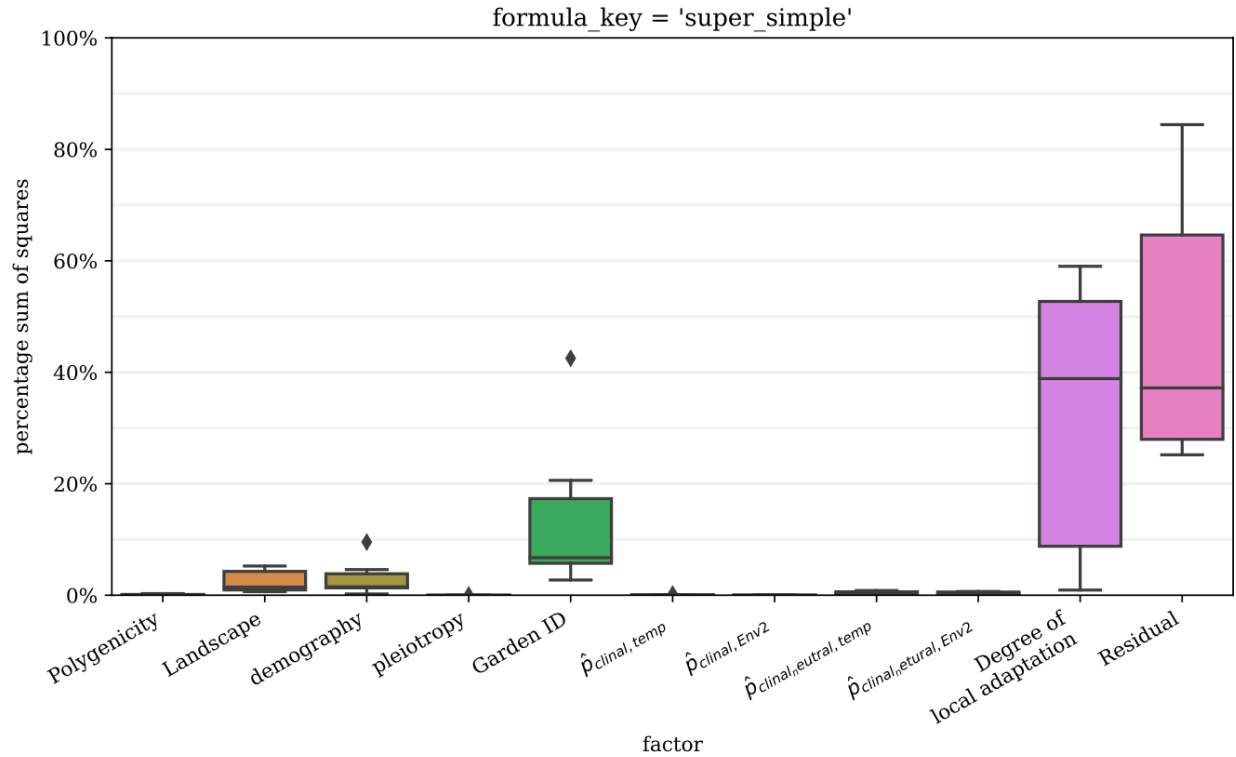
484 (Fig S4 continued)

485 6-trait



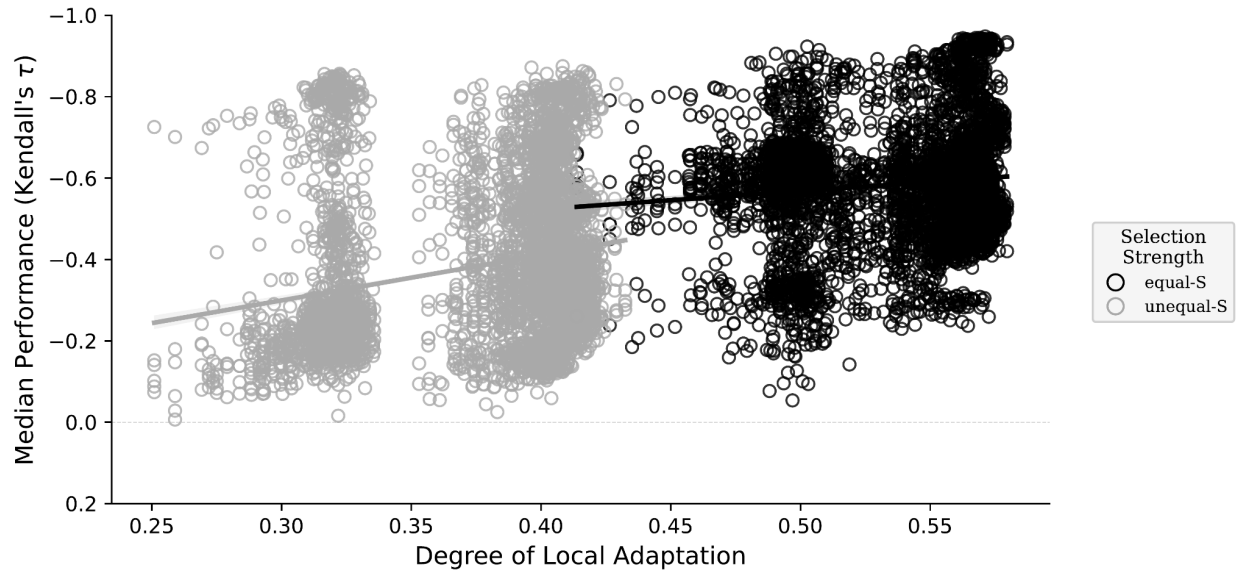
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487 Fig S5 is in Supplemental Note S3



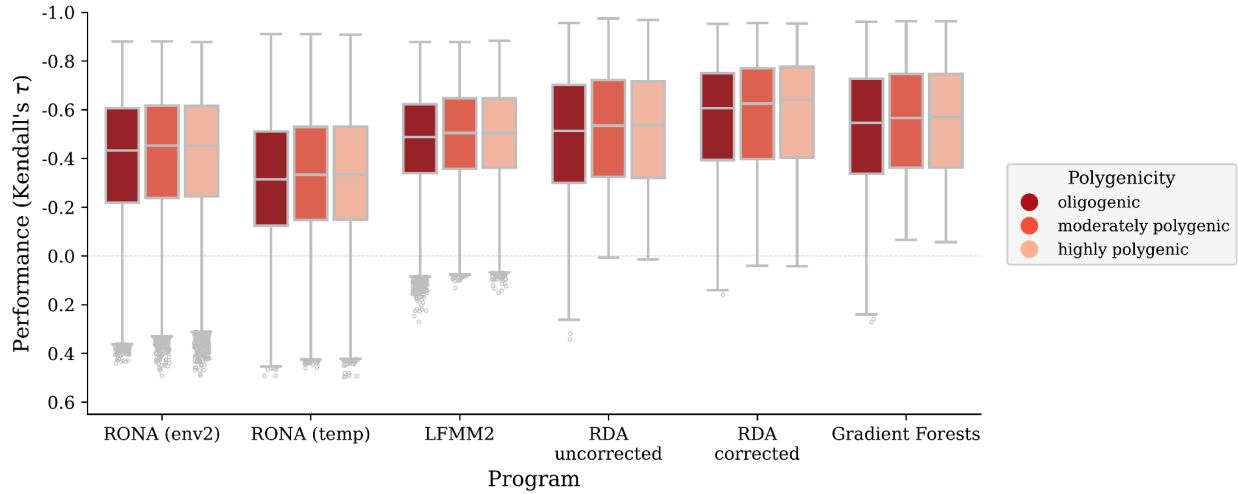
488

489 **Fig S6** Percent sum of squares of the various factors from the ANOVA model in  
 490 Table S1. Boxplots are created from the percent sum of squares from each method's  
 491 individual ANOVA model. Data included in this figure are from models trained  
 492 using all markers and simulations with two selective environments with  
 493 performance evaluated in all 100 common gardens. Code to create this table is in  
 494 02.02.01.



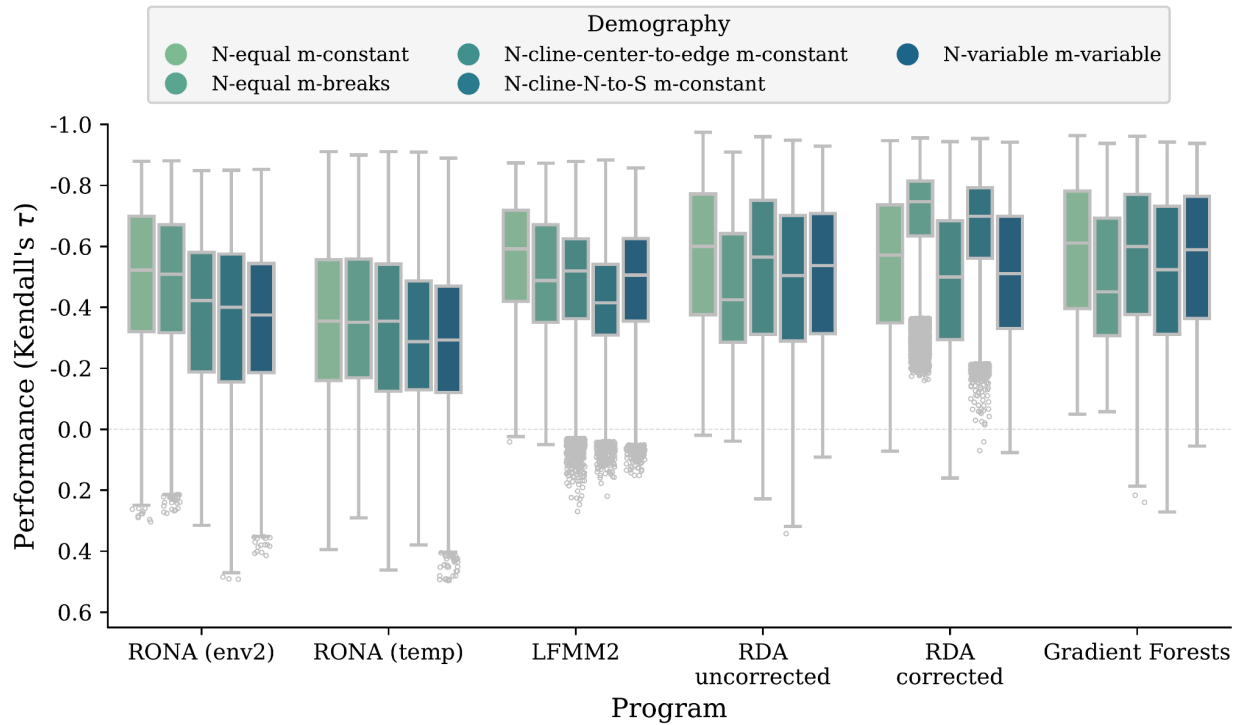
495

496 **Fig S7** Effect of the degree of local adaptation (x-axes) on method performance (y-  
 497 axes) colored by the relative strength of selection on the two traits. Shown are the  
 498 linear relationships between the median validation scores (circles, taken from  
 499 validation scores across all 100 common gardens on the landscape) and the  
 500 simulation's mean level of local adaptation (taken across all 100 populations). Data  
 501 included in this figure are from models trained using all markers and simulations  
 502 with two selective environments. Code to create this figure can be found in SC  
 503 02.02.02.



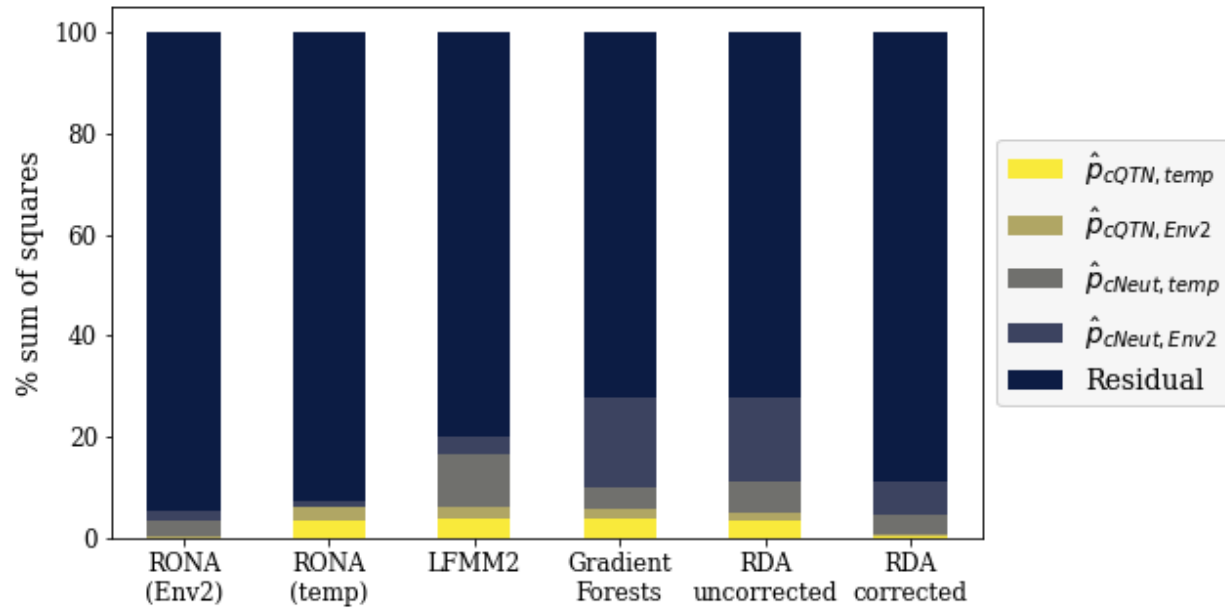
504

505 **Fig S8** Effect of polygenicity on performance of offset methods trained using all  
506 markers on simulations with two adaptive traits. Code to create this figure can be  
507 found in SC 02.02.01.



508

509 **Fig S9** Effect of demography on performance of offset methods trained using all  
 510 markers on simulations with two adaptive traits. Code to create this figure can be  
 511 found in SC 02.02.01.



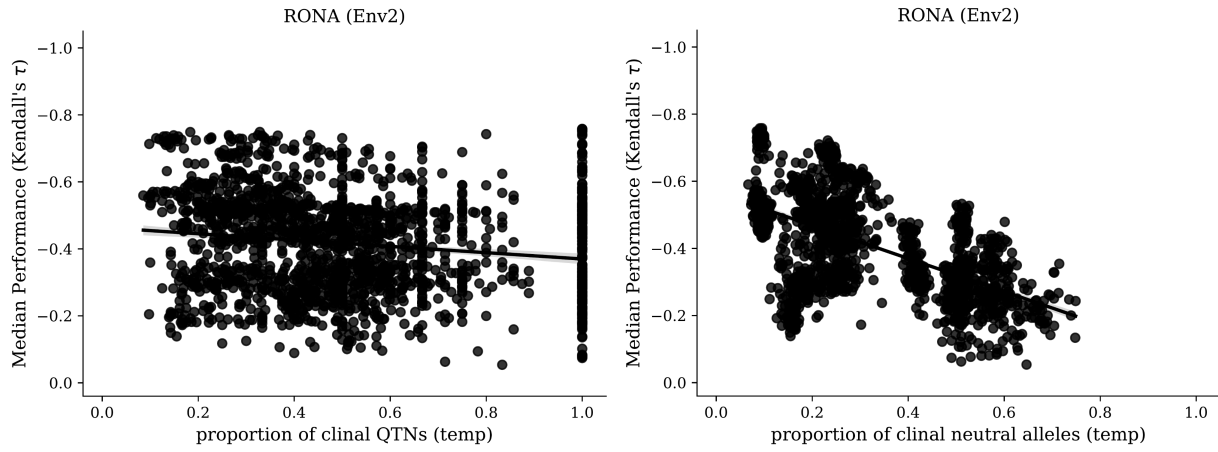
512

513 **Fig S10** Stacked bar plot of the percent sum of squares from Type II ANOVAs from  
 514 regressing the proportion of clinal QTNs and clinal neutral alleles on offset  
 515 performance (see Equation 2 of the main text). Code to create this table is in  
 516 02.02.05.

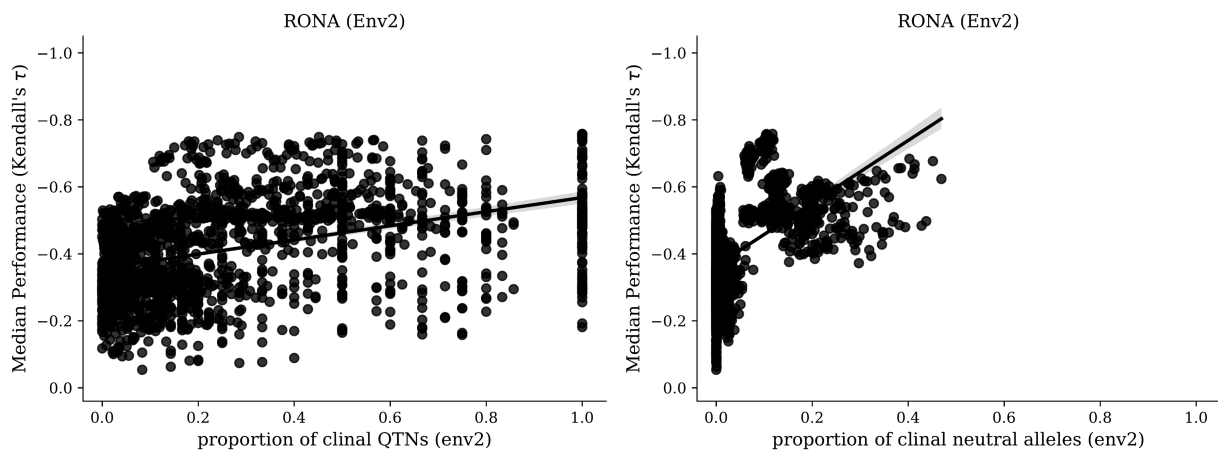


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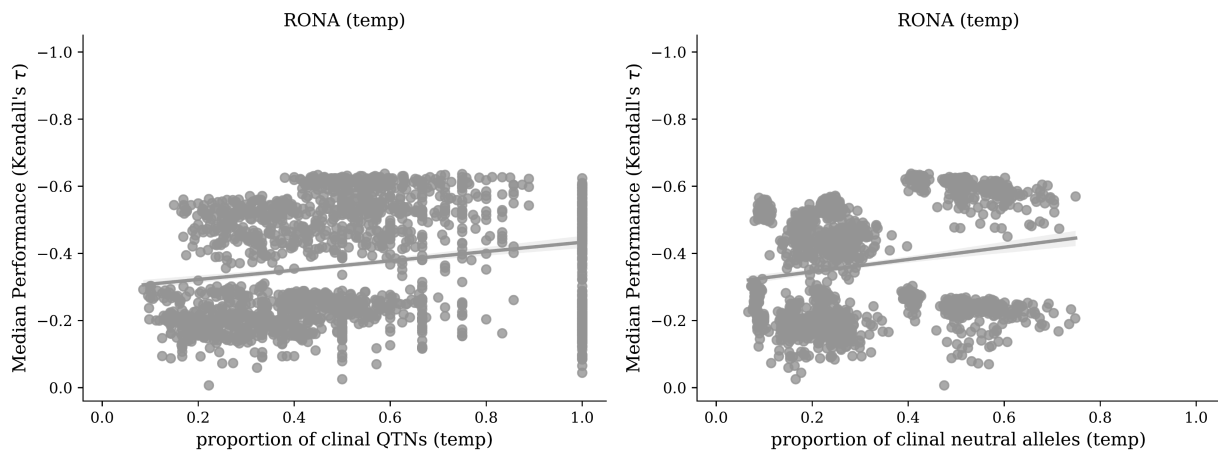
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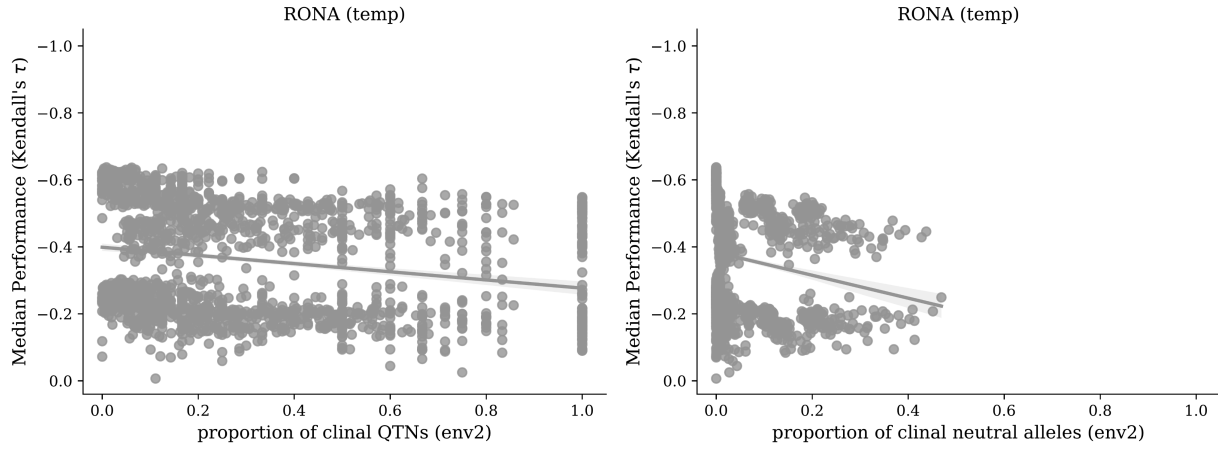


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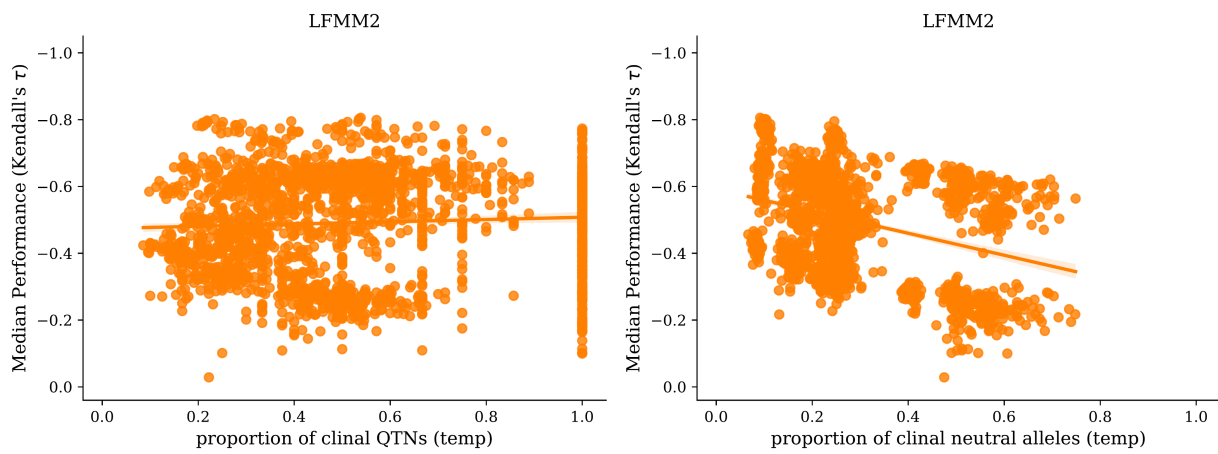


Supplement - *Lind, Lotterhos, and the limits of genomic offsets*

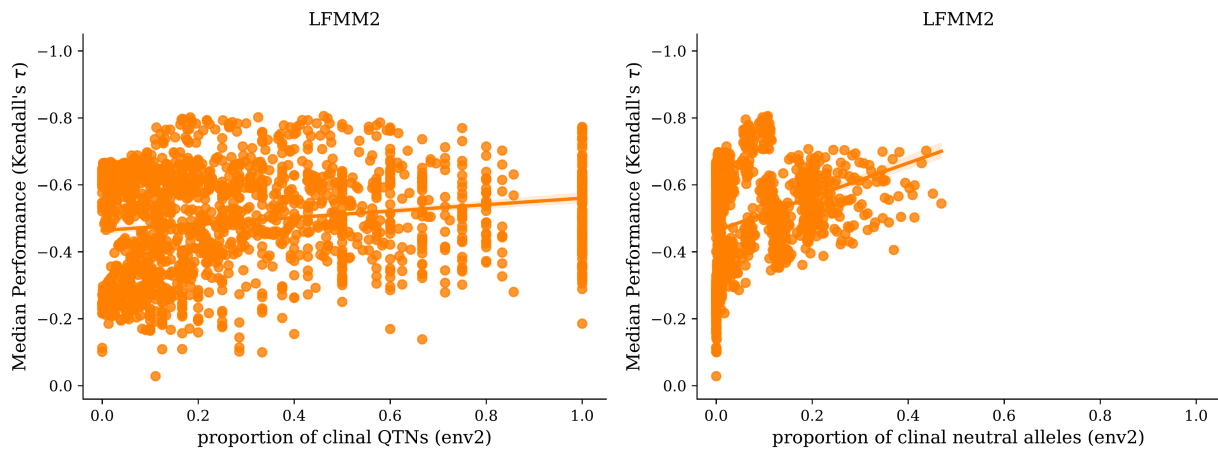
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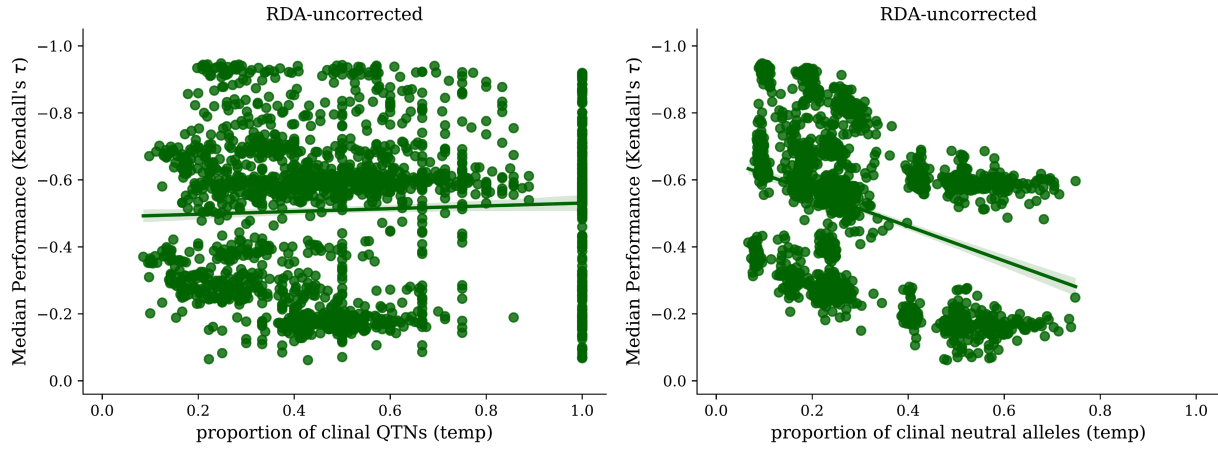


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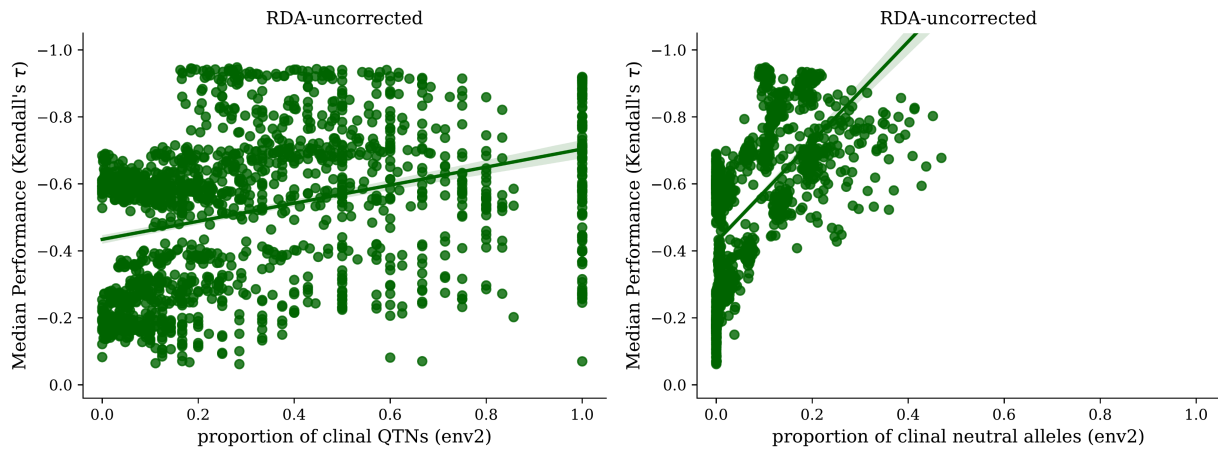


Supplement - *Lind, Lotterhos, and the limits of genomic offsets*

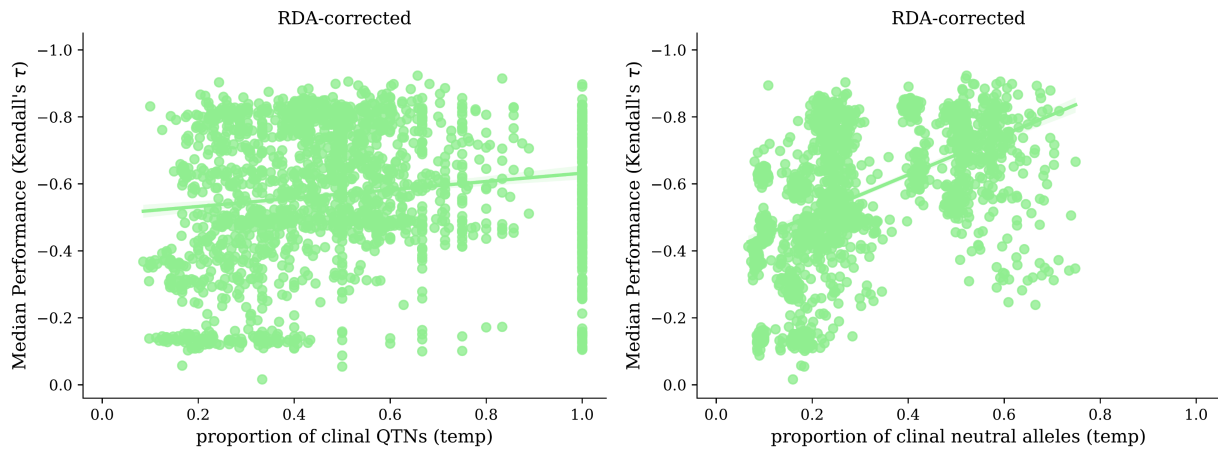
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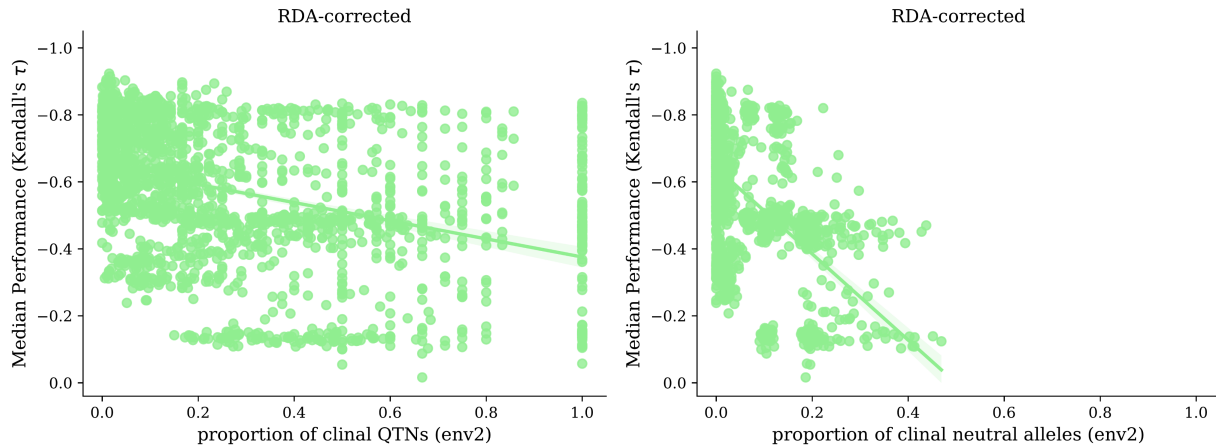


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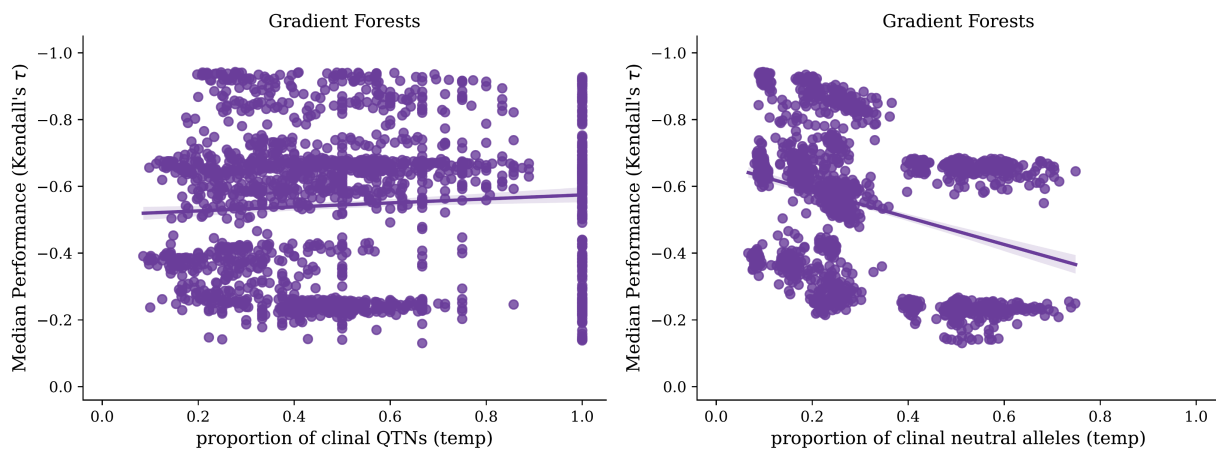


Supplement - *Lind, Lotterhos, and the limits of genomic offsets*

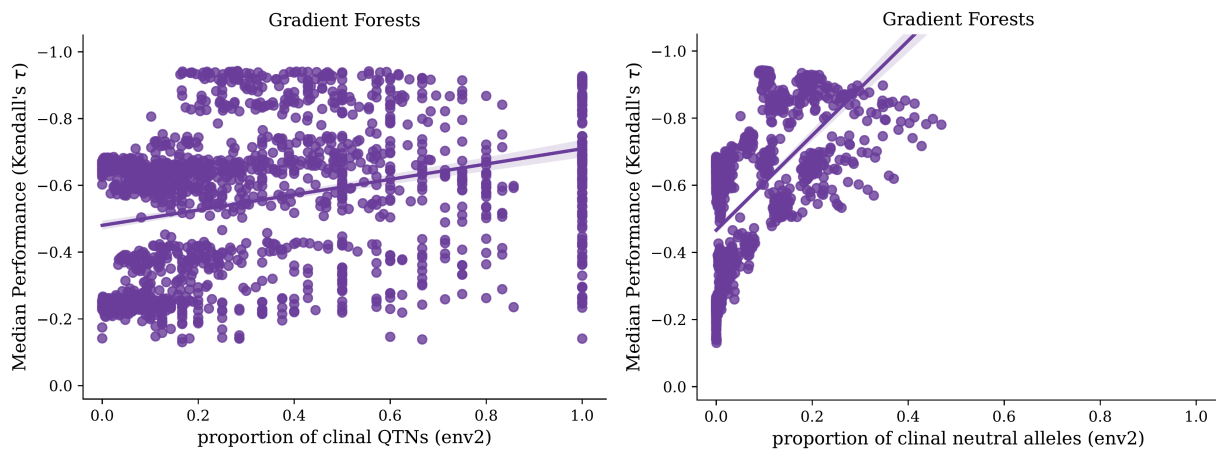
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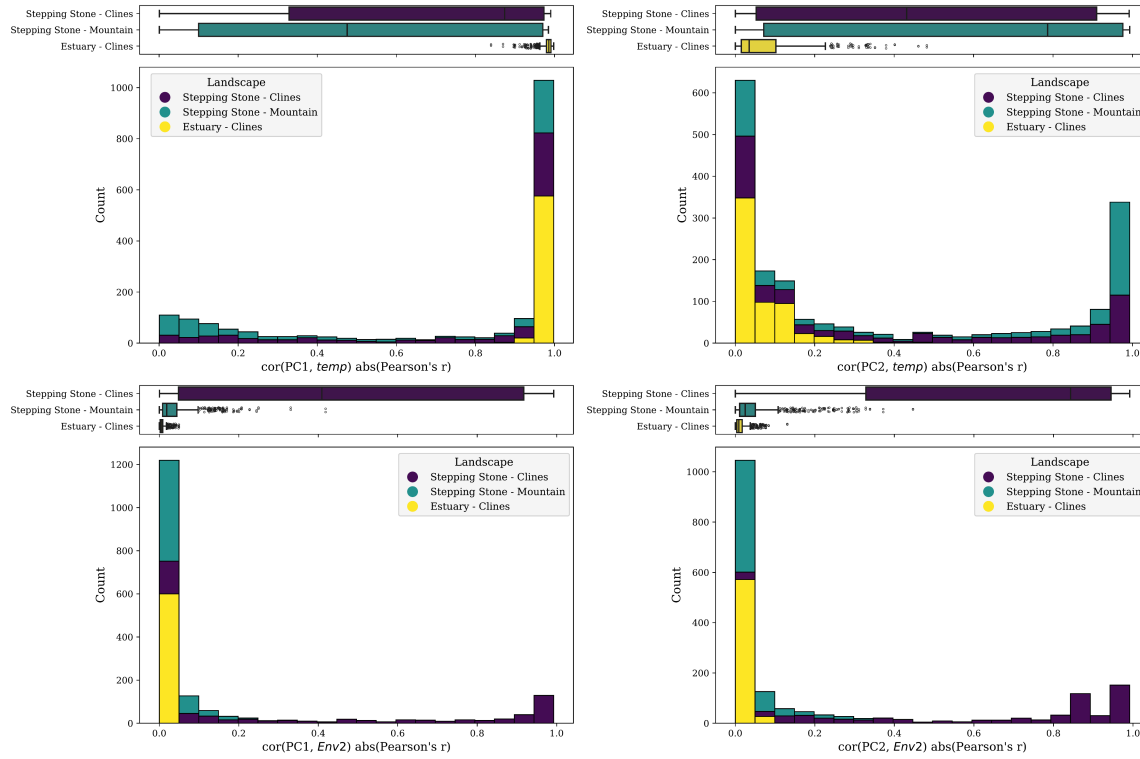
527



528



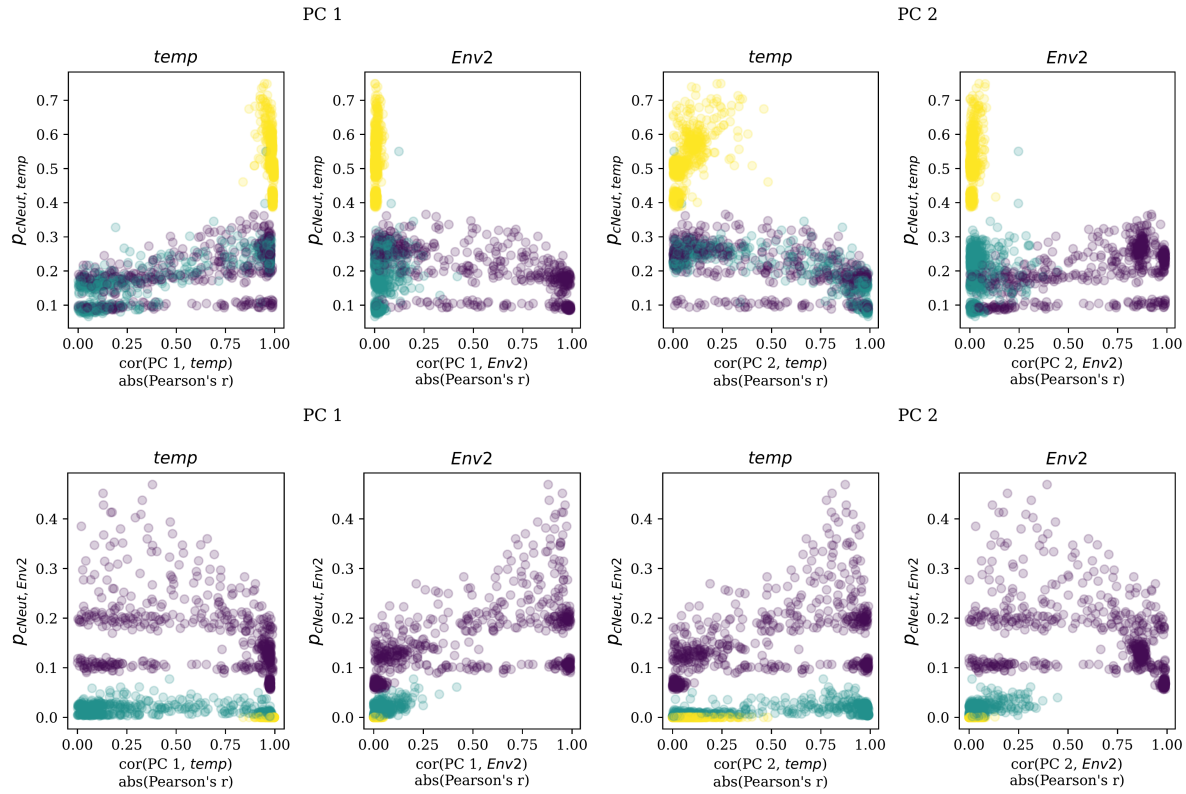
529 **Fig S11** Impact on method performance (y-axes) from the proportion of QTNs with  
 530 clinal relationships with *temp* (first column) or *Env2* (second column). Model  
 531 performance is quantified as Kendall's rank correlation between offset and fitness;  
 532 shown are median values from scores from 10 replicates per seed (100 common  
 533 gardens for each replicate). Data included in this figure is from evaluation of 2-trait  
 534 simulations using *all* markers. Code to create these figures can be found in 02.01.03.



535

536

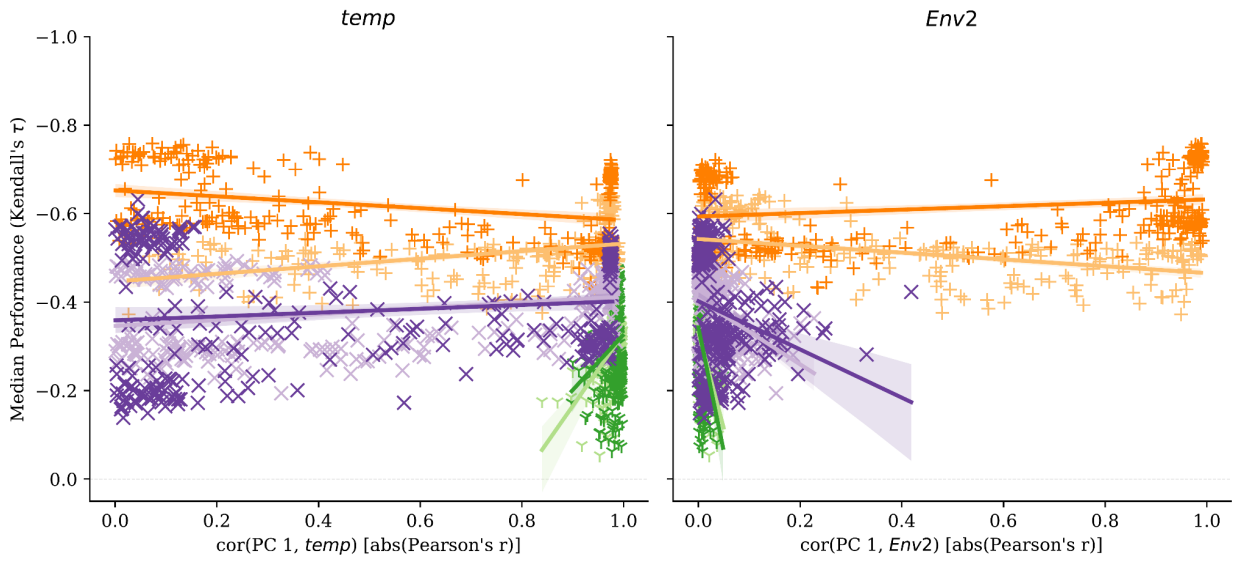
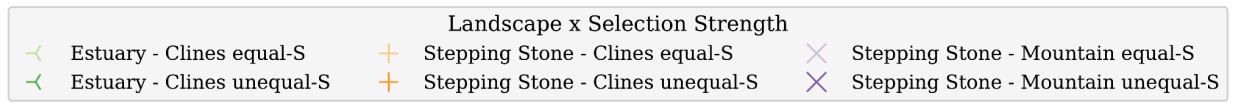
537 **Fig S12** Stacked bar plot showing correlation between environmental variables  
 538 (rows) and axes of population genetic structure (Principal Component Analysis  
 539 axes [PC axes]; columns). Data included in this figure is from all 2-trait simulations.  
 540 Code to create this figure can be found in SC 02.10.03.



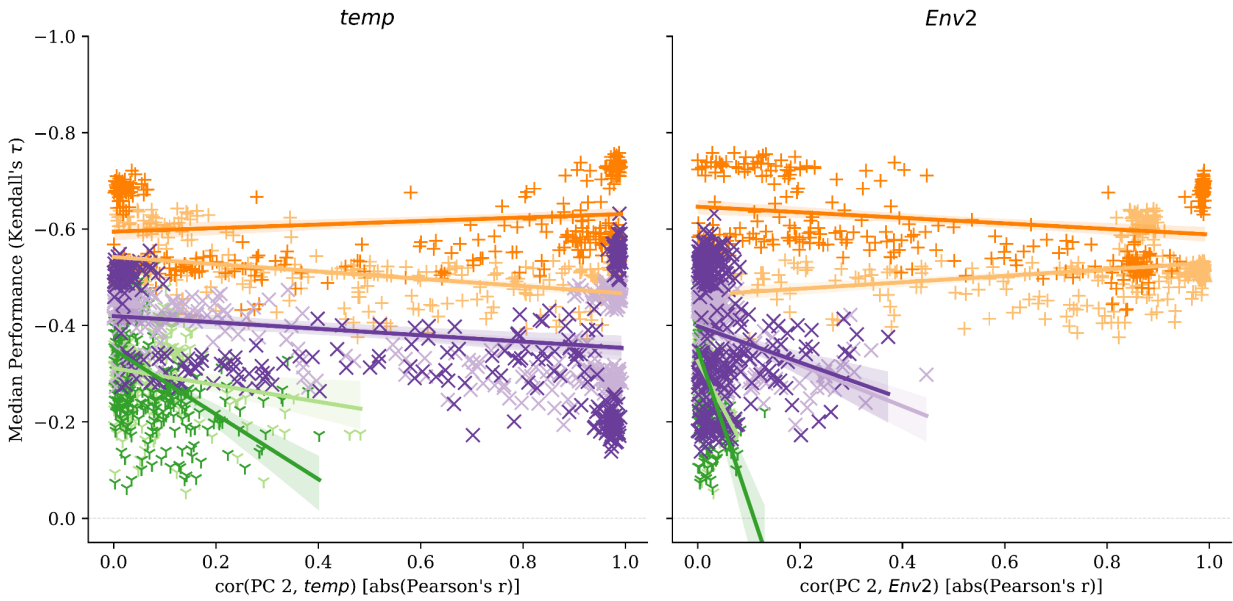
543 **Fig S13** Relationship between the proportion of clinal neutral loci for *temp* (y-  
 544 axes, first row) or *Env2* (y-axes, second row) with the strength of the relationship  
 545 between environmental variables and axes of population genetic structure. Purple  
 546 = *Stepping Stone - Clines*; teal = *Stepping Stone - Clines*; yellow = *Estuary - Clines*.  
 547 Data included in this figure is from all 2-trait simulations. Code to create this figure  
 548 can be found in 02.10.03.

549 (Fig S14)

RONA (Env2)



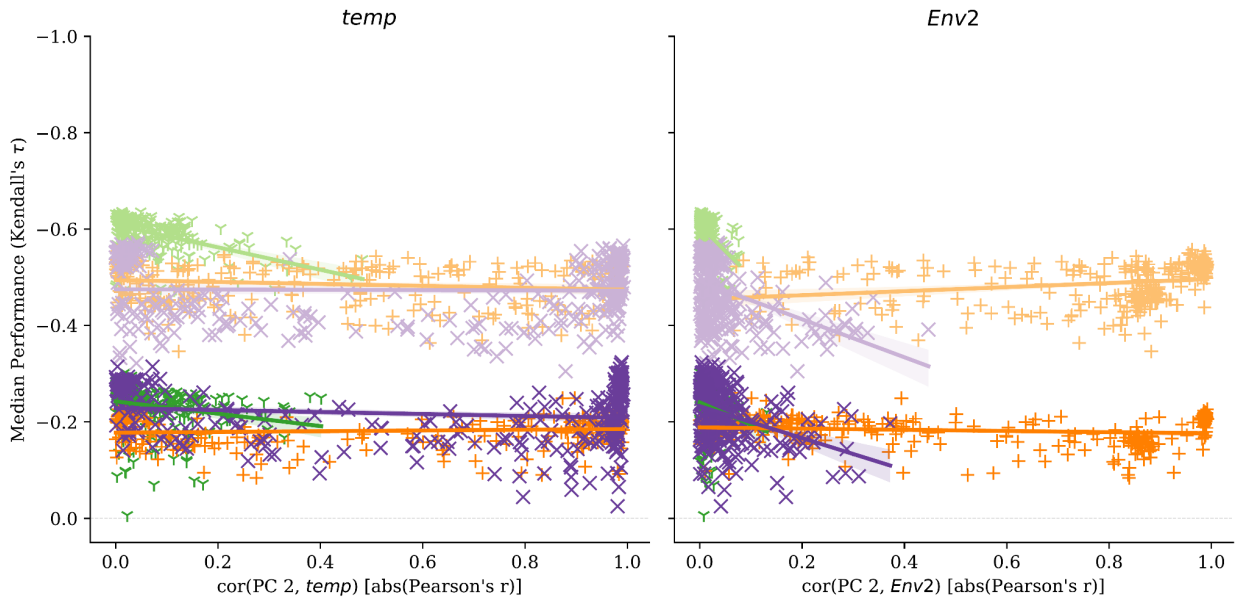
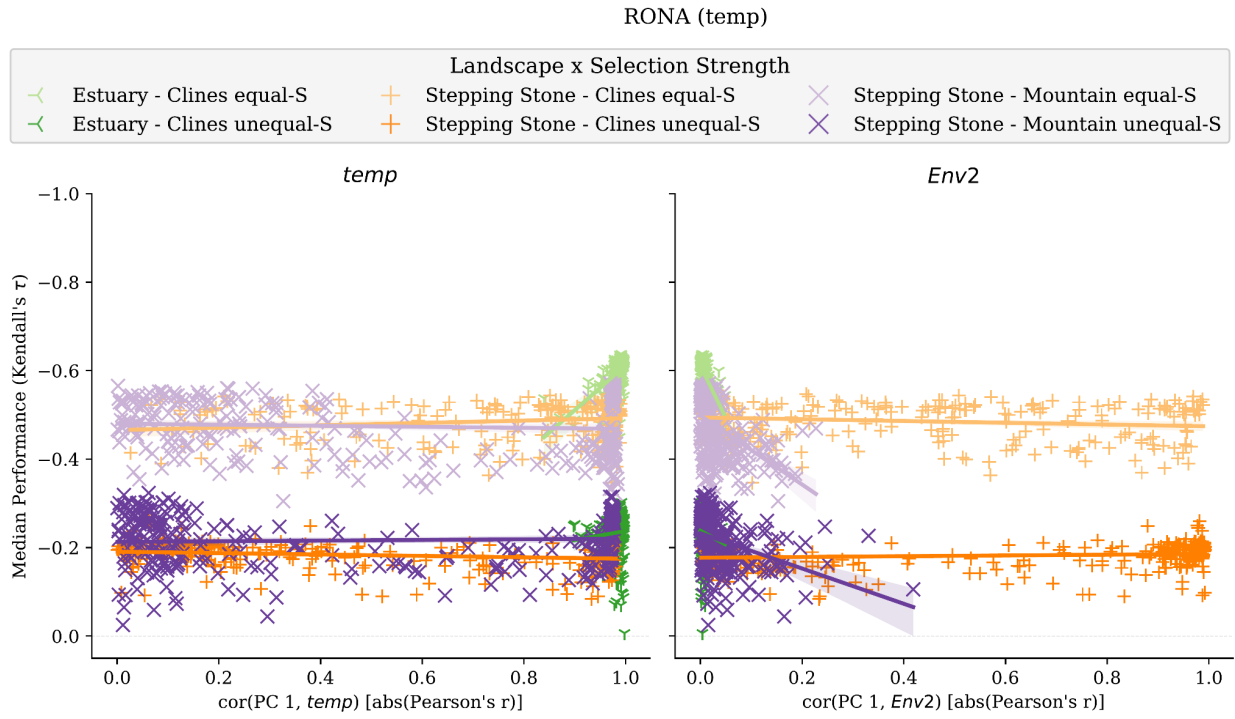
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551



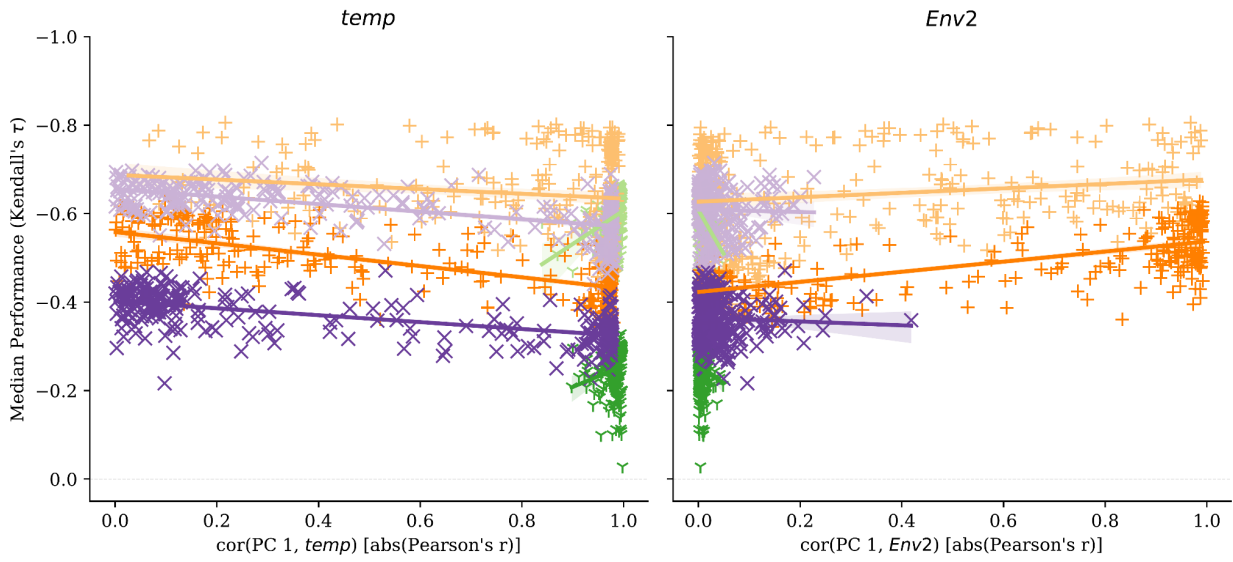
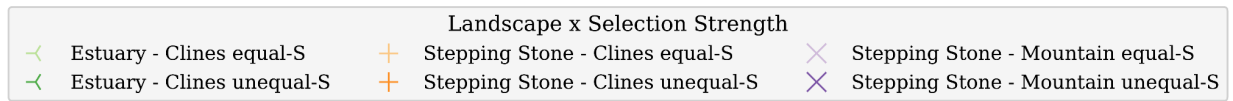
552 (Fig S14 continued)



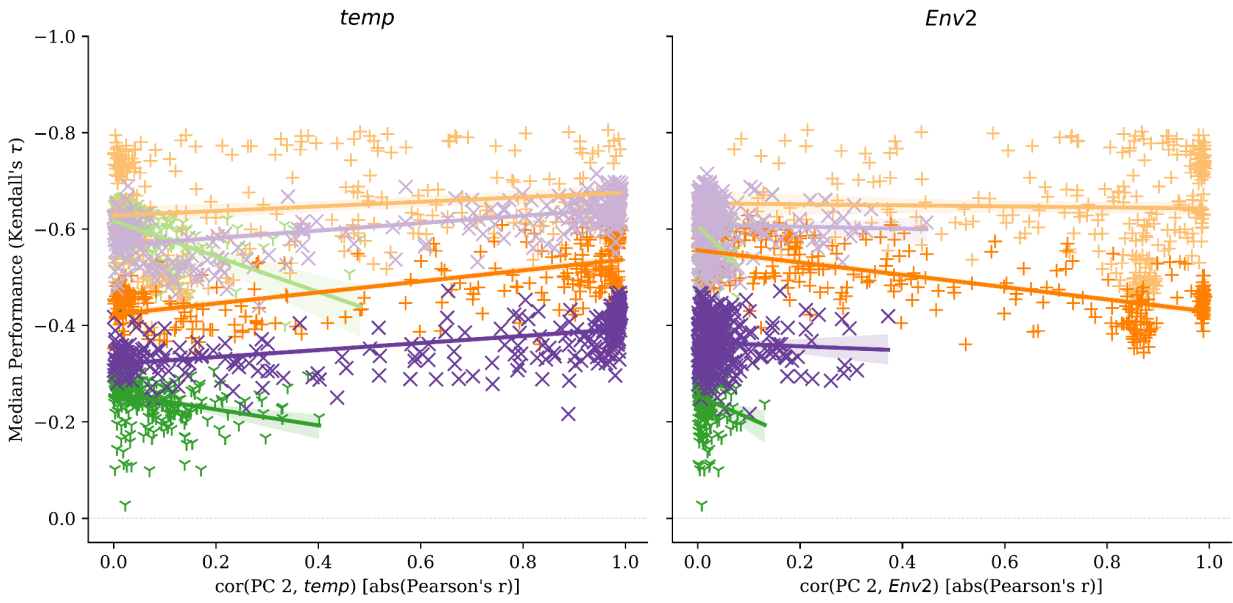


555 (Fig S14 continued)

LFMM2

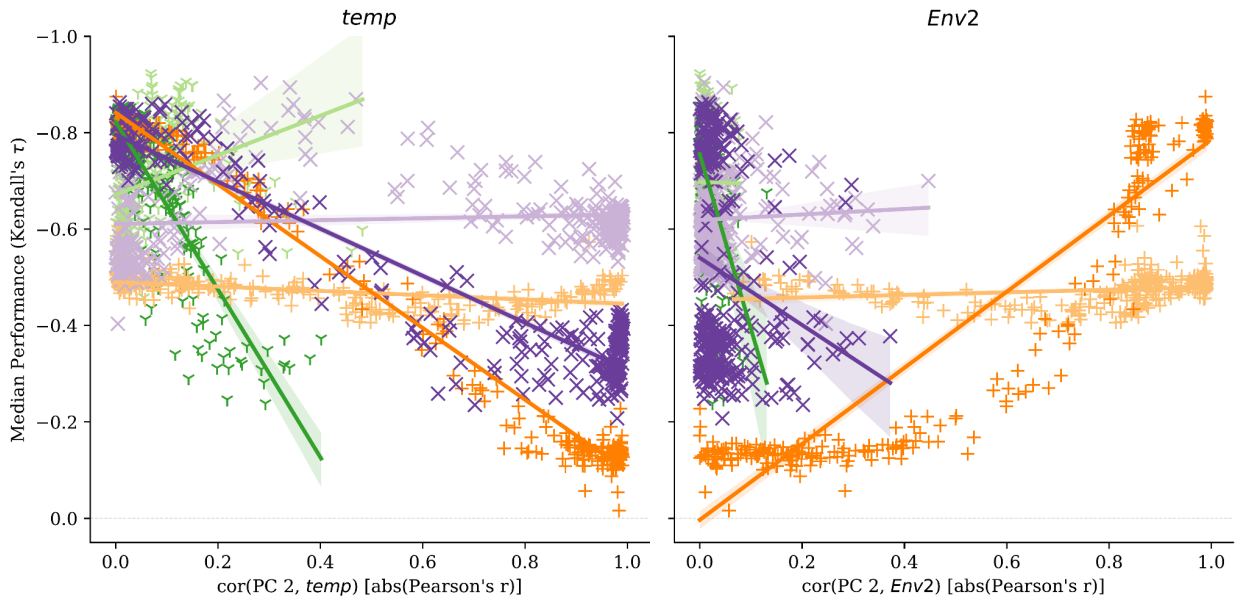
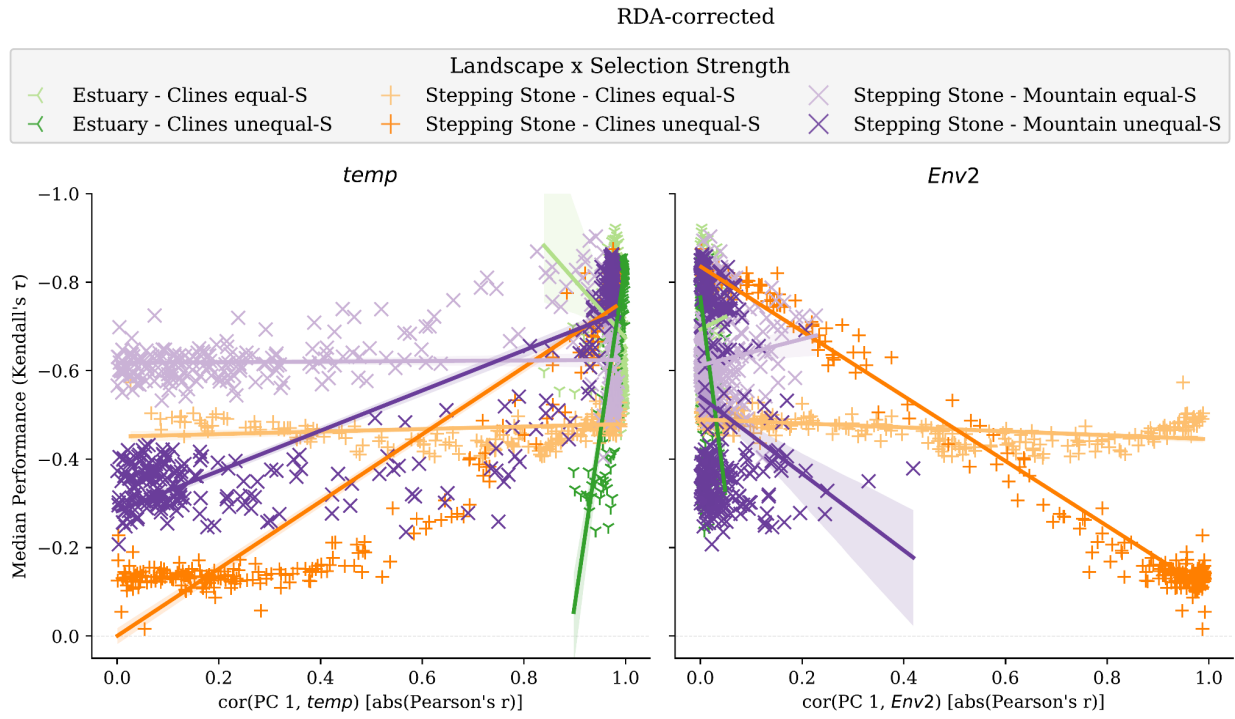


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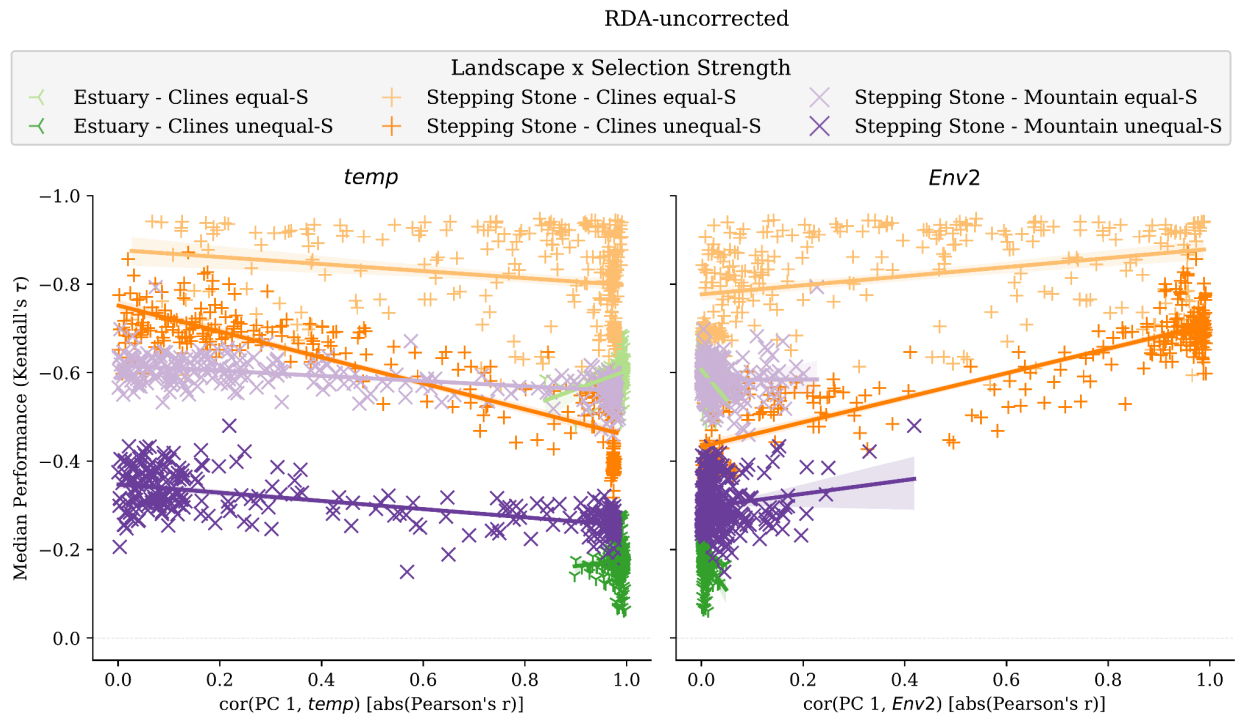


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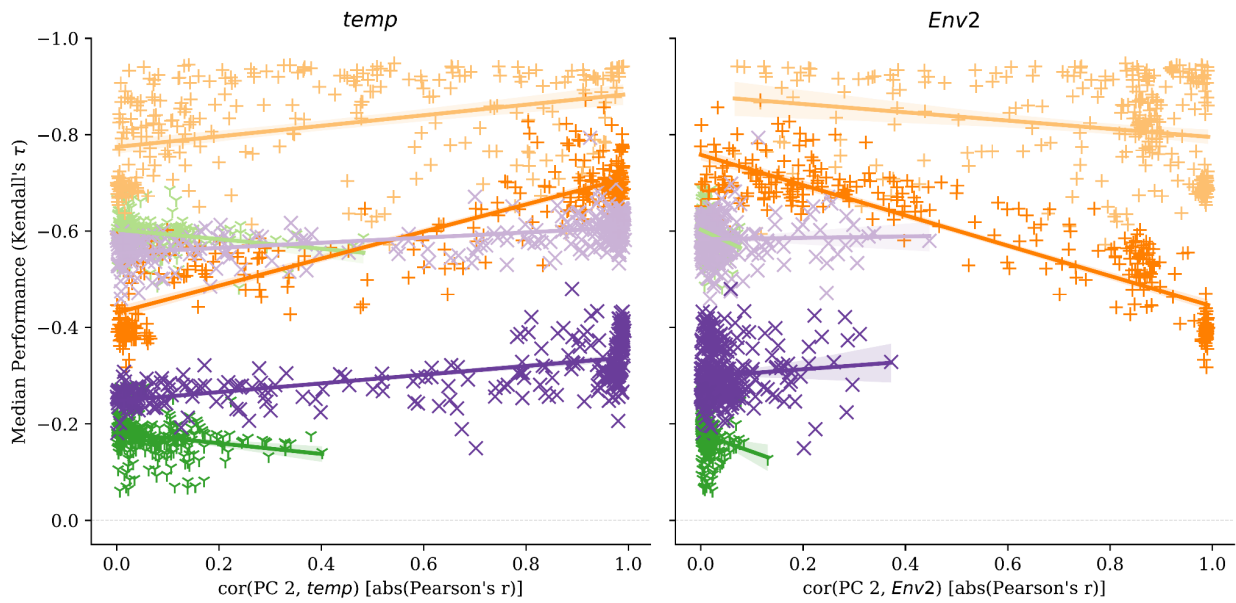
558 (Fig S14 continued)



561 (Fig S14 continued)

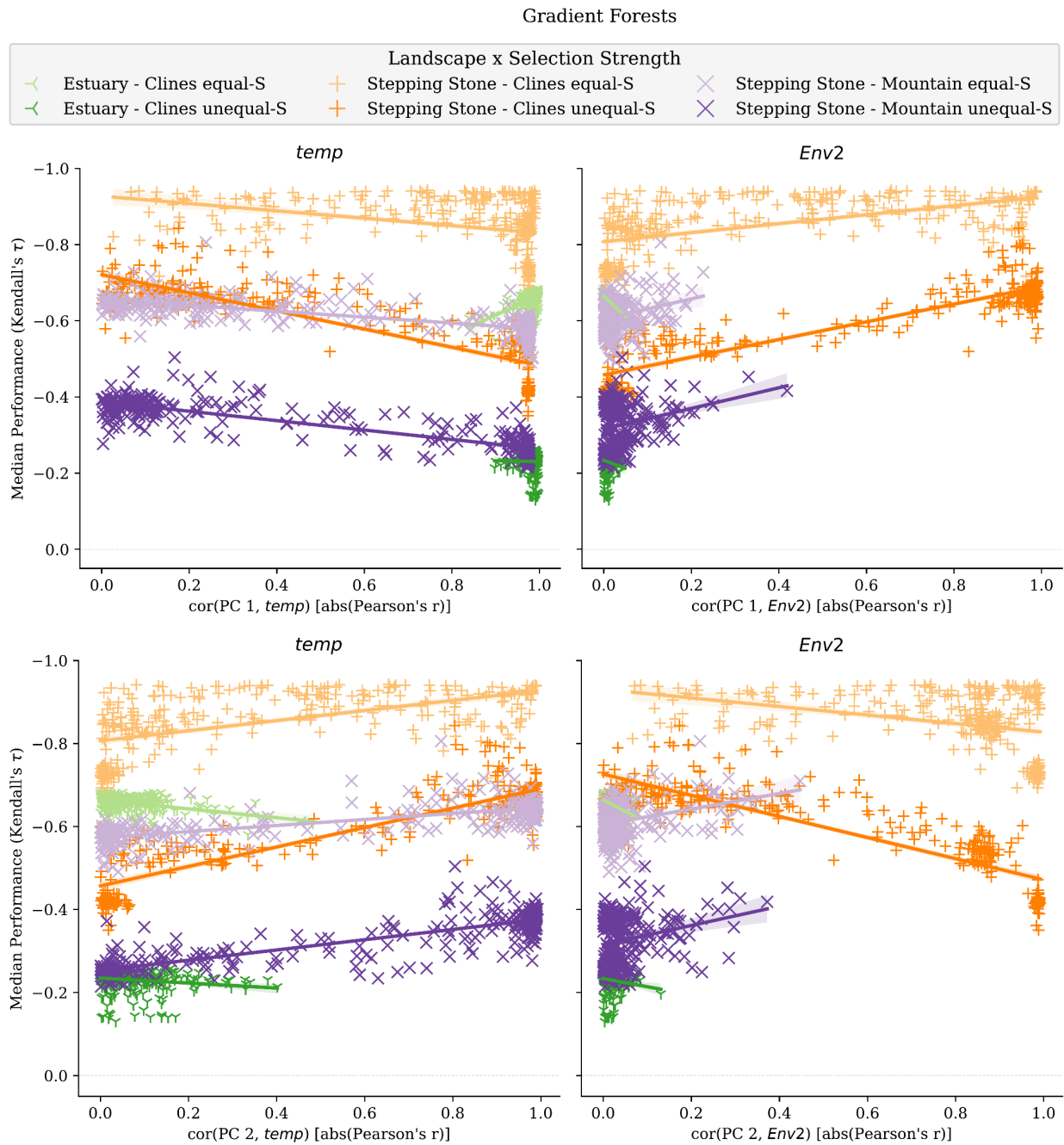


562



563

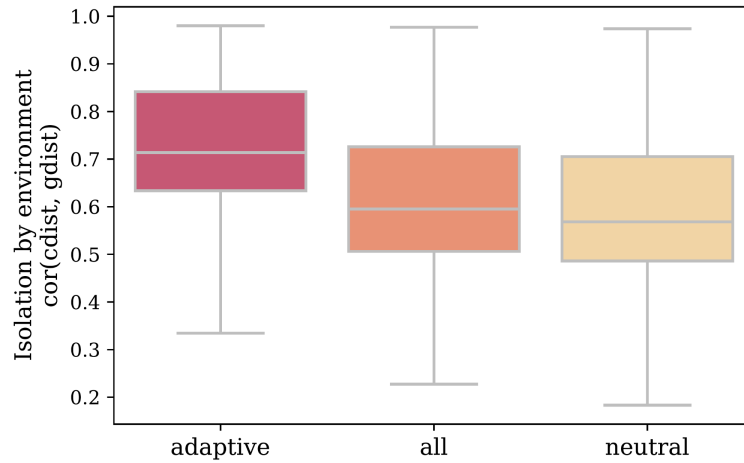
564 (Fig S14 continued)



565

566

567 **Fig S14** Relationship between median performance and absolute correlation  
 568 (Pearson's  $r$ ) between environmental variables and axes of population genetic  
 569 structure (principal component analysis axes). Each subfigure is for a different  
 570 method (see panel titles). Data used in this figure is from 2-trait simulations. Code  
 571 to create this figure can be found in SC 02.10.03.



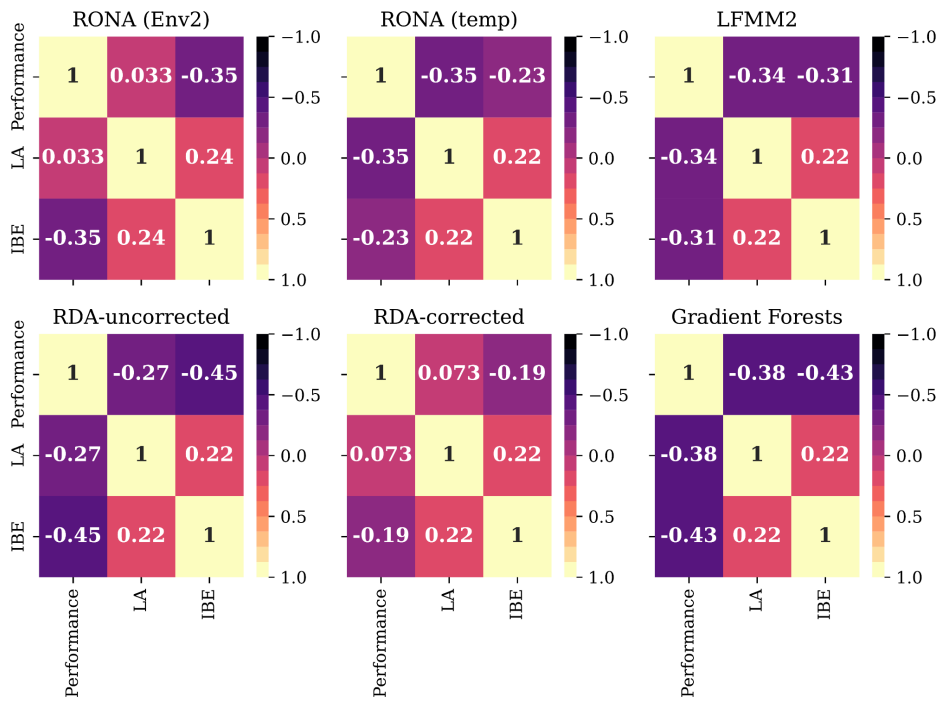
572

573 **Fig S15** *Adaptive* markers contain greater levels of isolation-by-environment  
574 (IBE) than other marker sets. *IBE* is quantified as Spearman's rank correlation  
575 between population pairwise  $F_{ST}$  and Euclidean distance of adaptive environments.  
576 Code to create this figure can be found in SC 02.02.10.

577 (Fig S16)

578

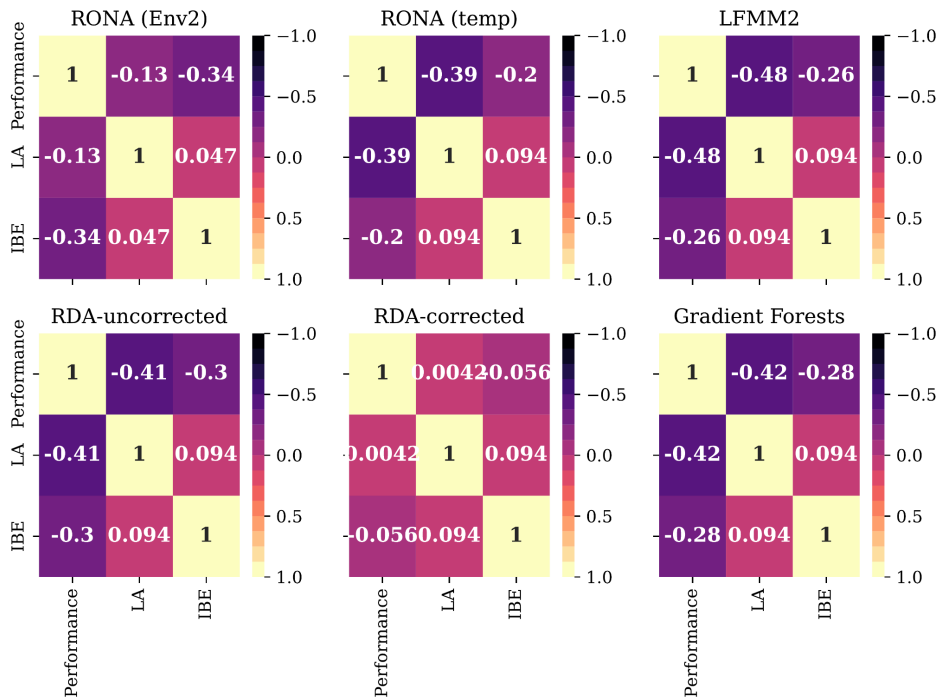
A) *adaptive markers*



579

580

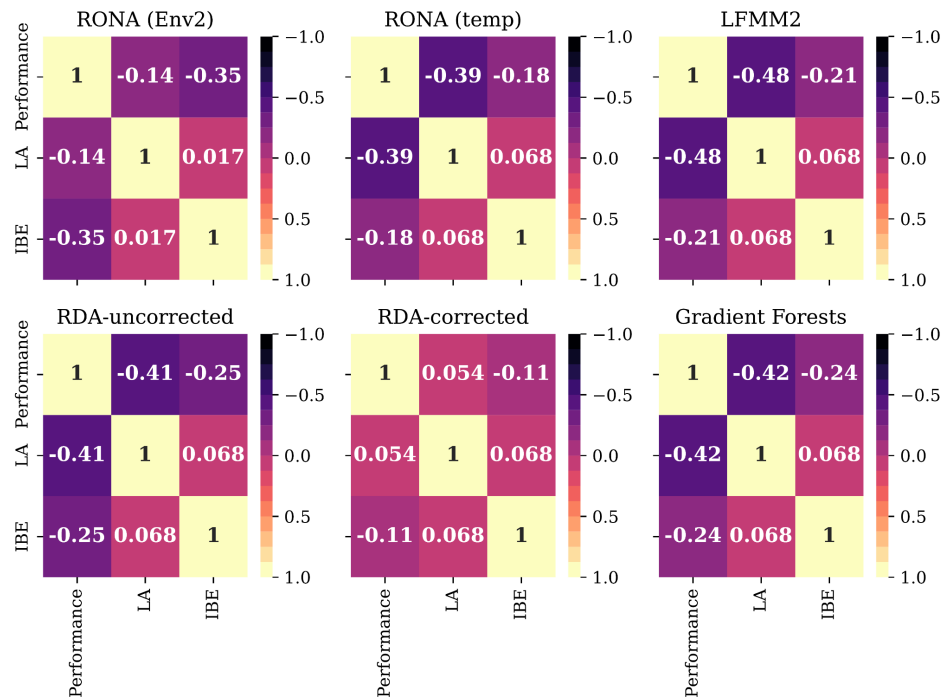
B) *all markers*



581

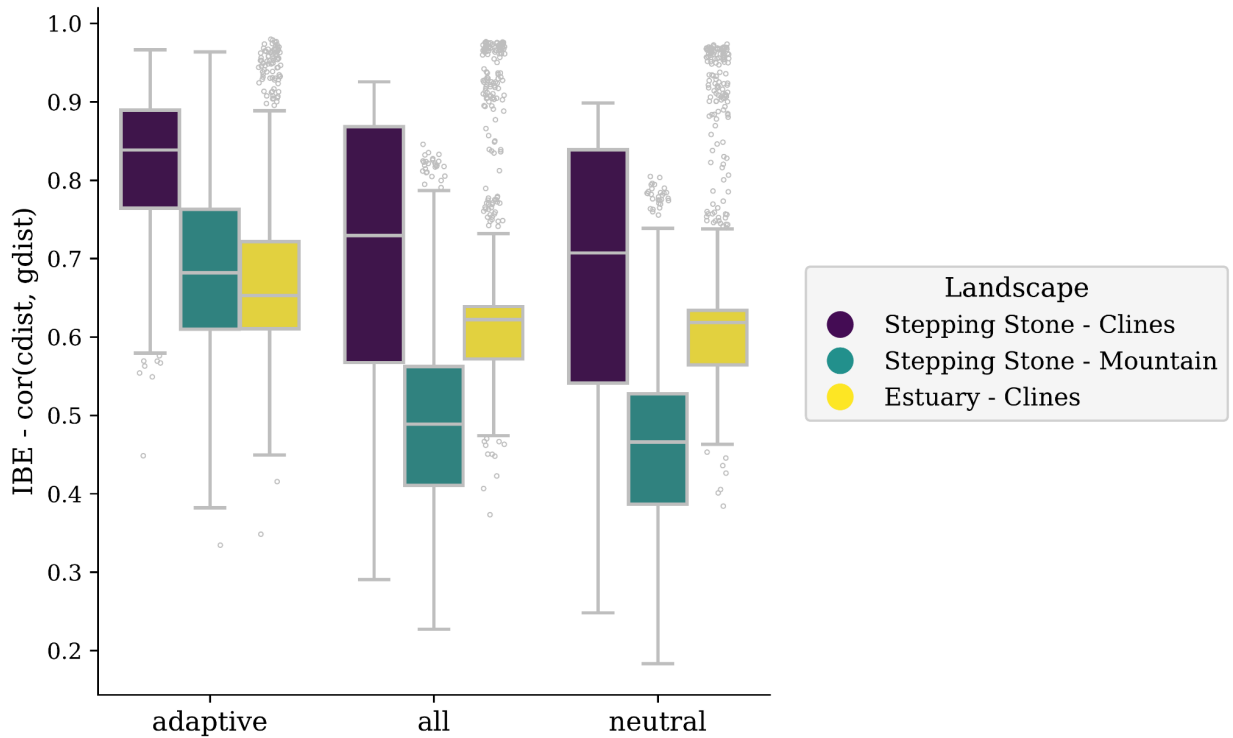
582

## C) neutral markers

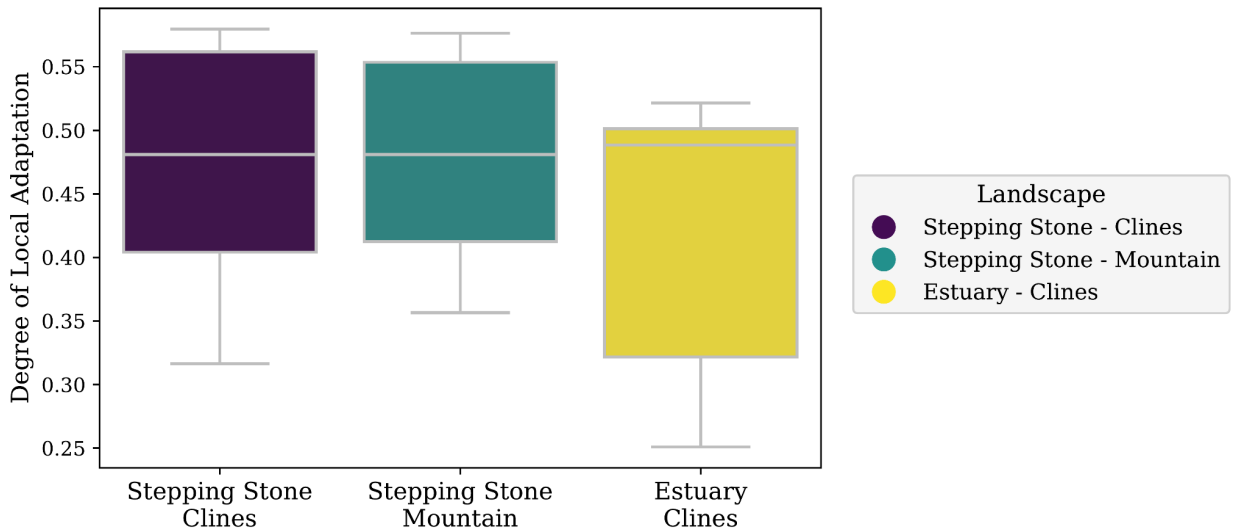


583

584 **Fig S16** The relationship between the degree of local adaptation ( $LA_{\Delta SA}$ ), levels of  
 585 *IBE* within marker sets, and median performance of models trained with one of the  
 586 three marker sets: (A) *adaptive*, (B) *all*, and (c) *neutral* marker sets. *IBE* is  
 587 quantified as Spearman's rank correlation between population pairwise  $F_{ST}$  and  
 588 Euclidean distance of adaptive environments. Data included in these figures are  
 589 from 1- and 2-trait simulations. Code to create these figures can be found in SC  
 590 02.02.10.



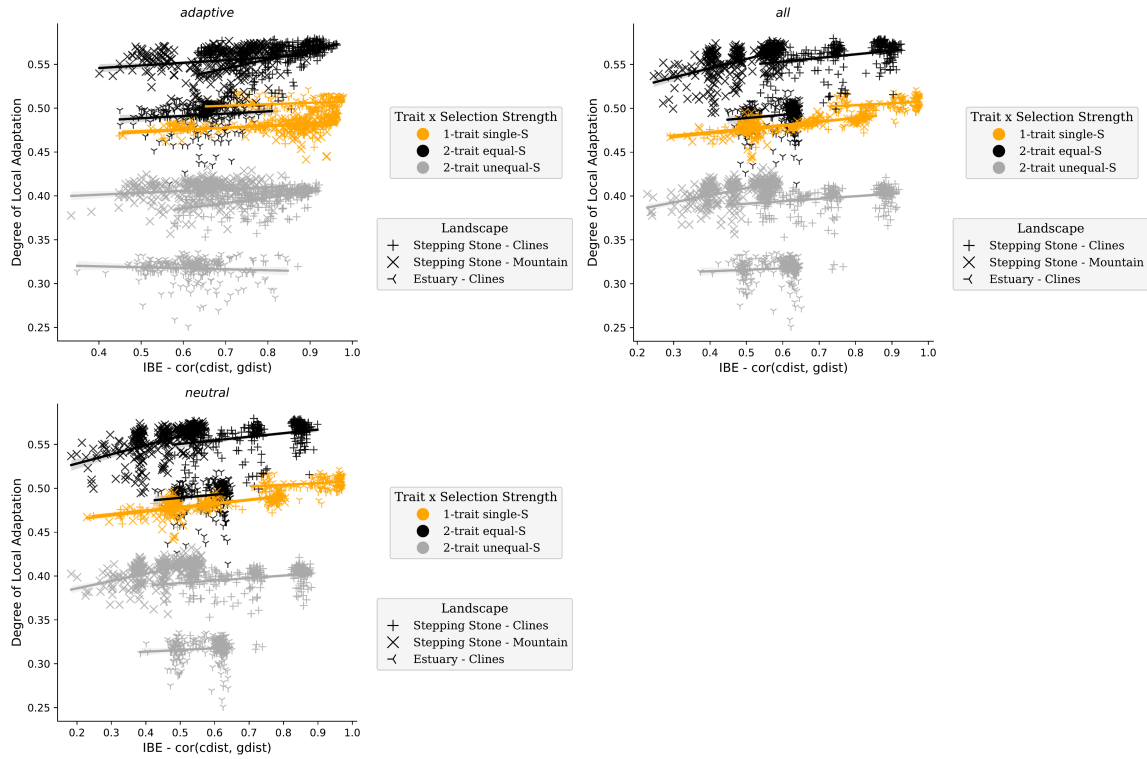
591



592

593 **Fig S17** Levels of isolation-by-environment in marker sets vary across landscapes  
 594 (A) and the degree of local adaptation reached by metapopulations on these  
 595 landscapes (B). The pattern in (A) given (B) is in contrast to patterns between levels  
 596 of IBE and the degree of local adaptation (Fig. S29). *IBE* is quantified as Spearman's  
 597 rank correlation between population pairwise  $F_{ST}$  (gdist) and Euclidean distance of  
 598 adaptive environments (cdist). Data in this figure is from all 1- and 2-trait  
 599 simulations. Code to create this figure can be found in SC 02.02.10.





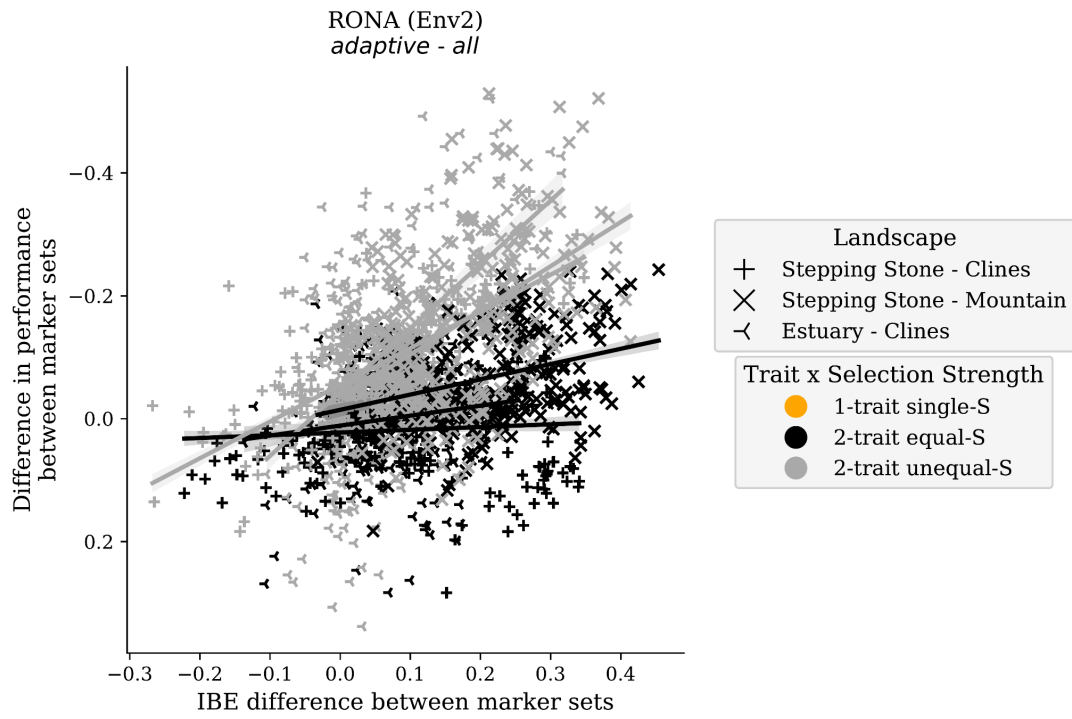
600

601

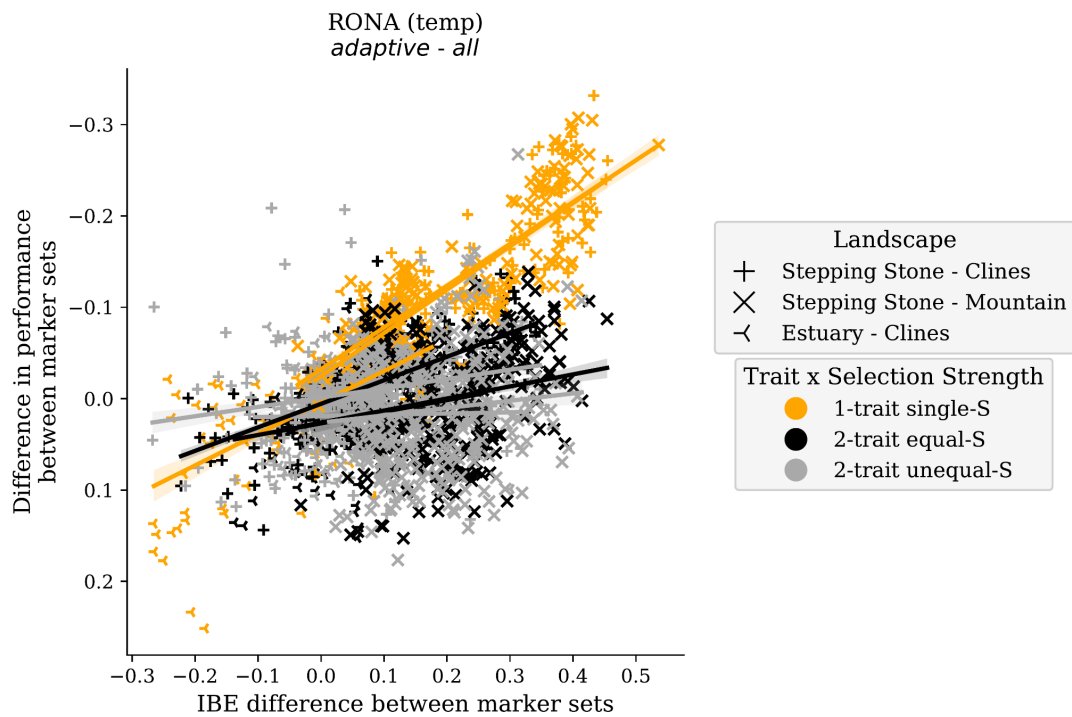
602 **Fig S18** The levels of isolation-by-distance in marker sets (panels) are weakly  
 603 correlated with the degree of local adaptation ( $LA_{ASA}$ ) within simulation levels.  $IBE$   
 604 is quantified as Spearman's rank correlation between population pairwise  $F_{ST}$   
 605 ( $gdist$ ) and Euclidean distance of adaptive environments ( $cdist$ ). Data included in  
 606 this figure is from all marker sets from 1- and 2-trait simulations. Code to create  
 607 this figure can be found in 02.02.10.

608 (Fig

S19)



609

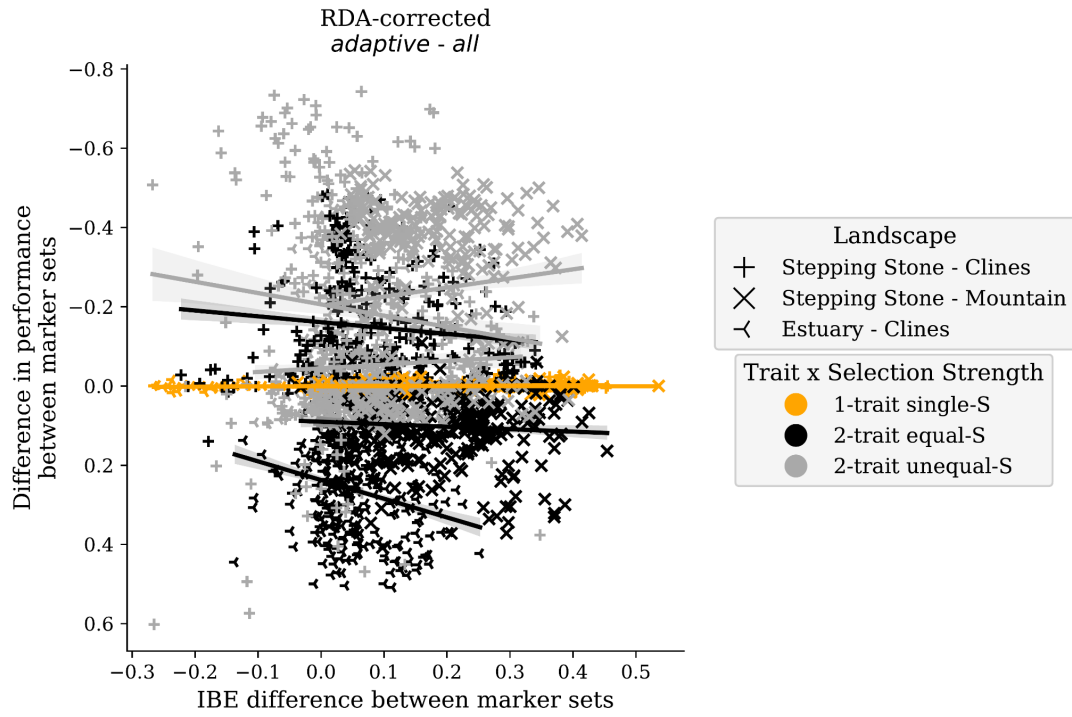


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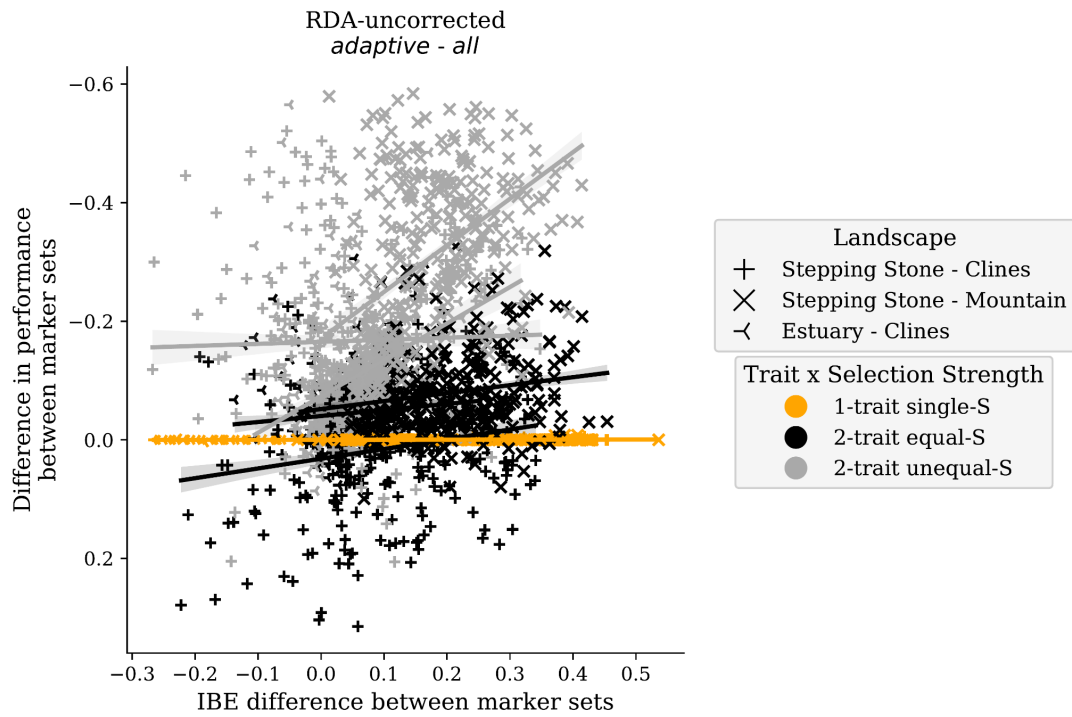
611 (Fig

S19

continued)



612



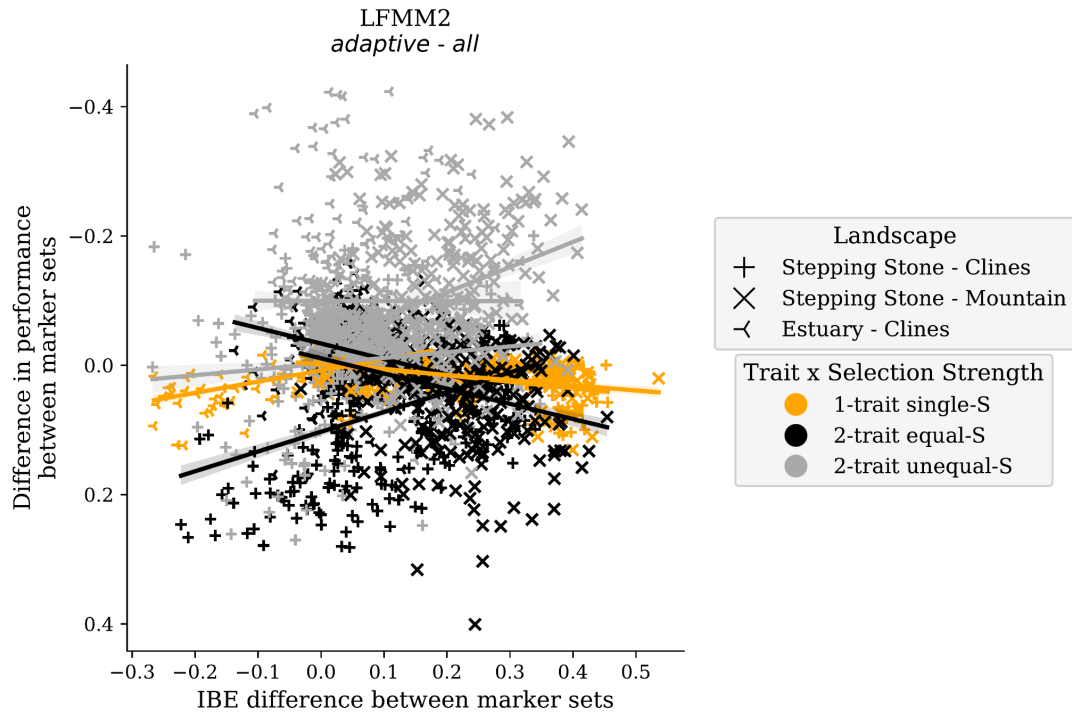
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614

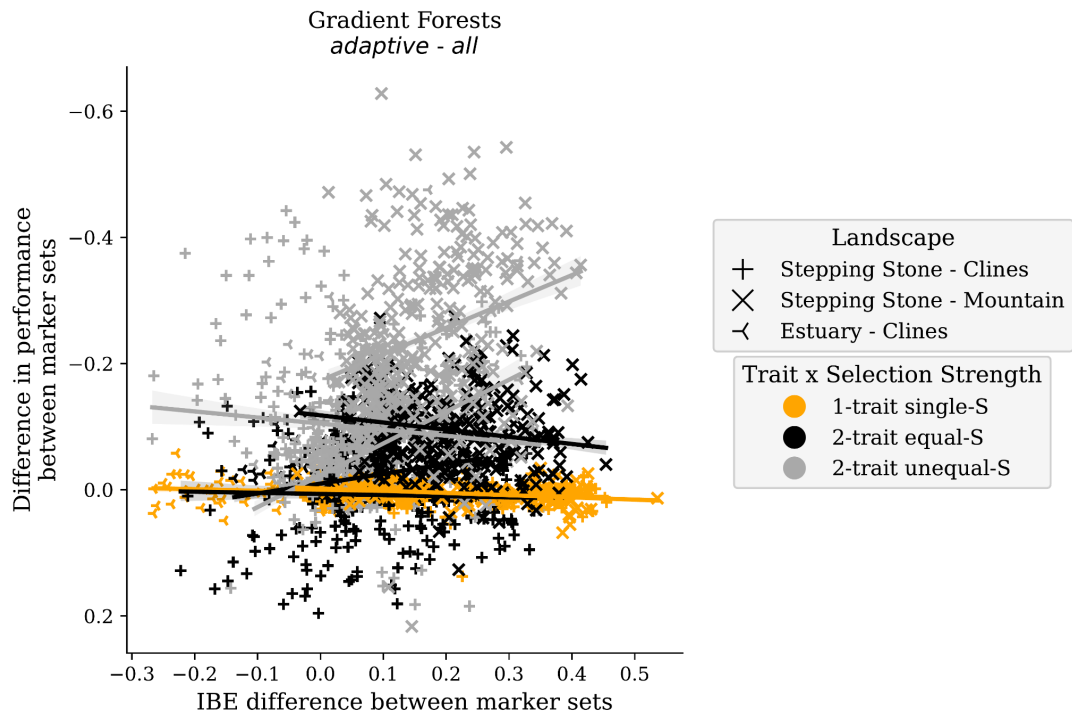
615 (Fig

S19

continued)



616



617

618 **Fig S19** Differences in levels of IBE between marker sets used to train models is  
619 generally unrelated to differences in model performances. Shown is the difference  
620 in median performance between *adaptive* and *all* marker sets and the difference in  
621 *IBE* between these marker sets. *IBE* is quantified as Spearman's rank correlation  
622 between population pairwise  $F_{ST}$  and Euclidean distance of adaptive environments.  
623 Data in this figure is from 1- and 2-trait simulations. Code to create these figures  
624 can be found in SC 02.02.12.

625

	<b>1</b>	<b>2</b>	<b>3</b>	<b>4</b>	<b>5</b>	<b>6</b>	<b>7</b>	<b>8</b>	<b>9</b>	<b>10</b>	
Latitude	<b>10</b>	91.0	92.0	93.0	94.0	95.0	96.0	97.0	98.0	99.0	100.0
	<b>9</b>	81.0	82.0	83.0	84.0	85.0	86.0	87.0	88.0	89.0	90.0
	<b>8</b>	71.0	72.0	73.0	74.0	75.0	76.0	77.0	78.0	79.0	80.0
	<b>7</b>	61.0	62.0	63.0	64.0	65.0	66.0	67.0	68.0	69.0	70.0
	<b>6</b>	51.0	52.0	53.0	54.0	55.0	56.0	57.0	58.0	59.0	60.0
	<b>5</b>	41.0	42.0	43.0	44.0	45.0	46.0	47.0	48.0	49.0	50.0
	<b>4</b>	31.0	32.0	33.0	34.0	35.0	36.0	37.0	38.0	39.0	40.0
	<b>3</b>	21.0	22.0	23.0	24.0	25.0	26.0	27.0	28.0	29.0	30.0
	<b>2</b>	11.0	12.0	13.0	14.0	15.0	16.0	17.0	18.0	19.0	20.0
	<b>1</b>	1.0	2.0	3.0	4.0	5.0	6.0	7.0	8.0	9.0	10.0

626

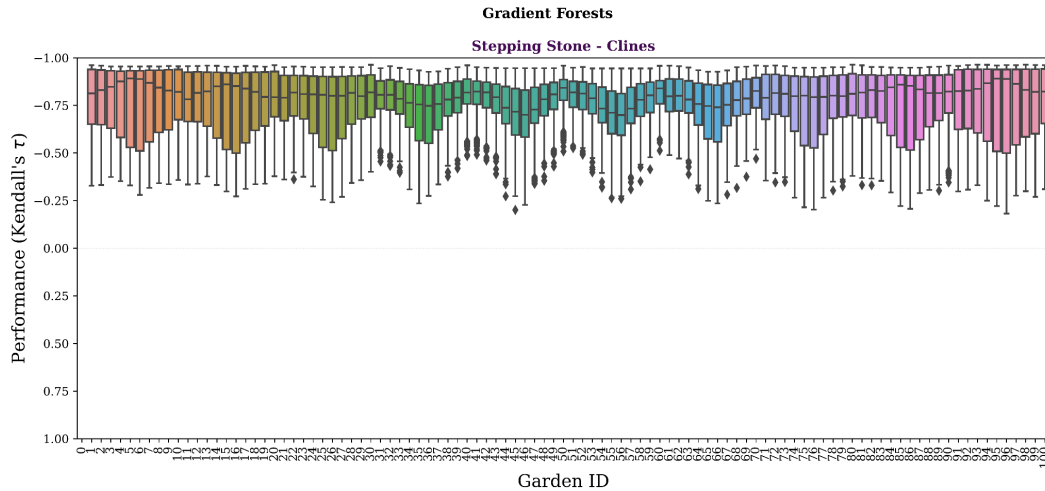
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Longitude

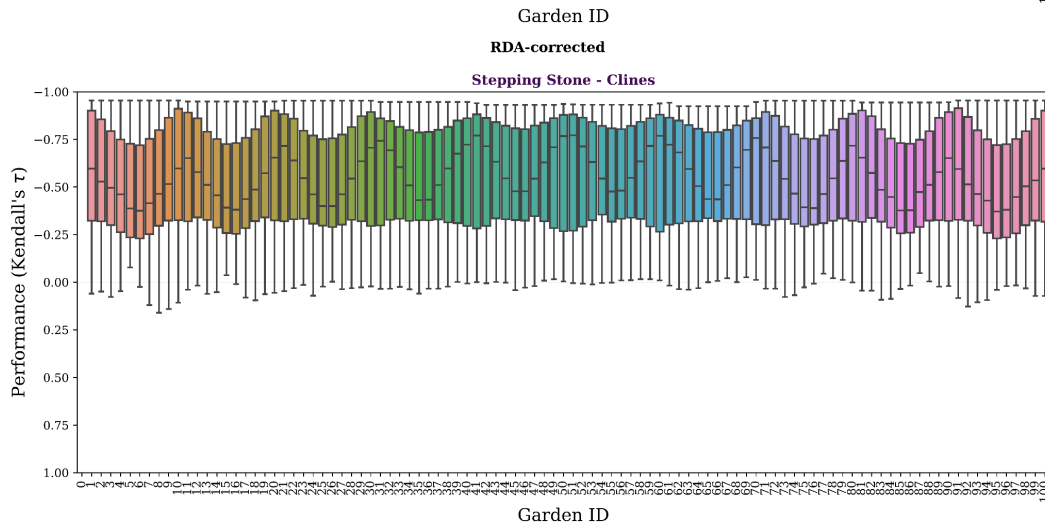
628

629 **Fig S20** A map of Garden ID (unbolded entries) across each landscape for 1-, 2- and  
 630 6-trait simulations (latitudinal and longitudinal grids are bolded). This map can be  
 631 used to interpret the ordering of gardens along x-axes of Figs. S21 S22 and S23. Code  
 632 used to create this figure can be found in SC 02.02.04.

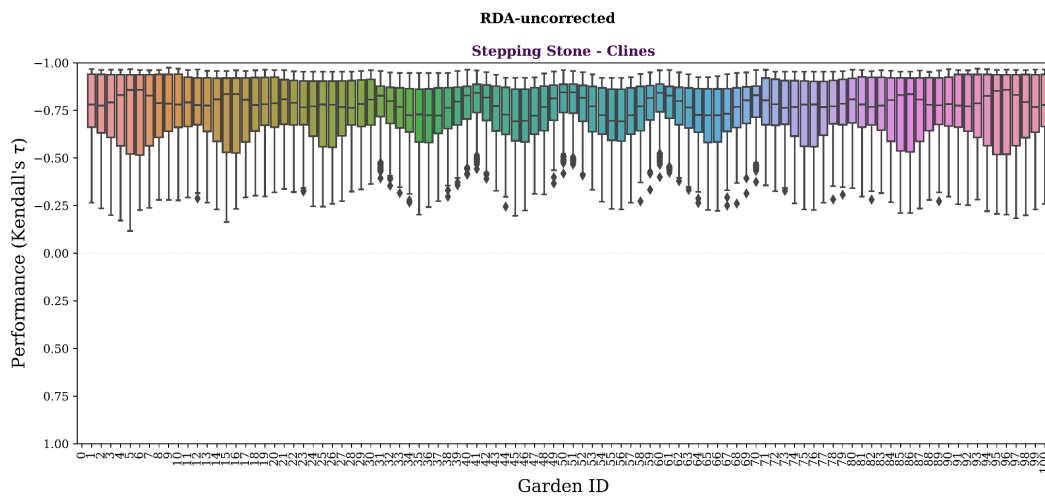
633 (Fig. S21)



634



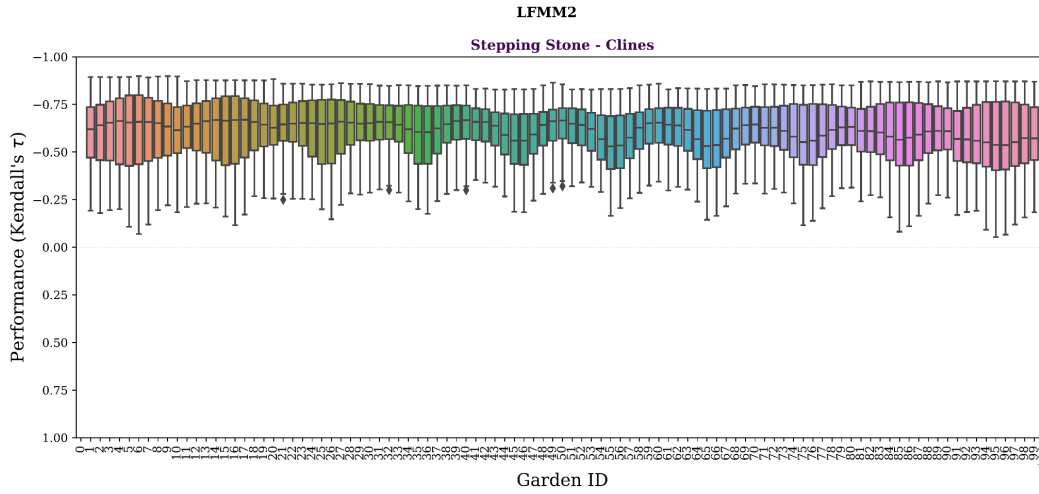
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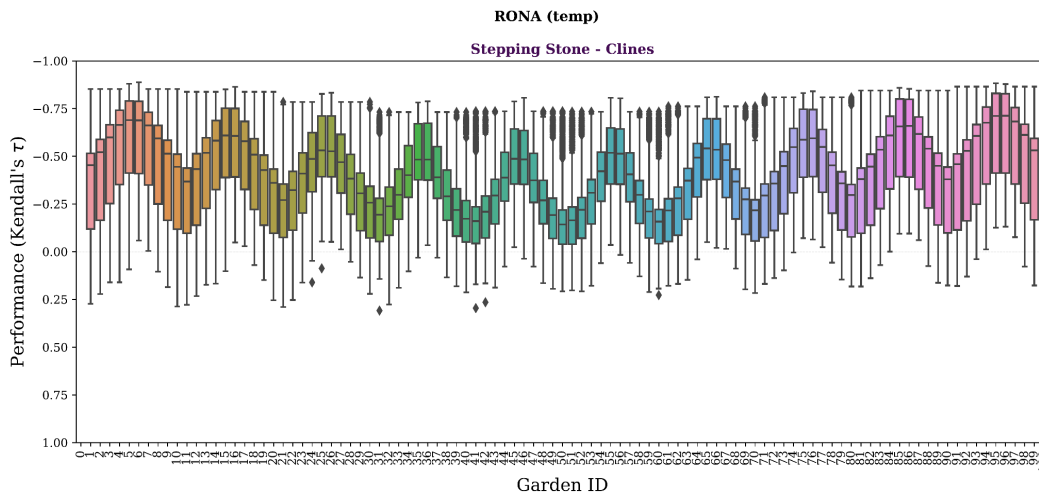
636

637 (Fig S21 continued)

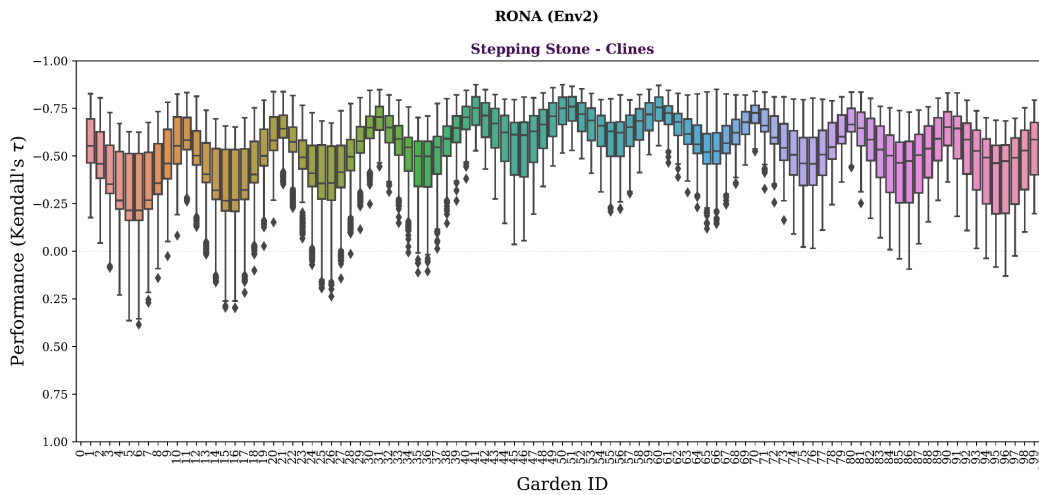
Supplement - *Lind, Lotterhos, and the limits of genomic offsets*



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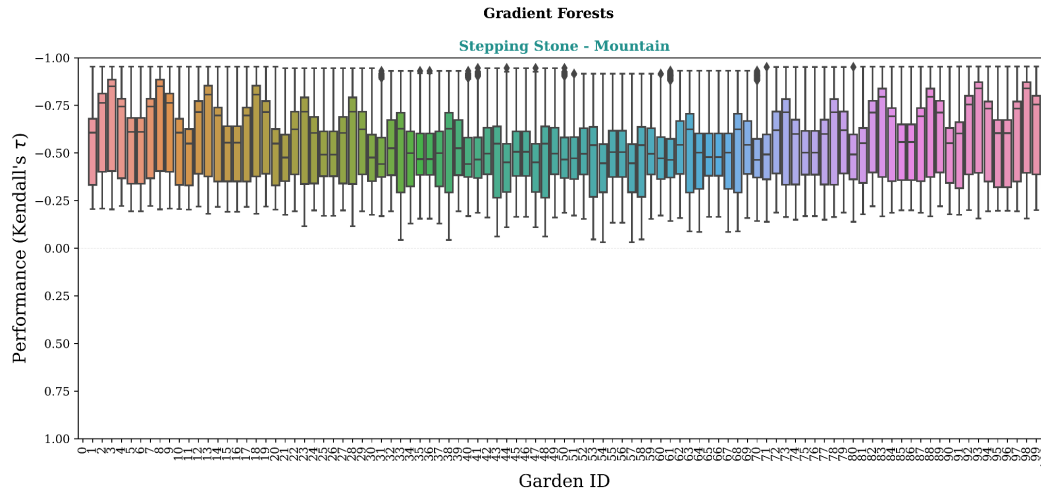


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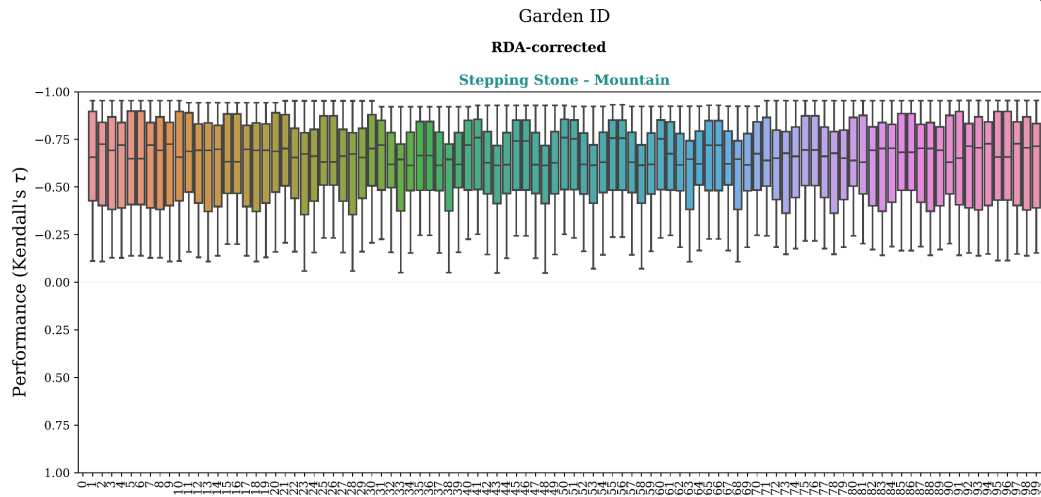


641 **Fig S21** Genomic offset methods have variable performance across the *Stepping-*  
642 *Stone Clines* landscape. Shown is the variability of each offset method performance  
643 (y-axes) across the 100 common gardens (x-axes). Gardens are ordered from left to  
644 right by garden ID. This ordering of gardens is equivalent to the southwestern-most  
645 garden first and northeastern-most garden last (see Fig. S20 for a map of garden ID  
646 across each landscape). Similar figures for *Stepping-Stone Mountain* and *Estuary-*  
647 *Clines* landscapes can be found in Fig S22 and Fig S23, respectively. Data included  
648 in this figure is from evaluation of 1- and 2-trait simulations using *all* markers. Code  
649 used to create these figures can be found in SC 02.02.04.

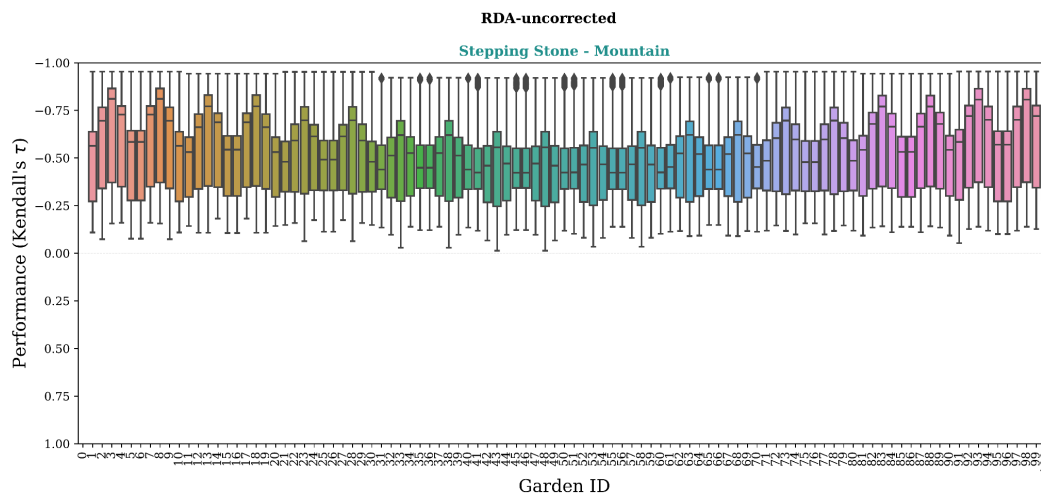
650 (Fig S22)



651



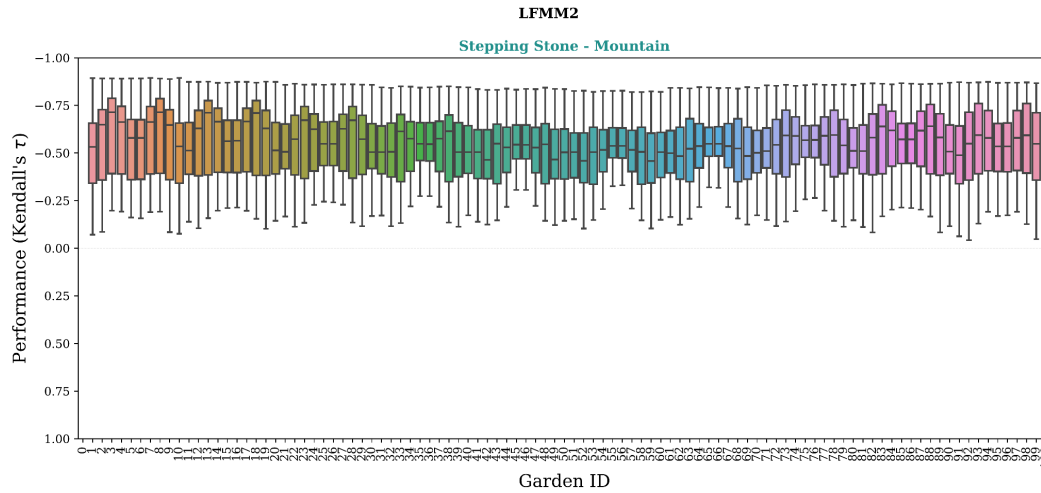
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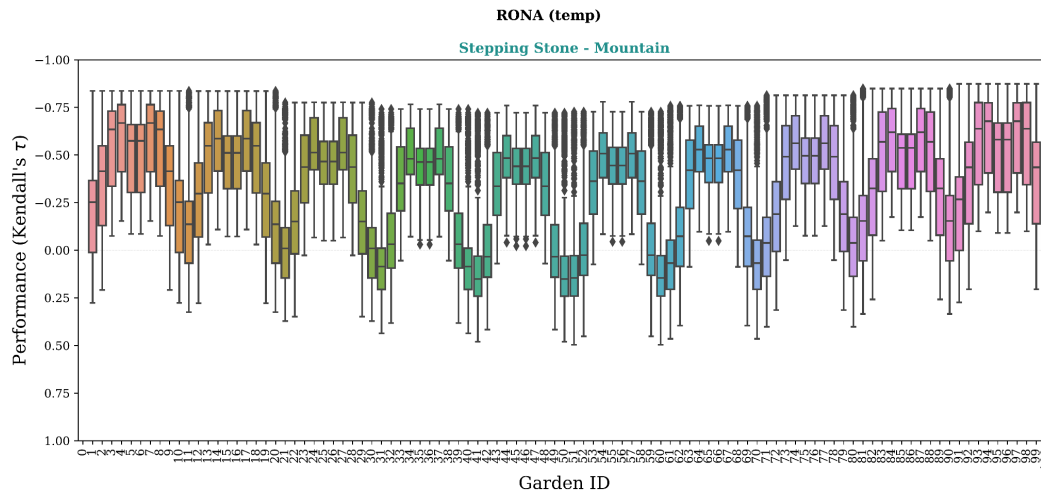
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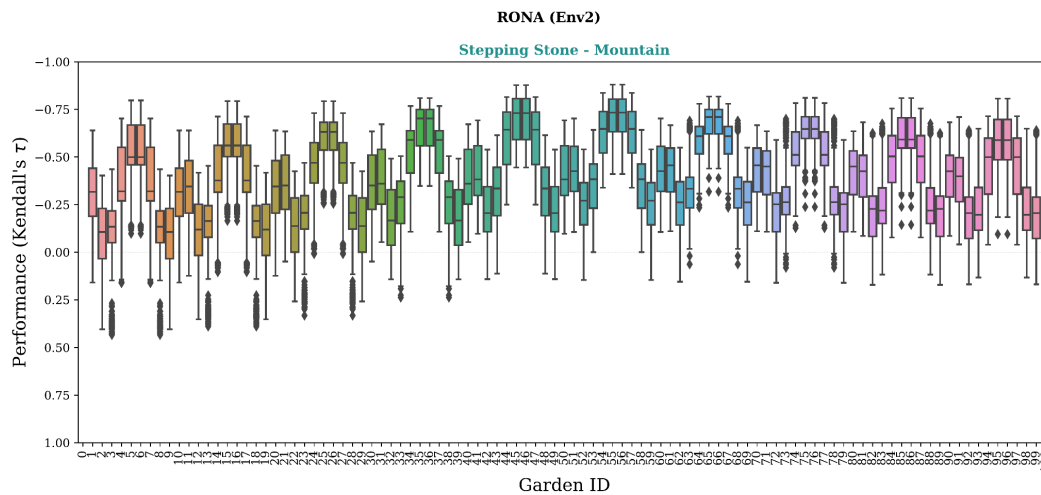
655 (Fig S22 continued)



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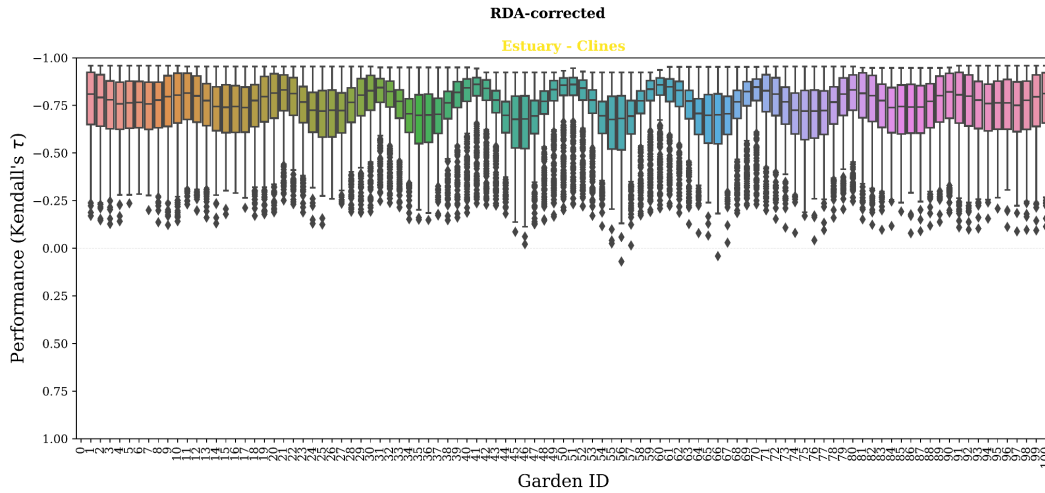
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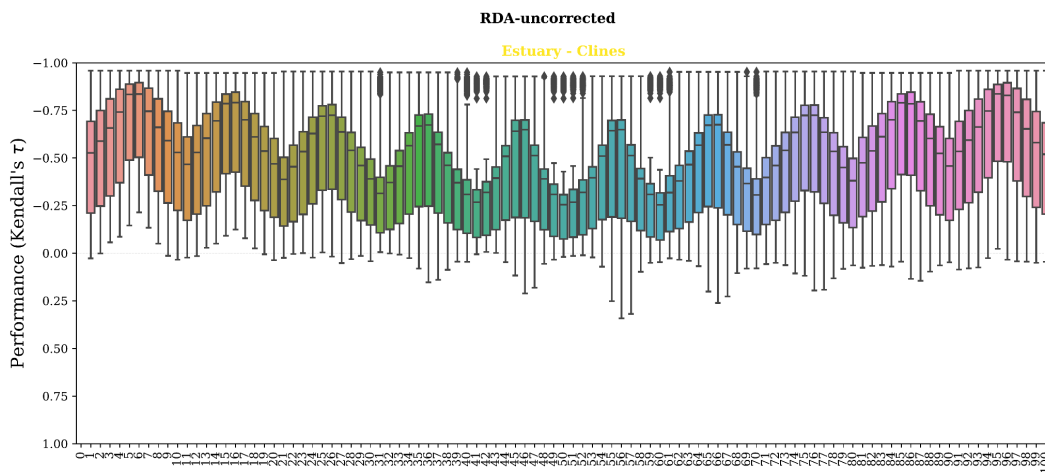
658

659 **Fig S22** Genomic offset methods have variable performance across the *Stepping-*  
660 *Stone - Mountain* landscape. Shown is the variability of each offset method  
661 performance (y-axes) across the 100 common gardens (x-axes). Gardens are  
662 ordered from left to right by garden ID. This ordering of gardens is equivalent to  
663 the southwestern-most garden first and northeastern-most garden last (see Fig. S20  
664 for a map of garden ID across each landscape). Similar figures for *Stepping-Stone -*  
665 *Clines* and *Estuary - Clines* landscapes can be found in Fig S21 and Fig S23,  
666 respectively. Data included in this figure is from evaluation of 1- and 2-trait  
667 simulations using *all* markers. Code used to create this figure can be found in SC  
668 02.02.04.

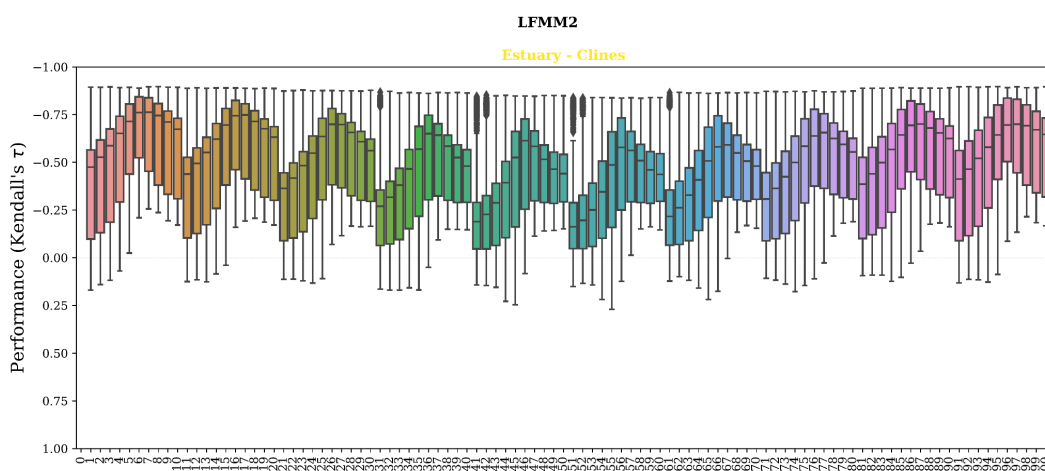
669 (Fig S23)



670

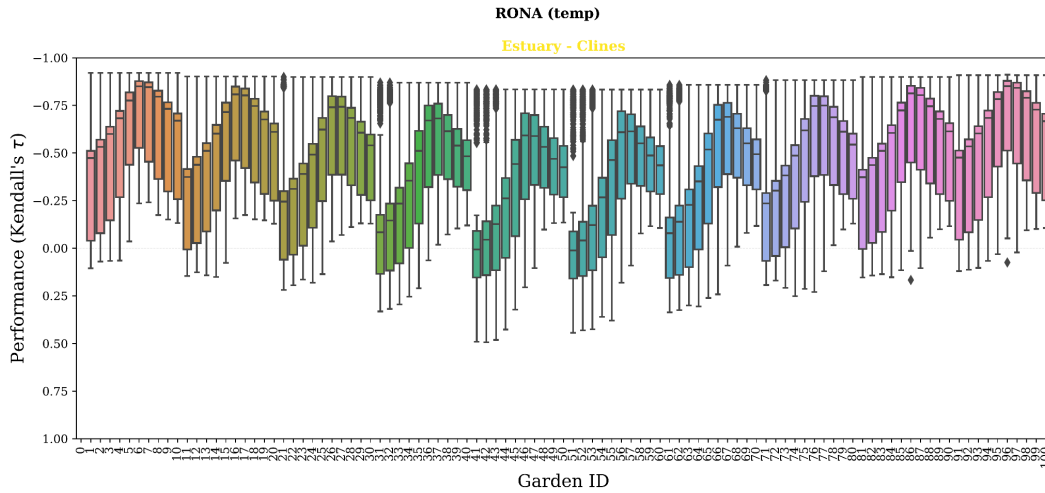


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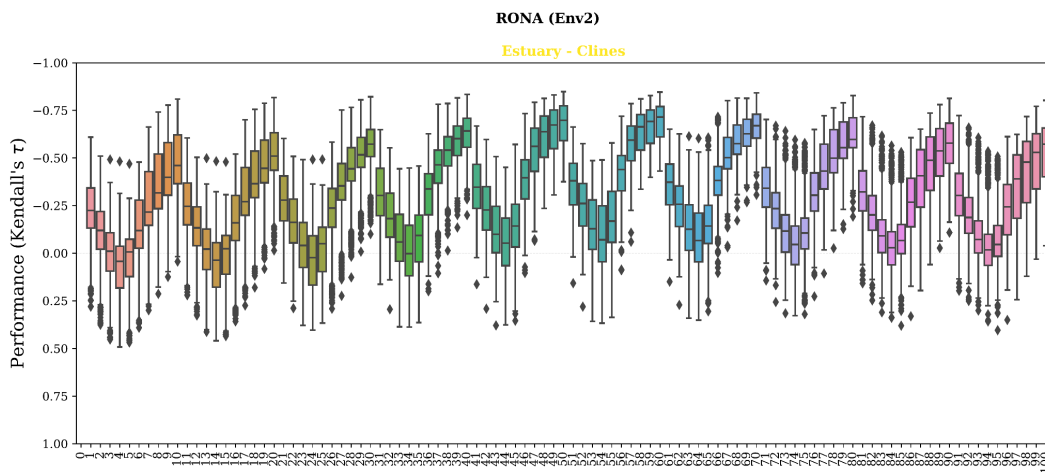


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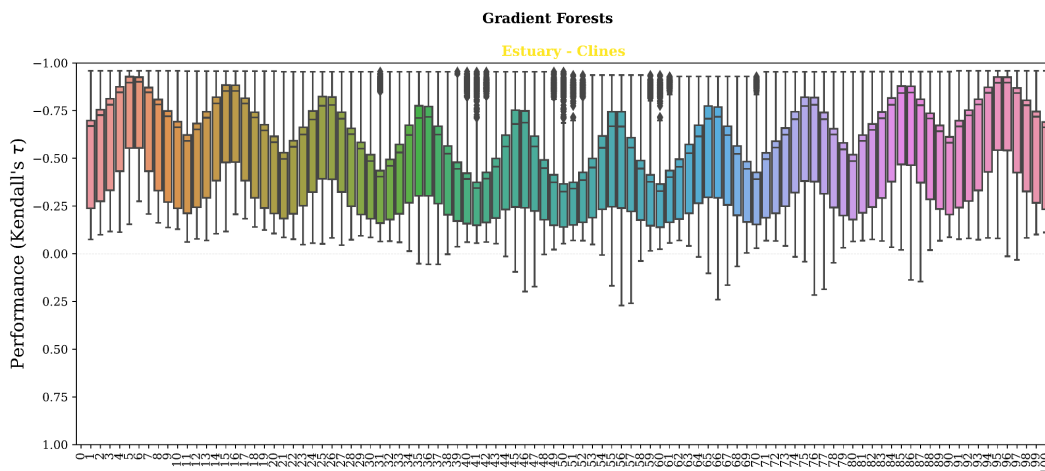
673 (Fig S23 continued)



674



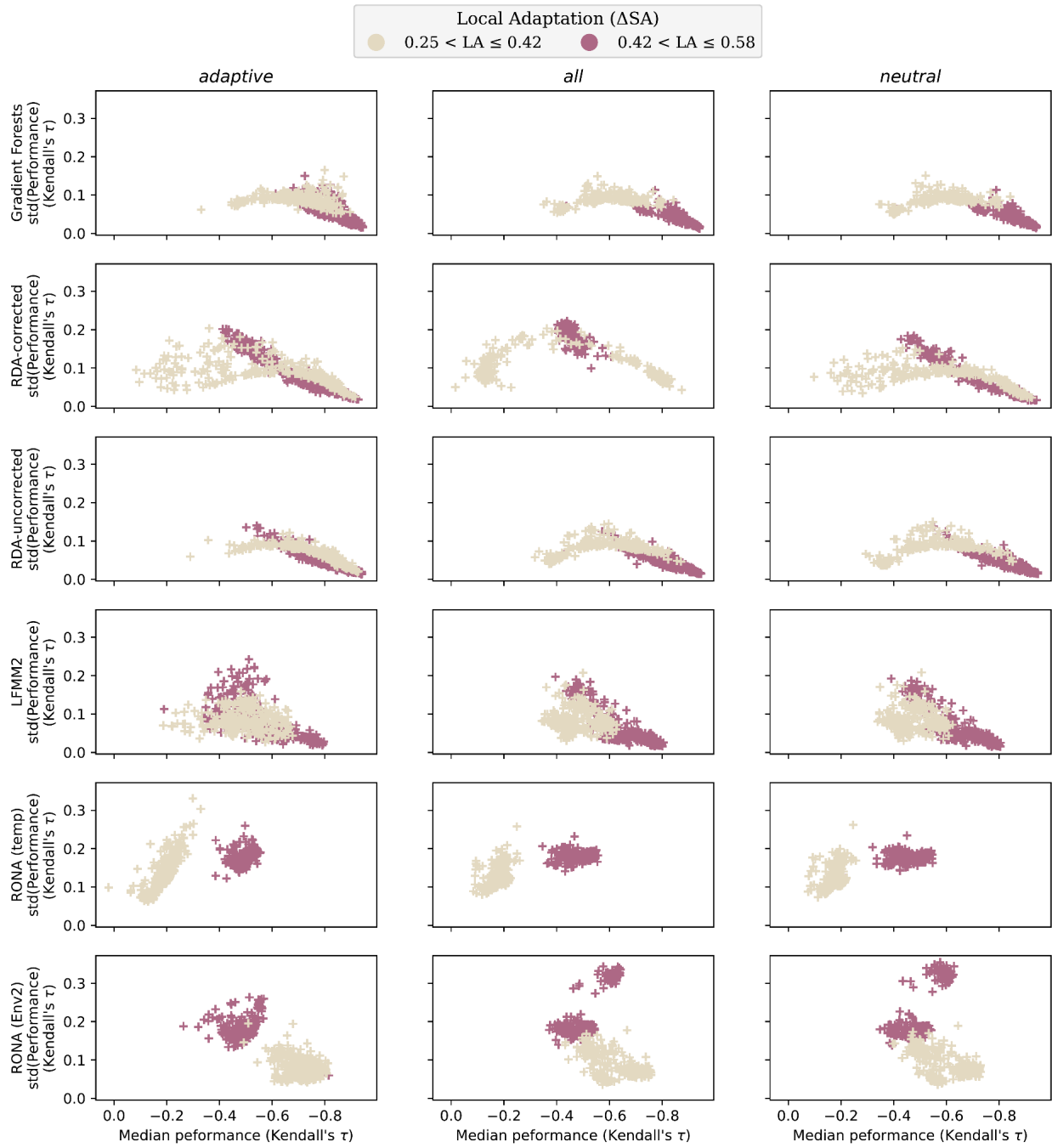
675



676

677 **Fig S23** Genomic offset methods have variable performance across the *Estuary -*  
678 *Clines* landscape. Shown is the variability of each offset method performance (y-  
679 axes) across the 100 common gardens (x-axes). Gardens are ordered from left to  
680 right by garden ID. This ordering of gardens is equivalent to the southwestern-most  
681 garden first and northeastern-most garden last (see Fig. S20 for a map of garden ID  
682 across each landscape). Similar figures for *Stepping-Stone - Clines* and *Stepping-*  
683 *Stone - Mountain* landscapes can be found in Fig S21 and Fig S22, respectively. Data  
684 included in this figure is from evaluation of 1- and 2-trait simulations using *all*  
685 markers. Code used to create this figure can be found in SC 02.02.04.

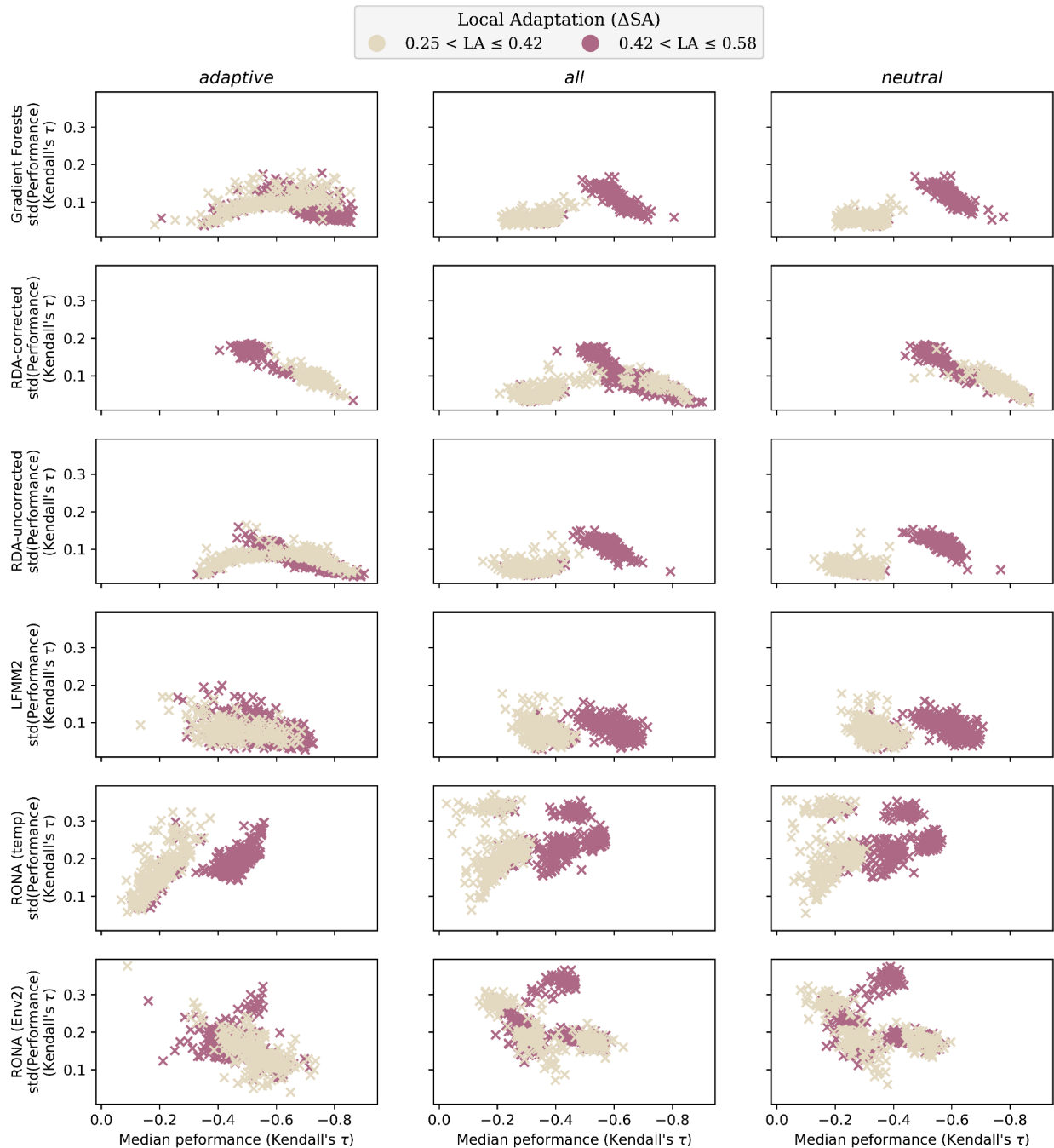
686 (Fig. S36)



687

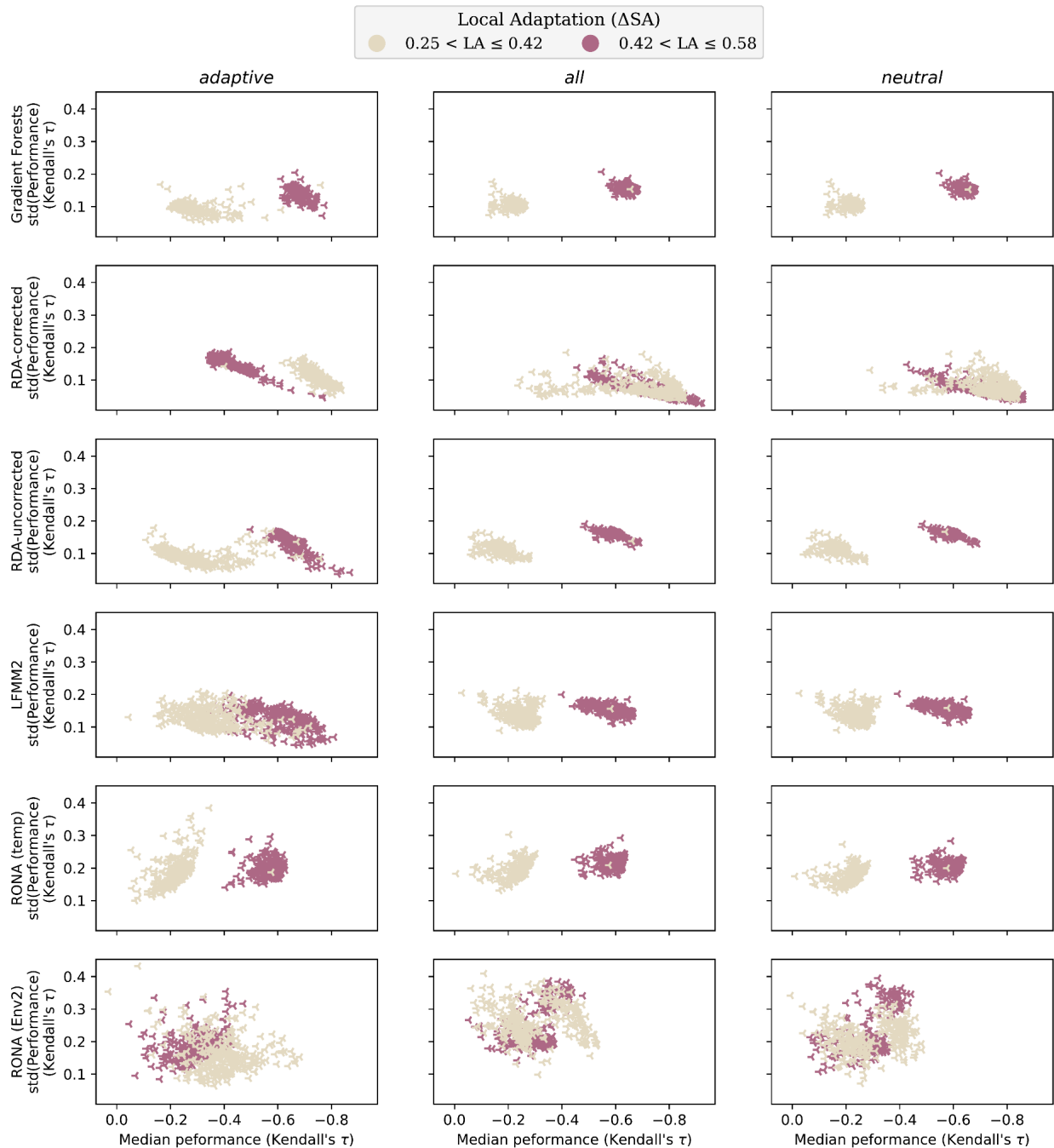


688 **Fig S24** Variability of genomic offset performance (y-axes) for a given model (+)  
689 often decreases with increasing median performance (x-axes). Shown are patterns  
690 from each offset method (rows) for each marker set (columns) used in training.  
691 Data included in this figure is from evaluation of 2-trait simulations from *Stepping-*  
692 *Stone - Clines* landscapes processed through the *Adaptive Environment* workflow.  
693 For similar figures for *Stepping-Stone - Mountain* and *Estuary - Clines* landscapes,  
694 see Figs. S25-S26, respectively. Code used to create these figures can be found in SC  
695 02.02.07.



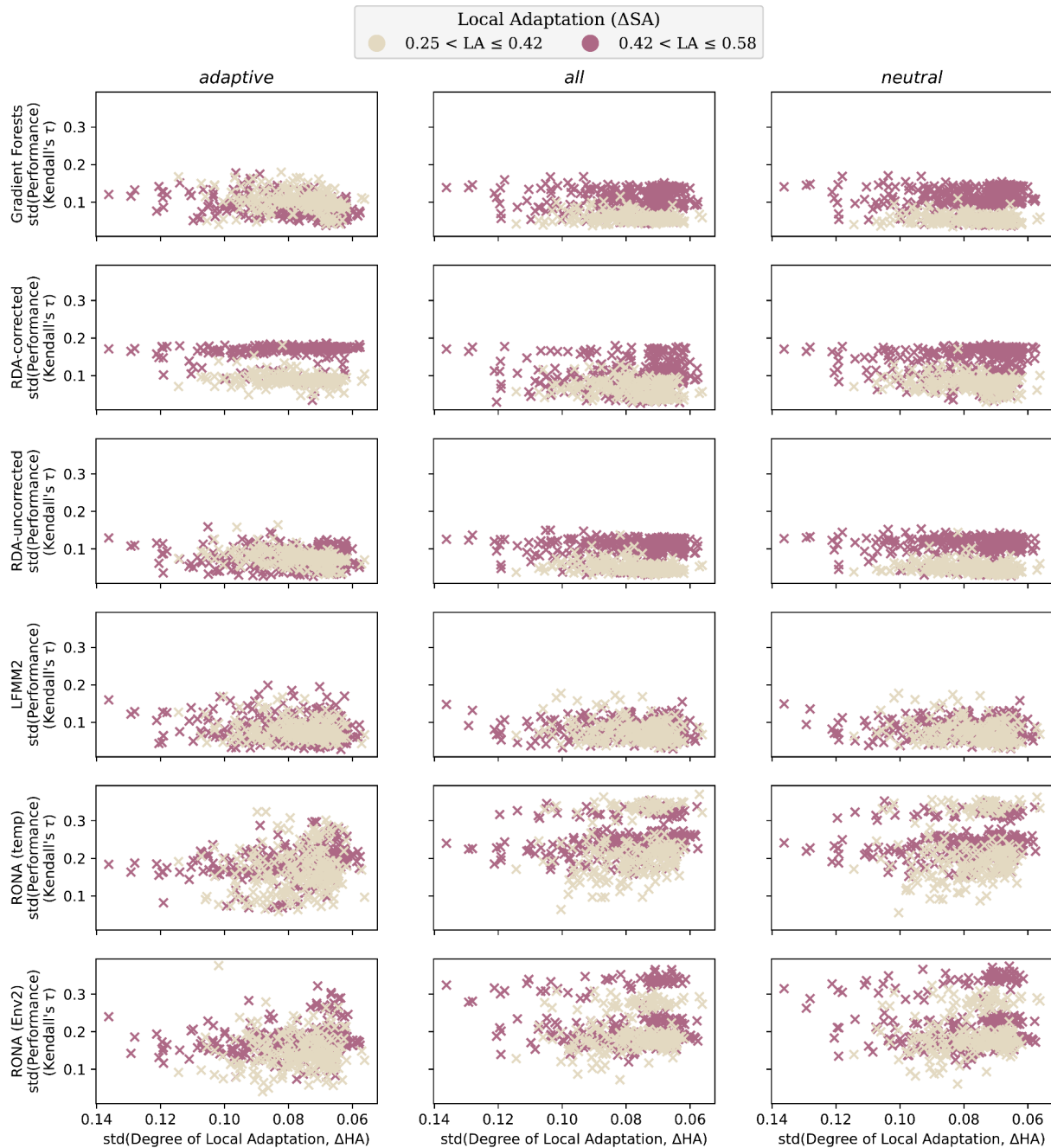
696

697 **Fig S25** Variability across evaluations of genomic offsets often decreases with  
 698 increasing average performance across marker sets. Data included in this figure is  
 699 from evaluation of 2-trait simulations from *Stepping-Stone - Mountain* landscapes.  
 700 For similar figures for *Stepping-Stone - Clines* and *Estuary - Clines* landscapes, see  
 701 Figs. 24 and S26, respectively. Code used to create these figures can be found in SC  
 702 02.02.07.



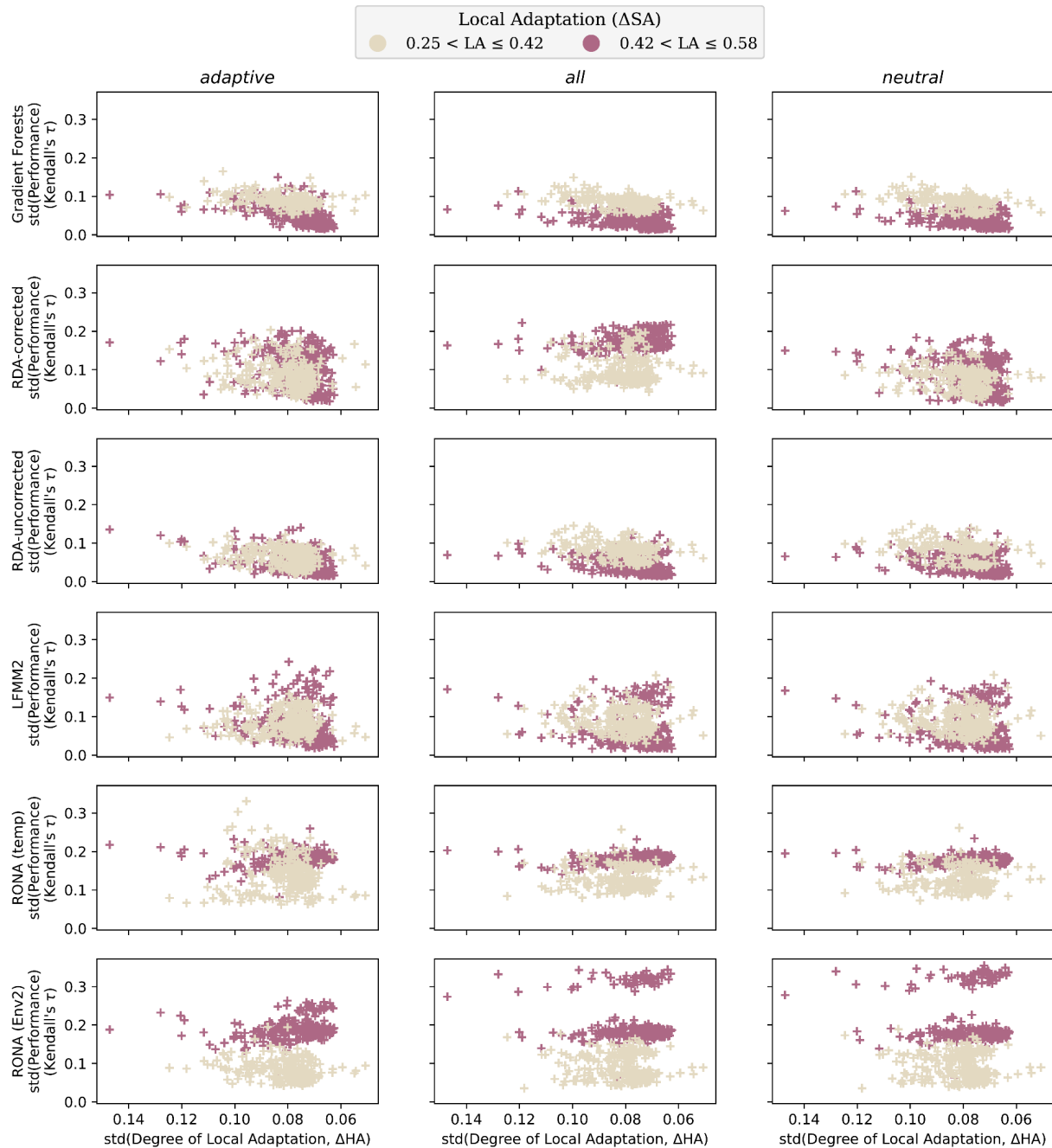
703

704 **Fig S26** Variability across evaluations of genomic offsets often decreases with  
 705 increasing average performance across marker sets. Data included in this figure is  
 706 from evaluation of 2-trait simulations from *Estuary - Clines* landscapes. For similar  
 707 figures for *Stepping-Stone - Clines* and *Stepping-Stone - Mountain* landscapes, see  
 708 Figs. S24 and S25, respectively. Code used to create these figures can be found in  
 709 SC 02.02.07.



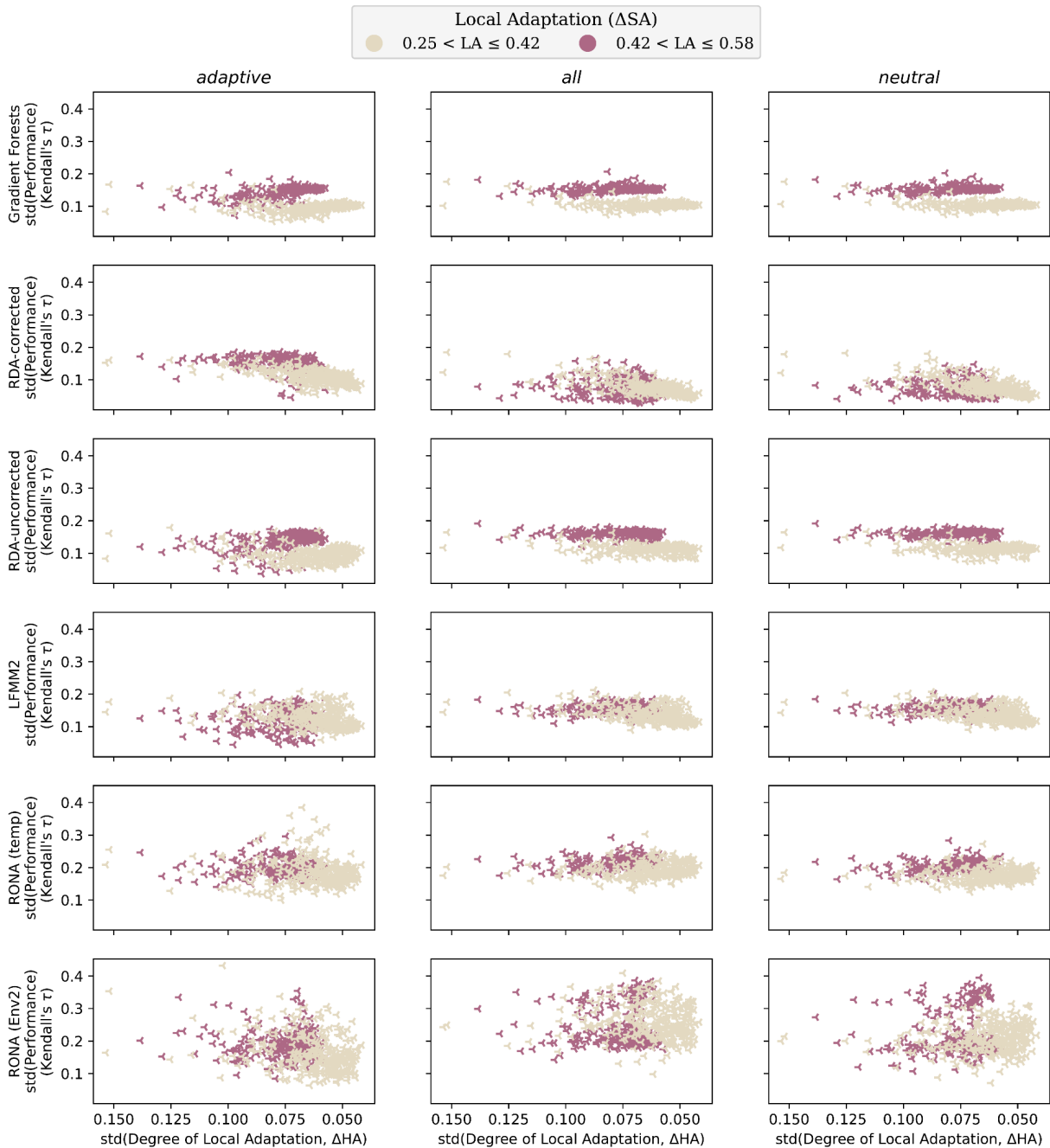
710

711 **Fig S27** Variability across evaluations of genomic offsets is often unrelated to the  
 712 variability in the degree of local adaptation across populations. Data included in  
 713 this figure is from evaluation of 2-trait simulations from *Stepping-Stone - Mountain*  
 714 landscapes. For similar figures for *Stepping-Stone - Clines* and *Estuary - Clines*  
 715 landscapes, see Figs. S27 and S28, respectively. Code used to create these figures  
 716 can be found in SC 02.02.07.



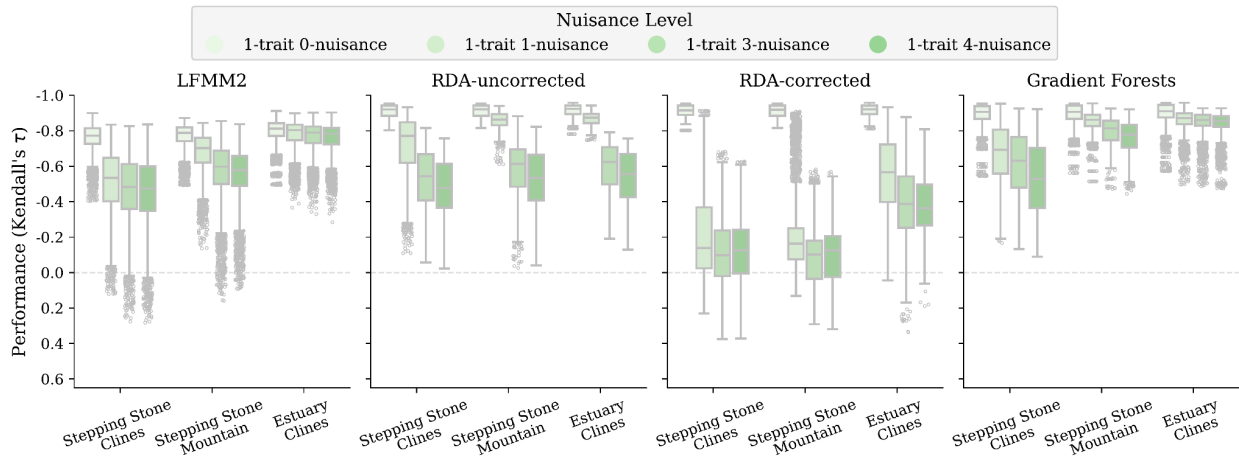
717

718 **Fig S28** Variability across evaluations of genomic offsets is often unrelated to the  
 719 variability in the degree of local adaptation across populations. Data included in  
 720 this figure is from evaluation of 2-trait simulations from *Stepping Stone - Clines*  
 721 landscapes. For similar figures for *Estuary - Clines* and *Stepping-Stone - Mountain*  
 722 landscapes, see Figs. S26 and S28, respectively. Code used to create these figures  
 723 can be found in SC 02.02.07.

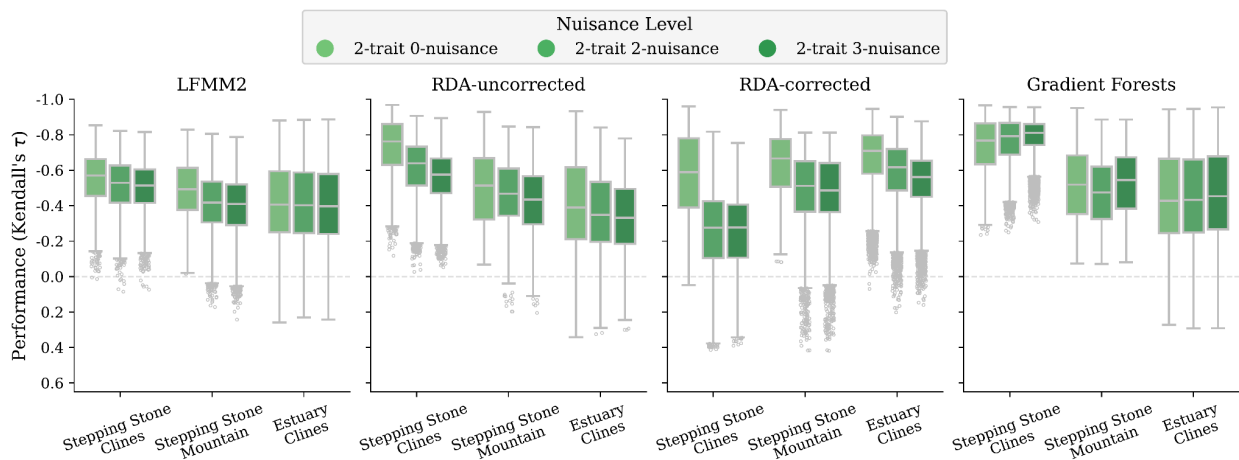


724

725 **Fig S29** Variability across evaluations of genomic offsets is often unrelated to the  
 726 variability in the degree of local adaptation across populations. Data included in  
 727 this figure is from evaluation of 2-trait simulations from *Estuary - Clines*  
 728 landscapes. For similar figures for *Stepping-Stone - Clines* and *Stepping-Stone -*  
 729 *Mountain* landscapes, see Figs. S26 and S27, respectively. Code used to create these  
 730 figures can be found in SC 02.02.07.



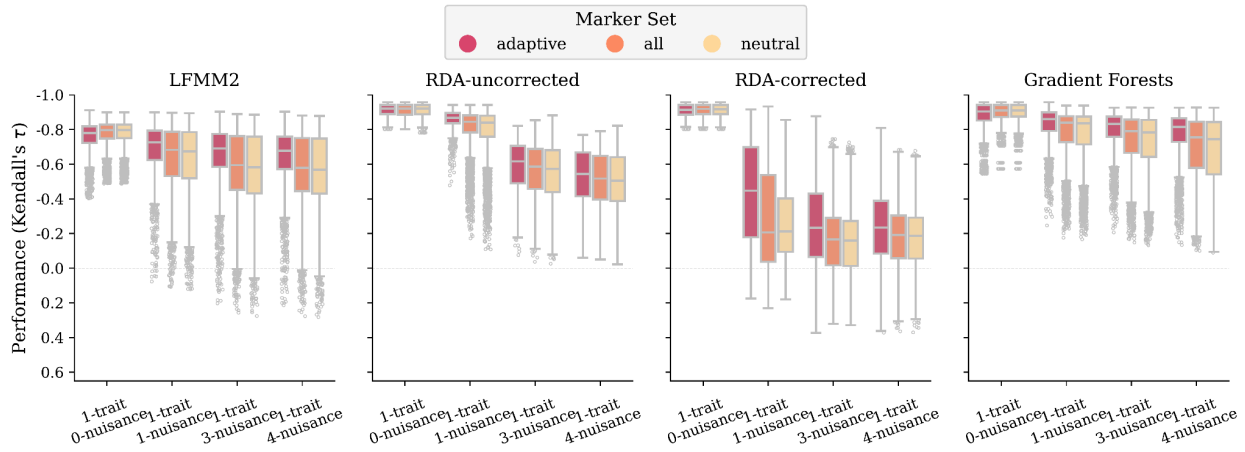
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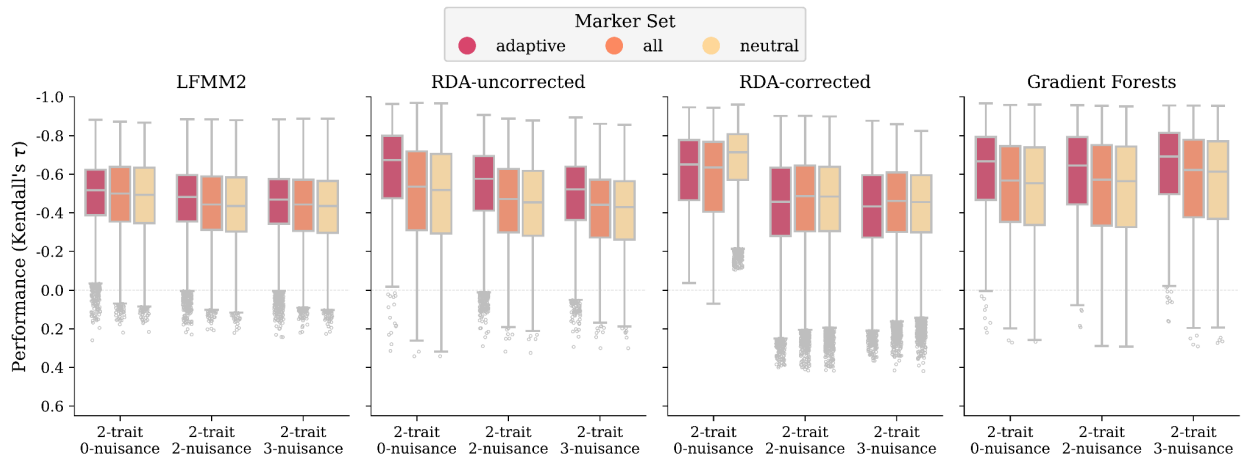
732

733 **Fig S30** Effect of non-adaptive nuisance environmental variables on offset  
 734 performance faceted by landscape. Shown are offsets from 1- and 2-trait  
 735 simulations trained using only adaptive environments (0-nuisance) or with  
 736 adaptive environments and the addition of  $N > 0$  non-adaptive environmental  
 737 variables ( $N$ -nuisance). RONA is not shown because it is univariate with respect to  
 738 environmental variables. The nuisance variables for 1-trait simulations are: Env2,  
 739 ISO, TSsd, PSsd; and for 2-trait simulations are ISO, TSsd, PSsd; see Table 2. The  
 740 *Nuisance Environment* workflow was used to produce this data. Code to create  
 741 these figures can be found in SC 02.02.06.

742



743



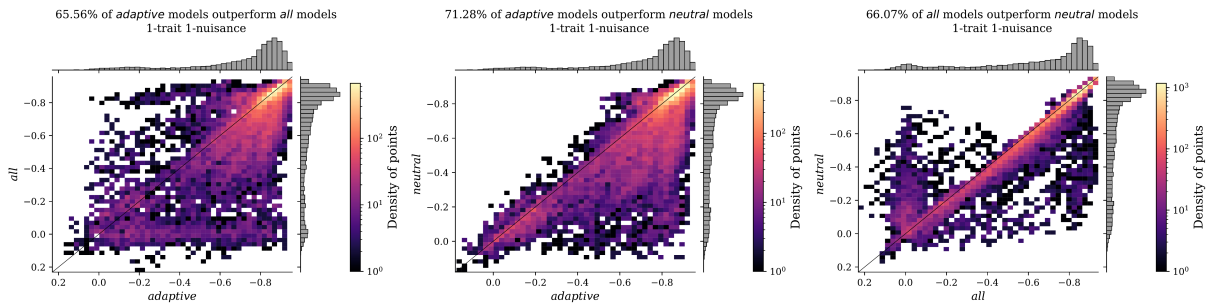
744 **Fig S31** Effect of non-adaptive nuisance environmental variables on offset  
 745 performance faceted by marker set. Shown are offsets from 1- (A) and 2-trait (B)  
 746 simulations trained using only adaptive environments (0-nuisance) or with  
 747 adaptive environments and the addition of  $N > 0$  non-adaptive environmental  
 748 variables ( $N$ -nuisance). RONA is not shown because it is univariate with respect to  
 749 environmental variables. The nuisance variables for 1-trait simulations are: Env2,  
 750 ISO, TSsd, PSsd; and for 2-trait simulations are ISO, TSsd, PSsd; see Table 2. Code to  
 751 create these figures can be found in SC 02.02.06.



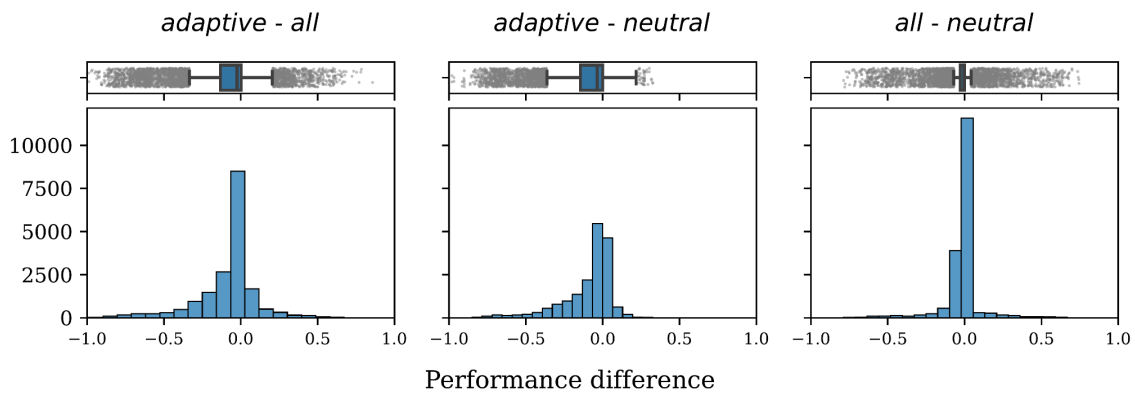
752 (Fig. S31)

753

**1-trait 1-nuisance**



754

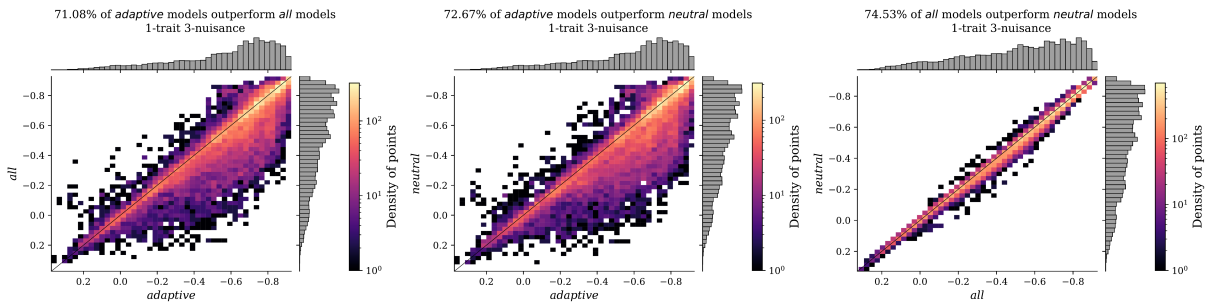


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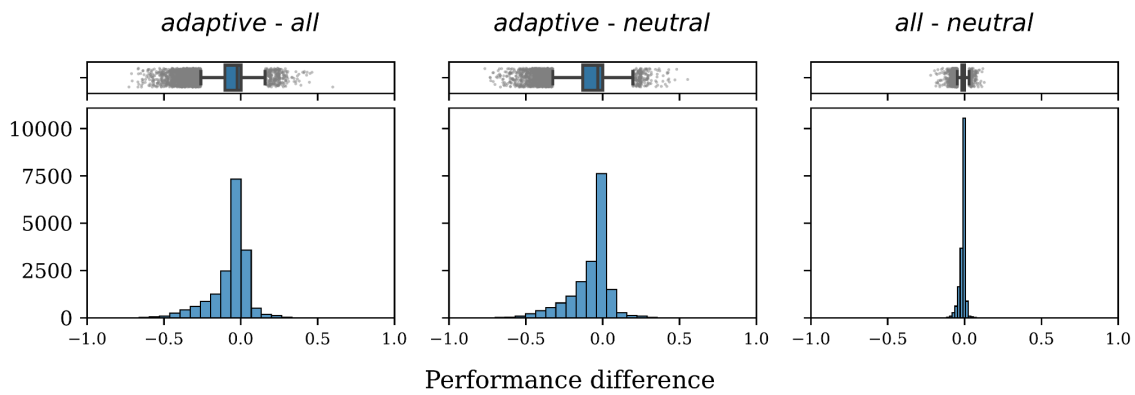
756 (Fig. S31 continued)

757

**1-trait 3- nuisance**



758

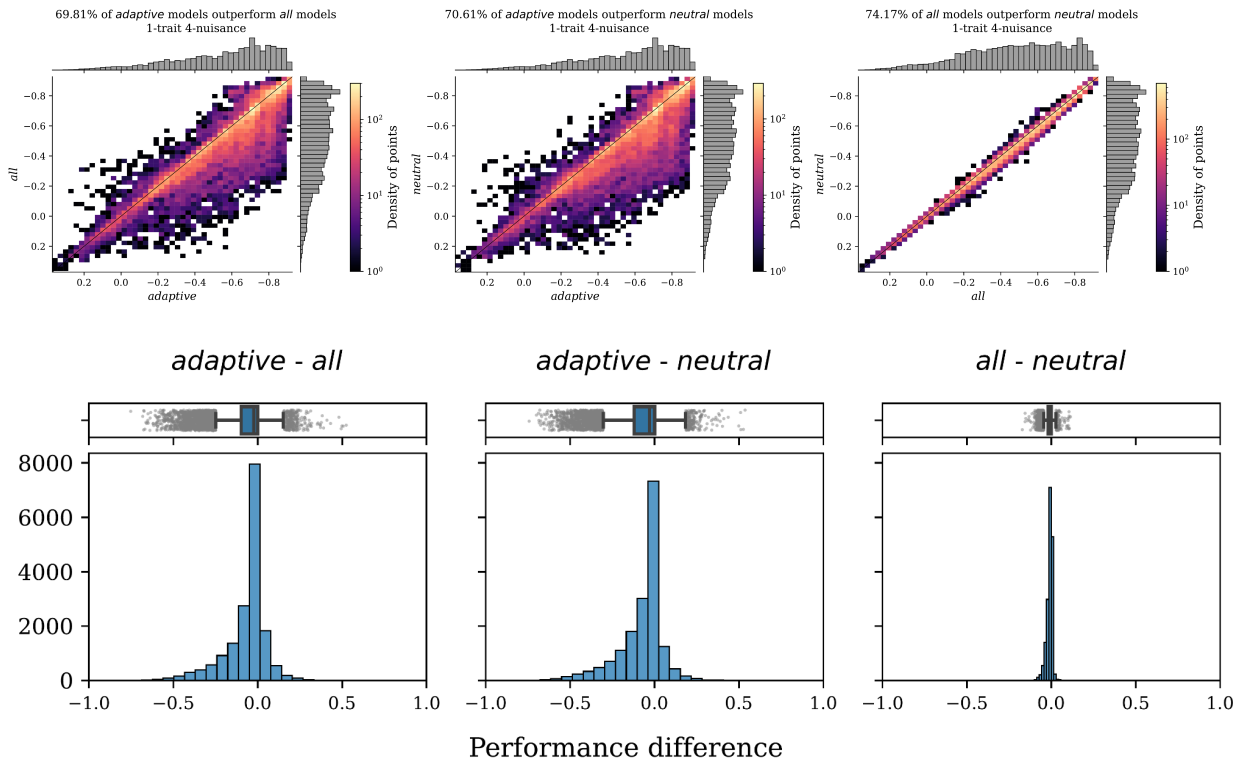


759

760 (Fig. S31 continued)

761

**1-trait 4-nuisance**



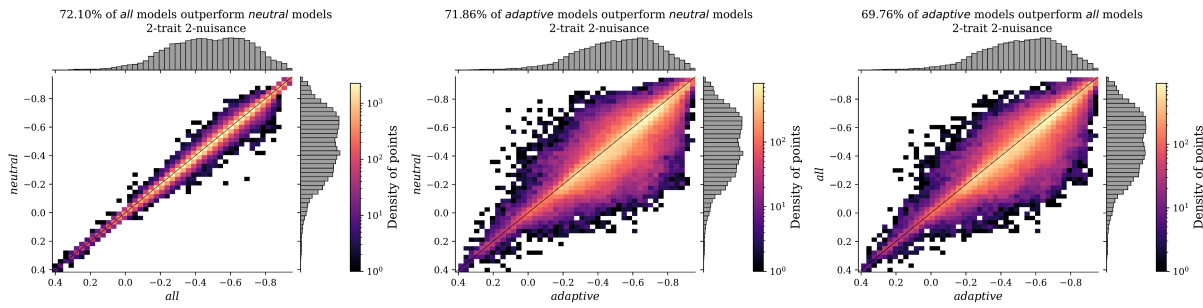
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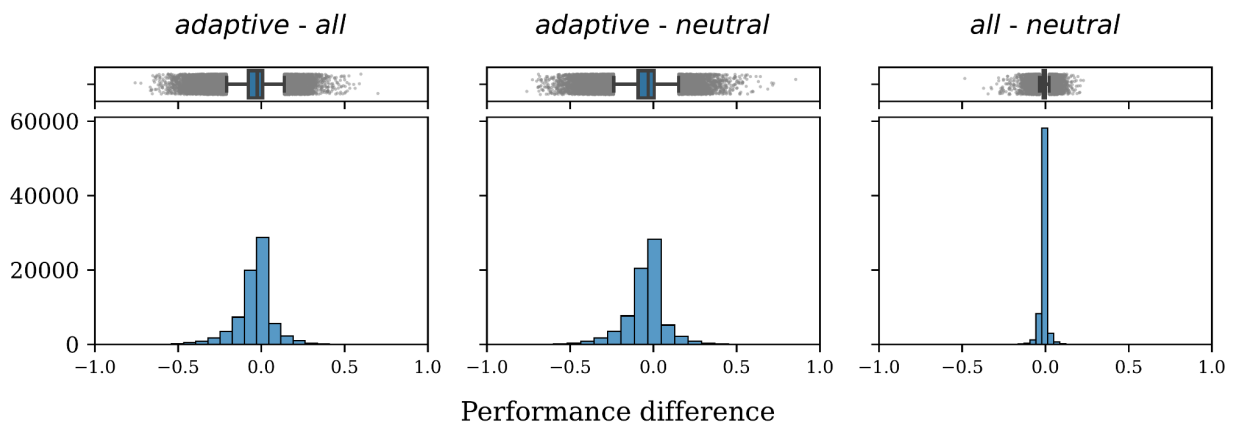
764 (Fig. S31 continued)

765

**2-trait 2- nuisance**



766



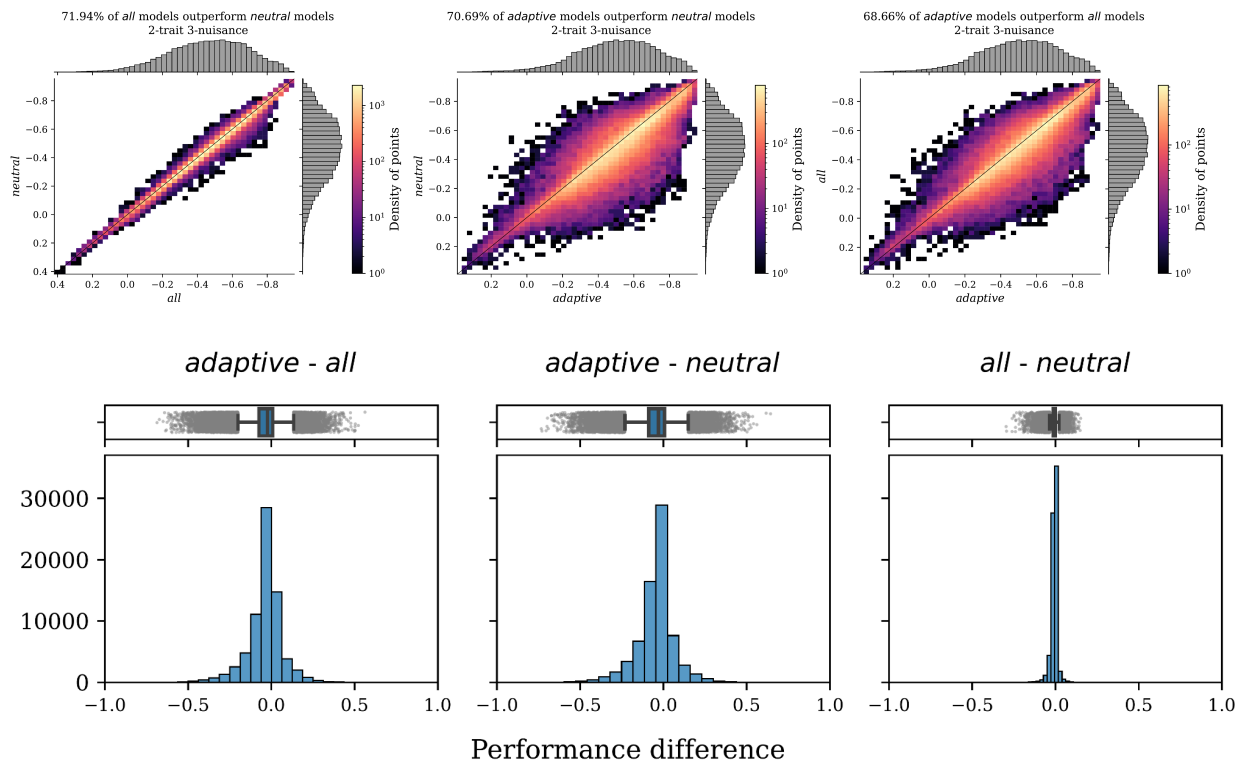
767

768 (Fig. S31 continued)

769

2-trait 3-nuisance

770

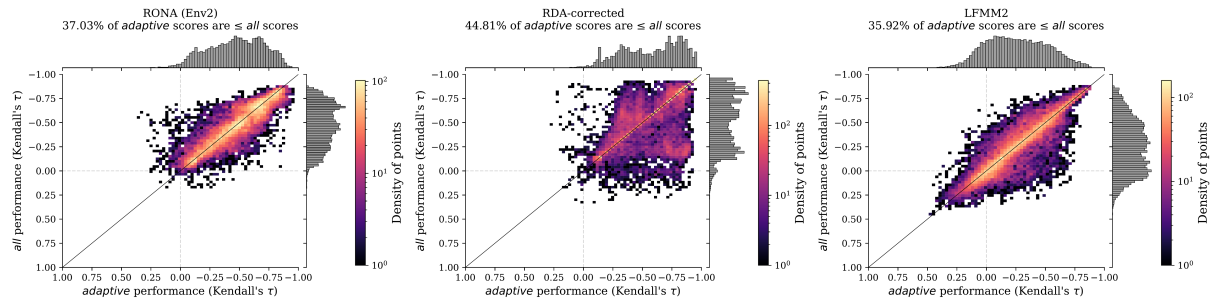


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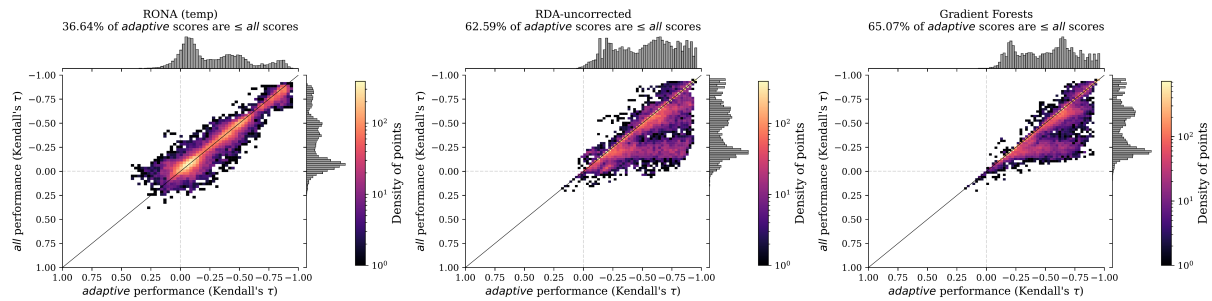
772 **Fig S32** Pairwise comparison of performance differences between marker sets for  
 773 *Nuisance Environment* scenarios. The first row for each nuisance level (*N-trait N-*  
 774 *nuisance*) are scatterplots of pairwise comparisons of performance between  
 775 marker sets (histograms in each margin) from both 1- and 2-trait models where  
 776 density of points is indicated by color in legend (note color scale is different for  
 777 each figure to accentuate patterns in data). The second row for each nuisance level  
 778 are histograms for the difference in performance between marker sets for a given  
 779 model. Method-specific figures are not shown except in SC 02.02.06. Data for these  
 780 figures includes 1- and 2-trait *Nuisance Environment* evaluations. Code to create  
 781 these figures can be found in SC 02.02.06.

782 Fig S33 is in Supplemental Note S4

783



784



785 **Fig S34** Pairwise comparison of performance differences between marker sets for  
 786 *Climate Novelty* scenarios. Shown are scatterplots of pairwise comparisons of  
 787 performance between marker sets (histograms in each margin) from both 1- and 2-  
 788 trait models where density of points is indicated by color in legend (note color scale  
 789 is different for each figure to accentuate patterns in data). Data for these figures  
 790 includes 1- and 2-trait *Climate Novelty* evaluations. Code to create these figures can  
 791 be found in SC 02.04.05.

792 **Supplemental References**

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